

Ambient Ratio Method Version 2
(ARM2) for use with AERMOD for
1-hr NO₂ Modeling

Development and Evaluation Report

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Executive Summary

The majority of the oxides of nitrogen (NO_x) emissions from typical air emission sources are in the form of nitric oxide (NO), whereas EPA has established a National Ambient Air Quality Standard (NAAQS) for nitrogen dioxide (NO_2). EPA has approved a three-tiered screening approach to calculating NO_2 concentrations based on AERMOD predictions of total NO_x concentrations, including the Tier 2 Ambient Ratio Method (ARM) and the Tier 3 Ozone limiting Method (OLM) and Plume Volume Molar Ratio Method (PVMRM).

This report discusses the development and evaluation of an updated Ambient Ratio Method, referred to as “ARM2”. EPA’s current recommendation for ARM uses a fixed ambient ratio of 0.8 for modeling of 1-hr averages, which will often overestimate 1-hr NO_2 concentrations. ARM2 incorporates a variable ambient ratio that is a function of model predicted 1-hr NO_x concentration, based on an analysis of hourly ambient NO_x monitoring data from approximately 580 stations over the period 2001-2010.

ARM2 has some advantages over the current EPA Tier 3 conversion methods for 1-hr NO_2 dispersion modeling. Unlike the Tier 3 methods, ARM2 does not require additional input data that is subject to case-by-case review and approval. The model execution time for ARM2 is faster than for these more computationally intensive refined methods. The ARM2 method performs better than the current ARM method, and is comparable to the more refined EPA modeling methods for 1-hour ambient NO_2 concentrations. ARM2 may help reduce the resources expended by both the regulated facility and the reviewing agency to perform and approve 1-hr NO_2 modeling analyses.

This report presents the data and procedures used to develop ARM2. Performance evaluations and sensitivity analyses of ARM2 and the current refined methods are also presented. The key findings of this study are:

1. Plots of ambient NO_2/NO_x ratios as a function of NO_x concentration from various ambient monitoring data sets show a consistent relationship of decreasing ambient ratios

with increasing NO_x concentrations. At NO_x concentrations above approximately 300 ppb, the observed ambient ratios cluster in a range of approximately 0.1 to 0.2.

2. The ARM2 conversion method was developed using 10 years of ambient monitoring data from 580 monitoring sites in EPA's AQS data base. ARM2 uses an empirically derived relationship between the upper limits of the observed NO₂/NO_x ambient ratio versus the ambient NO_x concentration.
3. ARM2 predicted ambient ratios derived from various geographical, land use, and time period data sets were evaluated and found to be similar to, and typically lower than, the ratios predicted by the "All AQS Sites" equation. This indicates that the ARM2 equation derived from the "All AQS Sites" data set is representative of a wide range of geographical or land use categories, and can be used as the basis of the ARM2 method.
4. The ARM2 conversion method has been programmed into the latest version of AERMOD (version 12345).
5. The performance of the ARM2 method has been compared to monitoring observations and predictions from the Tier 3 screening methods, using the same evaluation data sets that have previously been used to test the Tier 3 methods. Plots of ambient ratios versus NO_x concentration indicate that all methods overpredict the observed ambient ratios. Scatter plots of predicted versus observed NO₂/NO_x ratios indicate that all methods have little skill in predicting the ambient ratio on a "paired in space and time" basis. The Q-Q NO₂ plots also indicate that all three methods over-predict the highest NO₂ concentrations by factors ranging from approximately 1.2 to 2.0.
6. The Robust Highest Concentration (RHC) summary of the evaluation results is a measure of model performance based on the top 26 highest modeled and observed NO₂ concentrations (which range from 55 to 196 µg/m³). The RHC summary indicates that the AERMOD-ARM2, AERMOD-OLM, and AERMOD-PVMRM methods perform very similar, and all 3 methods are on average over-predicting the observed concentrations by factors of 1.8, 1.7, and 2.0 respectively. The ARM2 RHC result falls between the OLM and PVMRM results.
7. Sensitivity analyses were performed for the ARM2, PVMRM, OLM, and full conversion methods across a range of meteorology and source characteristics. The scenarios modeled included updated versions of those described in the MACTEC report

“Sensitivity Analysis of PVMRM and OLM in AERMOD”, 2004, for the diesel generator, gas turbine, and 35-meter stack “single source” scenarios, and the cumulative source scenario. For the single source sensitivity scenarios with low predicted NO_x concentrations (below $20 \mu\text{g}/\text{m}^3$), all of the conversion methods predicted NO_2/NO_x ratios near 0.9. When the predicted NO_x concentrations were higher (greater than $300 \mu\text{g}/\text{m}^3$), all the methods predict NO_2/NO_x ratios in the range of 0.2 to 0.4, with ARM2 conservatively predicting the highest NO_2/NO_x ratios of any of the methods. For the multi-source scenario, the model predicted NO_x concentration is $1,774 \mu\text{g}/\text{m}^3$, and the predicted NO_2/NO_x ratios are 0.13 for OLM, 0.18 for PVMRM, and 0.2 for ARM2.

8. Some additional multi-source scenarios were also analyzed. These scenarios were based on real-world configurations of large diesel IC generators, a refinery, a gas pipeline compressor station, natural gas production fields and processing plants, and a large boiler in complex terrain. The ARM2 method predicts higher or similar ratios when compared to the Tier 3 methods. Two of the cases were further evaluated because of suspiciously high PVMRM predicted ratios, and based on comparisons of PVMRM versus OLM “source-by-source” predicted ratios, it was concluded that PVMRM is predicting unrealistically high concentrations and ratios for those cases. In summary, the additional multi-source sensitivity analysis indicates that the ARM2 method predicts higher or similar ratios when compared to the two Tier 3 methods.

The performance evaluations and sensitivity analyses presented in this report document that the general performance of the ARM2, PVMRM, and OLM methods is similar. The relative performance ranking between the three methods varies depending upon the data set, and there are cases when PVMRM is predicting unrealistically high NO_2/NO_x ratios. PVMRM formulation issues have previously been identified¹, and it is possible that the over-estimation of plume volumes and the resulting number of moles of ozone available for NO conversion may be responsible for the anomalous performance of PVMRM.

¹ Hendrick, E., S. Hanna, B. Egan, V. Tino, 2011, Technical Review of the Ozone Limiting Method (OLM) and Plume Volume Molar Ratio Method (PVMRM) Codes in the ISC3 and AERMOD Models.

Based on the analyses presented herein, it is recommended that ARM2 be approved by EPA as a refinement to the current fixed-ratio ARM method for performing 1-hr NO₂ analyses, or be approved as an additional Tier 3 screening method. Consistent with other recent updates to AERMOD, EPA could implement this refinement by posting a memorandum on the SCRAM website authorizing the use of ARM2 for AERMOD modeling analyses and referencing this report as supporting documentation. An updated version of AERMOD should also be posted that includes the ARM2 model option.

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1.0 Introduction

The majority of the oxides of nitrogen (NO_x) emissions from air emission sources are in the form of nitric oxide (NO), whereas EPA has established a National Ambient Air Quality Standard (NAAQS) for nitrogen dioxide (NO_2). EPA's "Guideline on Air Quality Models" (GAQM), 40 CFR Part 51 Appendix W, describes a three-tiered screening approach to calculating NO_2 concentrations based on dispersion model predictions of NO_x concentrations (NO_x is modeled as if it is a conserved or non-reactive tracer). The three tiers, arranged in order from simplest to most refined, are:

- Tier 1 – Assume full conversion of NO to NO_2 , so that the NO_x predicted by AERMOD is 100 % NO_2
- Tier 2 – Ambient Ratio Method (ARM), where model predicted NO_x concentrations are multiplied by a NO_2/NO_x ambient ratio, derived from ambient monitoring data
- Tier 3 – More detailed methods that account for the plume dispersion and chemistry may be considered on a case-by-case basis, including the Ozone Limiting Method (OLM) and the Plume Volume Molar Ratio Method (PVMRM).

EPA's Appendix W guidance regarding NO_2 modeling addresses only annual averages. EPA has issued additional guidance² on NO_2 conversion methods for 1-hr NO_2 National Ambient Air Quality Standard (NAAQS) modeling (hereafter referred to as the 2011 Additional Clarifications memo). EPA has stated that the Appendix W tiered methodologies do apply to the 1-hour NO_2 NAAQS, but additional issues must be considered. For example, EPA currently recommends the use of an ambient NO_2/NO_x ratio of 0.80 as a default ARM value for 1-hour NO_2 analyses without additional justification by applicants, and EPA has provided some default values for inputs to the Tier 3 OLM and PVMRM methods. However, use of the Tier 3 screening methods still requires justification of key inputs such as in-stack NO_2/NO_x ratios, and approval for use on a case-by-case basis.

² "Additional Clarification Regarding Application of Appendix W Modeling Guidance for the 1-hour NO_2 NAAQS", dated March 1, 2011.

NO_x is emitted from typical combustion sources primarily as nitric oxide (NO), and in smaller quantities directly as NO₂³. NO emissions react with ozone and other oxidants in the atmosphere to convert NO to NO₂. During the early stages of plume dispersion (i.e., at distances ranging from approximately 1 to 10 km over time intervals of 10 to 300 minutes), the principal reaction mechanism for NO₂ formation is oxidation by ozone⁴. Janssen describes how the NO to NO₂ oxidation rate in the atmosphere is limited by both the chemical kinetics and the mixing rate of the plume with ambient air⁵.

The ARM method, originally developed by Chu and Meyers⁶, hypothesized that the complex NO_x chemistry and mixing processes occurring in the atmosphere can be empirically parameterized through the use of simple relations derived from large amounts of ambient monitoring data. They reviewed data from a number of NO_x monitoring sites throughout the US for the period 1987-1989, and calculated the annual average ambient NO₂ and NO_x concentrations and the annual NO₂/NO_x ratios at each site. They then selected the 90th percentile as an upper bound estimate of the conversion, and recommended the default ambient ratio of 0.75. In the annual average ARM method, the highest modeled annual NO_x concentration is multiplied by the ambient ratio to determine the modeled annual NO₂ concentration, which is added to a representative background concentration and compared to the annual NAAQS.

Chu and Meyers also presented plots of the observed hourly ambient ratio as a function of distance from the emission source, based on measurements reported for power plant plume studies and a wind tunnel experiment. These plots illustrated that higher NO_x concentrations and

³ Rethinking the Ozone Problem in Urban and Regional Air Pollution; ISBN: 0-309-56037-3; National Academy of Sciences, 1991.

⁴ Prakash Karamchandani, Annie Koo, and Christian Seigneur. Reduced Gas-Phase Kinetic Mechanism for Atmospheric Plume Chemistry. *Environ. Sci. Technol.* 1998, 32 (11), 1709-1720.

⁵ L. H. J. M. Janssen, T F. T. M. Nieuwstadt, and M. Donze. Time Scales of Physical and Chemical Processes in Chemically Reactive Plumes. *Atmospheric Environment* 1990, 24A (11), 2861-2874.

⁶ Chu and Meyers, "Use of Ambient Ratios to Estimate Impact of NO_x Sources on Annual NO₂ Concentration", presented at the 1991 Air and Waste Management Association annual meeting.

lower ambient ratios were observed near the source, and as the distance increased the NO_x concentration decreased and ambient ratio increased. Other field studies also show this same pattern. EPA cited two ambient monitoring studies in their 2011 Additional Clarifications memo. The Wang study⁷ presents results from four short-term monitoring tests near roadways. The observed ambient NO_2/NO_x ratios varied from approximately 0.3 to 0.8, and the NO_2/NO_x ratios increased with distance from the emission sources while overall NO_x concentrations decreased with distance. Janssen's study⁸ of NO_x chemistry in power-plant plumes also reports that NO_2/NO_x ratios increase as a function of distance (time) from source. Day time data plots show ambient ratios varying from about 0.2 within 2.5 km of sources up to about 0.8 at 10 km. Night time data plots show ambient ratios of about 0.1 within 4 km of sources, increasing up to about 0.4 at 20 km.

These observations of increasing NO_2/NO_x ratios and decreasing NO_x concentrations with distance are consistent with the simplified plume dispersion and NO_x chemistry mechanisms that are the basis of the PVMRM method. As a plume containing NO_x is transported downwind over time, there is increased plume dispersion and entrainment of ambient air. The increased entrainment brings additional ambient ozone into the plume, which causes additional conversion of NO to NO_2 through the fast oxidation reaction. These processes result in increased NO_2/NO_x ratios and decreased NO_x concentrations with the plume's increased dispersion over time.

The above studies suggest that using a fixed ARM ratio (NO_2/NO_x) of 0.8 will overestimate 1-hr NO_2 concentrations in the near field. Therefore, an updated version 2 of ARM has been

⁷ Wang, Y.J., A. DenBleyker, E. McDonald-Buller, D. Allen and K. Zhang, 2011. Modeling the chemical evolution of nitrogen oxides near roadways. *Atmos. Env.*, 45, 43-52.

⁸ Janssen, L.M.J.M., F. Van Haren, P. Bange, and H. Van Duuren, 1991. Measurements and modelling of reactions of nitrogen oxides in power-plant plumes at night. *Atmos. Env.*, 25A, No. 5/6, 829-840.

developed, referred to as “ARM2”. ARM2 is designed to predict more realistic ambient NO_2 concentrations because it incorporates a variable ambient NO_2/NO_x ratio.

Section 2 of this report describes the ambient monitoring data and data processing procedures used to compile 1-hr ambient NO_2/NO_x ratios. Section 3 describes the development of the ARM2 methodology. Section 4 presents evaluations of ARM2 using the same performance evaluation data sets that have previously been used to test the OLM and PVMRM methods. Section 5 presents sensitivity studies and comparisons of ARM2, OLM, and PVMRM for various modeling scenarios. Section 6 presents the overall study conclusions and recommendations.

2.0 Ambient NO_x Data Sources and Analysis

The analysis of ambient monitoring data for ARM2 development used the same basic concepts as Chu and Meyers in their development of the original ARM. The observed ambient NO₂/NO_x ratios from a large number of NO_x monitoring sites throughout the US have been calculated, and these data are used to develop an equation that calculates an upper bound of the ambient ratio as a function of the total NO_x concentration.

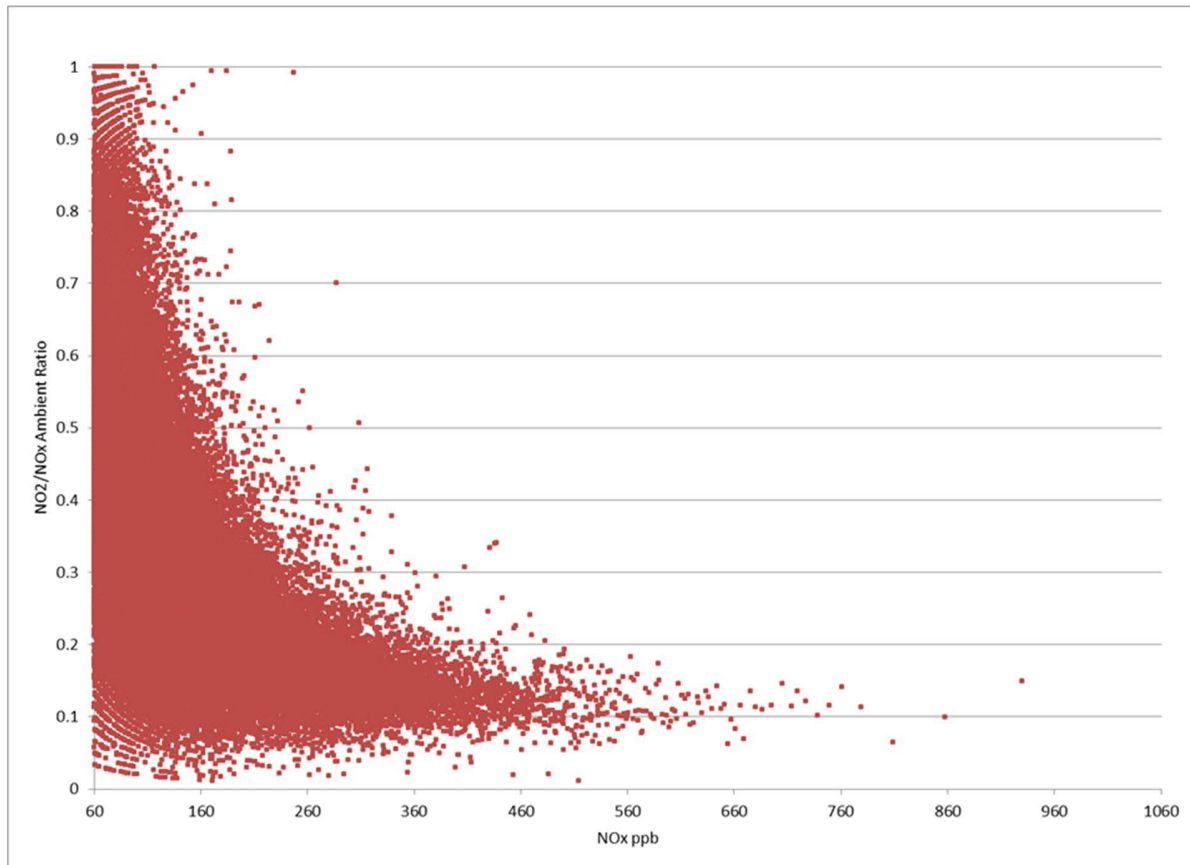
Recent 1-hr NO₂ and NO_x data from numerous monitoring sites throughout the US have been evaluated. These include three main monitoring data sets:

1. EPA's Air Quality System (AQS) 1-hr averaged NO₂ and NO_x ambient air quality data for the entire country from 2001 through 2010. This data is available in AQS "input transaction format" at <http://epa.gov/ttn/airs/airsaqs/detaildata/downloadaqsddata.htm>. Data from approximately 580 monitoring sites are available in this data set, and the number of valid hours of data with NO_x concentrations greater than 1 ppb is approximately 23,000,000.
2. Monitoring data from the performance evaluation data sets that have previously been used to test the Tier 3 methods, including the Empire Abo, NM, Palaau, HI, and Wainwright, AK, data sets.
3. A monitoring data set collected in the Athabasca Oil Sands region of Alberta, Canada, as reported by Jain⁹. These ambient data have been collected at 16 monitoring stations located throughout the Oil Sands Region, where numerous compressor engines and other NO_x emission sources are located.

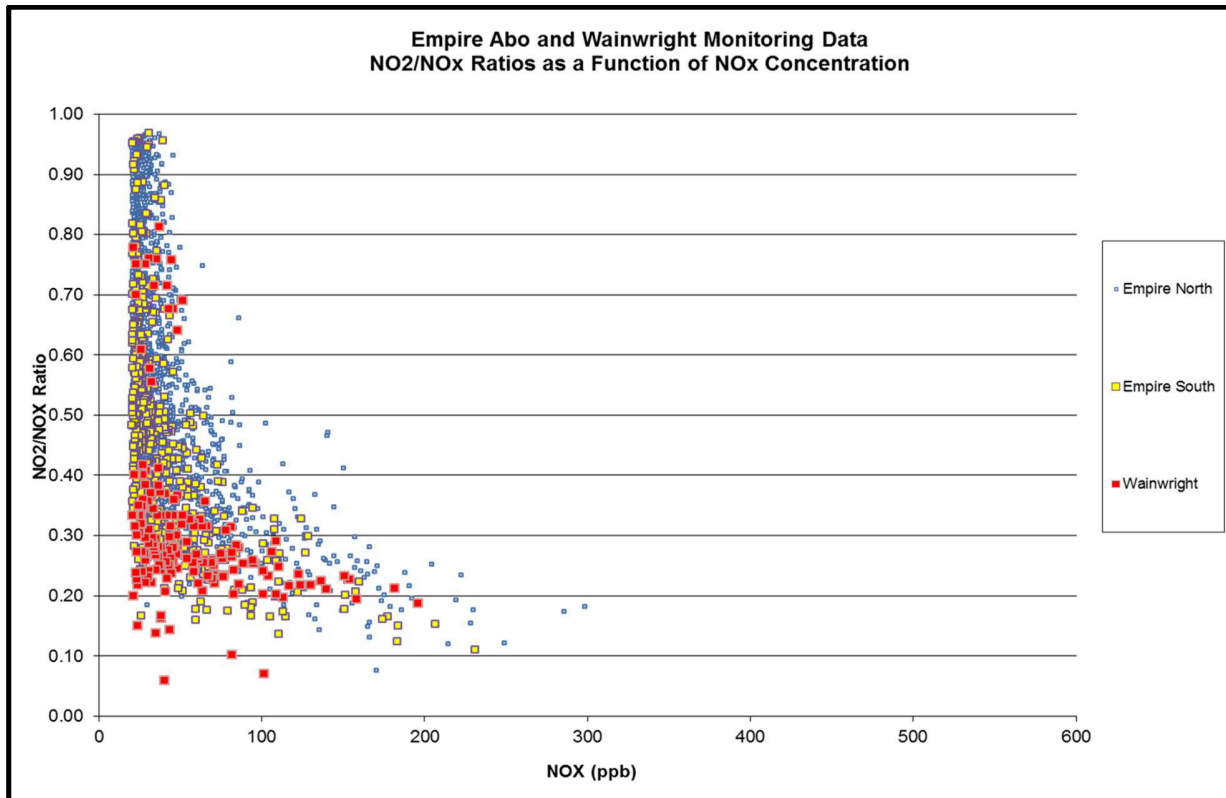
⁹ Jain, Önder, and Hoeksema in "Comparison of Ambient Ratio Method and Ozone Limiting Method for determining NO₂ Concentrations in the Athabasca Oil Sands Region", presented at the 2011 Air and Waste Management Association annual meeting.

The ambient monitoring data were first evaluated by calculating and plotting the hourly ambient NO_2/NO_x ratio. The Wang and Jannsen studies were designed so that ambient ratios were directly measured as a function of distance from the specific emission sources, so that distance was used as the X axis to represent the amount of plume dispersion and entrainment that has occurred. The use of distance as the X variable is not possible for the data sets analyzed herein, because detailed information on the distance to sources is not available for the numerous ambient monitoring sites analyzed (in addition, each monitoring site may be impacted by emission sources at various distances and directions). Therefore, the ambient ratios were plotted against the measured NO_x concentration, which is a variable that, like distance, is directly related to the amount of plume dispersion and entrainment that has occurred.

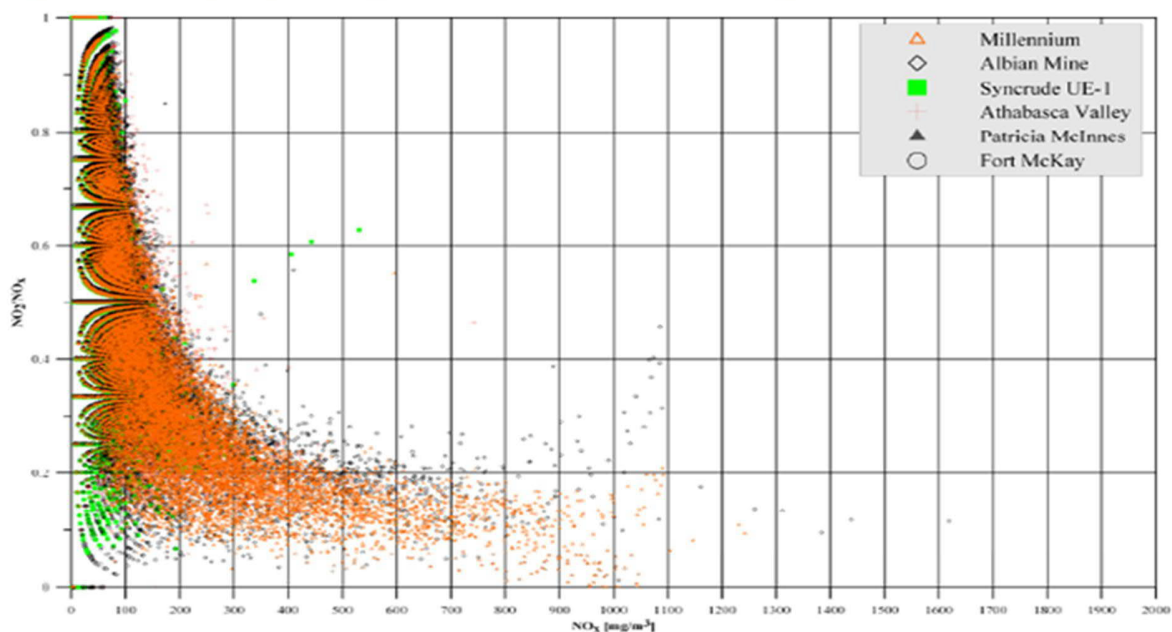
The plots of ambient ratios as a function of NO_x concentration from the three sets of data all show similar relationships, as illustrated in Figures 1 through 3. All plots indicate the same trend in decreasing ambient ratio with increasing NO_x concentrations. At higher NO_x concentrations (greater than 350 ppb), the observed ambient ratios tend to cluster in a range of approximately 0.1 to 0.2, which is similar to a typical “in-stack” NO_2/NO_x ratio and suggests very little NO_2 conversion beyond the in-stack levels. At lower NO_x concentrations (below 150 ppb) there is a much wider range of ambient ratios observed; this is a result of the varying transport and conversion times that are associated with low NO_x concentrations (low NO_x concentrations may be caused by nearby sources with low emissions, as well as more distant sources with higher emissions).

Figure 1– NO₂/NO_x Ratios for All AQS Monitoring Sites for 2001-2010

NOTE: There are approximately 6,800,000 valid hourly data points from 530 monitoring stations in this data plot. The X-axis for this plot begins at 60 ppb to limit the number of hourly data points to a size that the spreadsheet program can process.

Figure 2 – NO₂/NO_x Ratios for Empire Abo and Wainwright Data

NOTE: There are approximately five years of monitoring data plotted in this graph. The minimum NO_x concentration begins at 20 ppb, which was used as a threshold to address the noise and variability in the measurements at low NO_x concentration ranges.

Figure 3 – NO₂/NO_x Ratios for Athabasca Oil Sands Data Set**Figure 1. Hourly NO₂ and NO_x Measurements at WBEA monitoring Stations**

Note: Data from six monitors for the study year of 2006 are plotted. There appears to be an “upward tail” or spike of ambient ratios for NO_x concentrations around 900-1100 µg/m³ (about 500-600 ppb) at the Albion mine (blue) and Millennium (orange) monitoring stations. Based on discussions with the study author, it is likely that these ambient ratio “spikes” at the 500 ppb NO_x concentration level are caused by the 500 ppb full scale setting of some NO_x analyzers, with NO_x measurements being “saturated” or “flattening out” at concentrations above 500 ppb. If the NO_x concentration measurements are saturated, this will result in increasing NO₂/NO_x ratios that are not representative of actual ambient ratios.

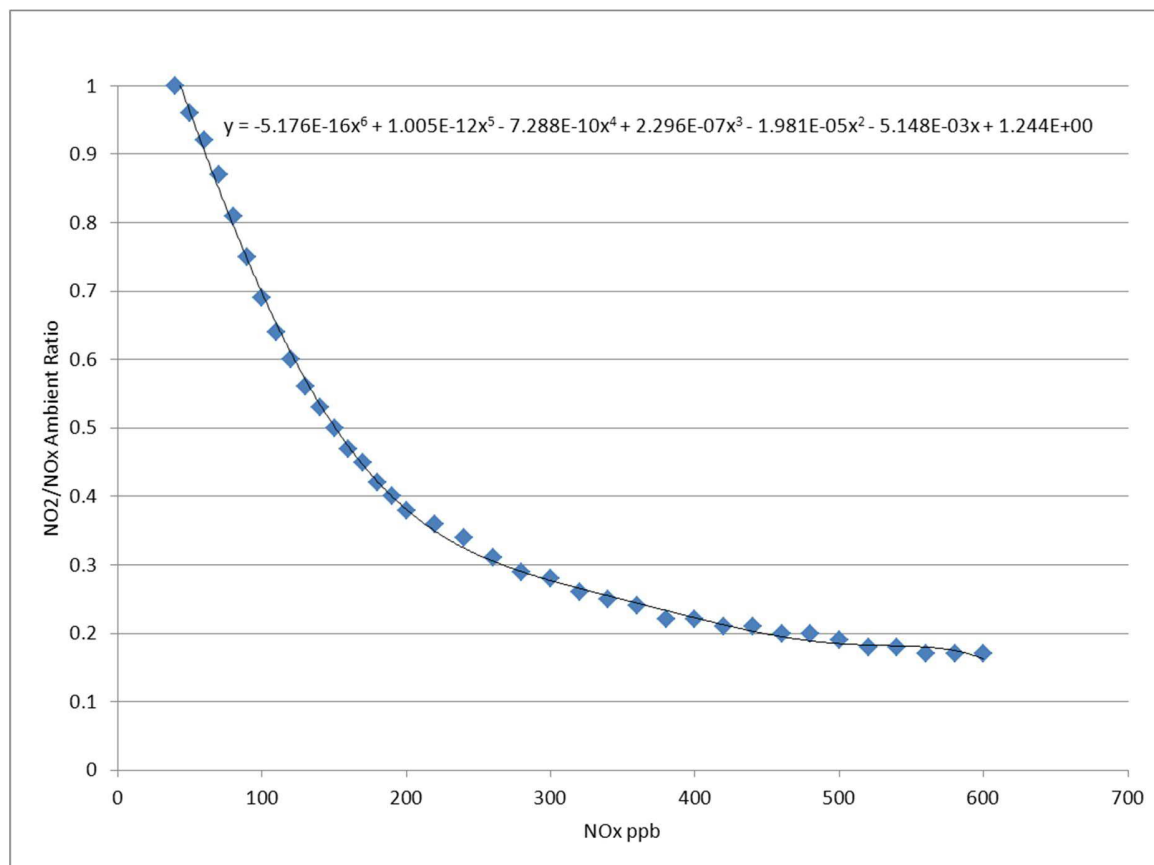
3.0 Development of ARM2 Method

3.1 Initial Data Analysis

Chu and Meyers originally calculated the annual average ambient NO_2/NO_x ratios from a large ambient monitoring data set, and chose the 90th percentile value as a reasonable upper bound estimate for the default ARM ambient ratio. They also used 20 ppb as a lower NO_x concentration threshold to avoid the noise inherent in ambient monitor measurements at low concentration ranges.

For the ARM2 data analysis, hourly ambient NO_2/NO_x ratios were calculated from EPA's large, nationwide AQS data base and sorted into NO_x concentration "bins". A NO_x threshold of 20 ppb was again used as a lower concentration "cut-off". The NO_x concentration bins were 10 ppb wide over the range 20 to 200 ppb, and 20 ppb wide over the range 200 to 600 ppb (note there were very few NO_x measurements greater than 600 ppb). The upper bound of the observed ambient ratios was estimated for each NO_x concentration bin by calculating the 98th percentile of the observed ratios in that bin (see Section 3.3 of this report for a discussion of the use of the 98th percentile value). This results in a conservatively high estimate of the ambient ratio. To address small sample sizes, the data processing program reported the second highest ambient ratio as the 98th percentile value for the cases when there are less than 25 observations in a bin.

Figure 4 presents a plot of the calculated 98th percentile bin values for the entire "All AQS Sites" data set. The 98th percentile NO_2/NO_x ratios for each bin were fitted to a polynomial equation to develop the ARM2 conversion factor equation, which is also shown in Figure 8. ARM2 uses this empirically derived relationship between the upper limits of the observed NO_2/NO_x ambient ratio versus the ambient NO_x concentration. Appendix B presents detailed data from the analysis, including the number of observations in each NO_x concentration bin and the calculated percentile value for each concentration bin.

Figure 4 – 98th Percentile Ambient Ratios and ARM2 Equation for “All AQS Sites” Data

The ambient ratios calculated by the ARM2 equation were further constrained to a maximum value of 0.9 and a minimum value of 0.2. The maximum value is based on the current EPA recommendation for the maximum “equilibrium” ratio of 0.9 for the refined Tier 3 conversion methods. The minimum ambient ratio is based on two criteria, the analysis of ambient NO₂/NO_x ratios, and consideration of in-stack ratios for typical sources (which represents the minimum ambient ratio “floor”). First, the ambient monitoring data in Figures 1 through 3 were reviewed. At NO_x concentrations above 300 ppb, the observed ambient ratios tended to cluster in a range of approximately 0.1 to 0.2. Second, in-stack ratio information was reviewed to determine typical

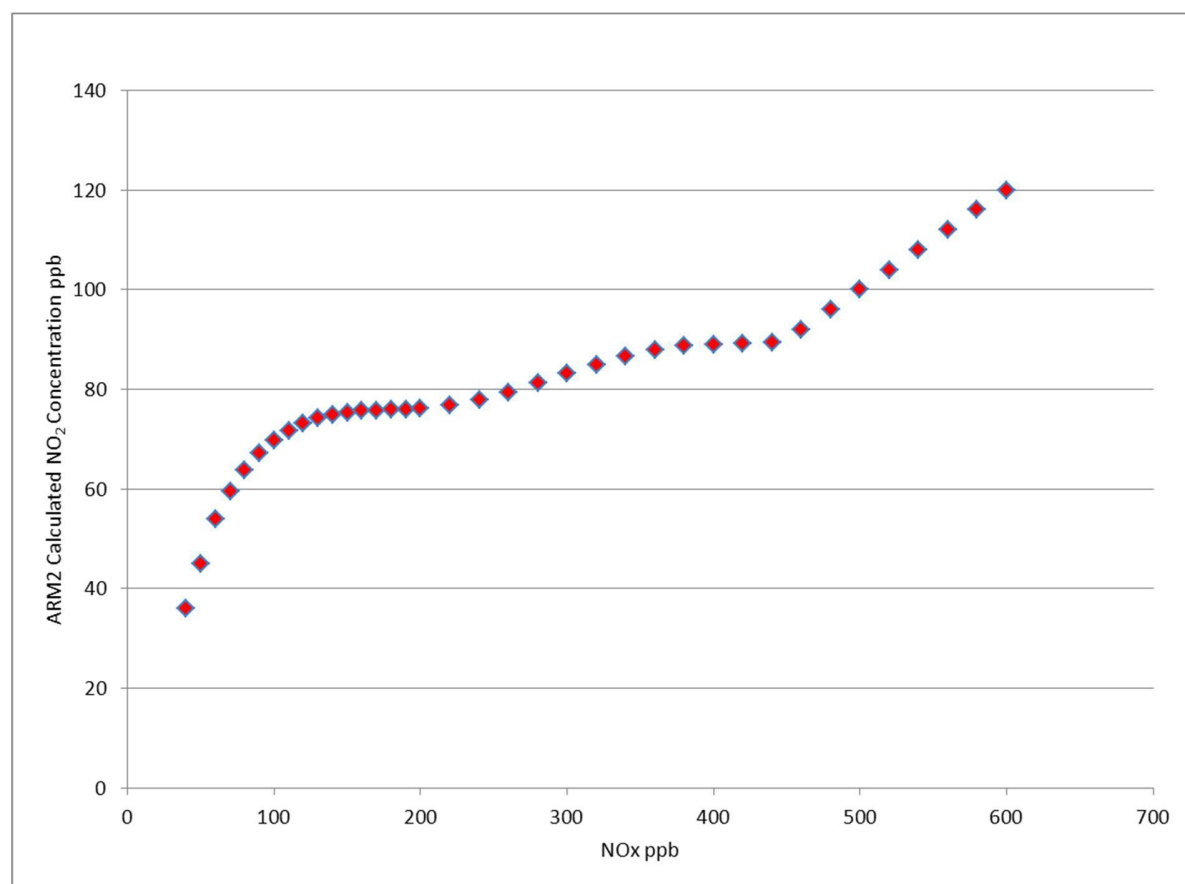
in-stack ratios for common source types. EPA's "alpha" version of the in-stack ratio data base¹⁰ was reviewed. The available data were sorted by source category and a reasonable upper limit (90th percentile) of the test data was determined. The resulting in-stack ratio for boilers and combustion turbines is approximately 0.1, and for IC engines approximately 0.2. This is consistent with data compiled by both the Alaska DEC¹¹ and the California Air Pollution Control Officers Association (CAPCOA)¹², which indicate that most in-stack ratios vary between 0.05 and 0.2 for these three source categories. Both the ambient ratio curves and the "consensus" in-stack ratio data indicate that a minimum ambient ratio value of 0.2 would address most applications of ARM2 (as discussed later, ARM2 has been implemented in AERMOD with an optional input value for the minimum ambient ratio, to address unique situations where the in-stack ratio for the modeled source may be higher).

Figure 5 is a plot of the resulting ARM2 calculated NO₂ concentrations as a function of NO_x concentrations. The shape of this graph is consistent with Chu and Meyer's and Janssen's observations that NO₂ formation is limited at higher NO_x concentrations by both chemical kinetics and plume mixing rates. The ARM2 equation results in steadily increasing NO₂ concentrations at NO_x concentrations below approximately 150 ppb. The NO₂ concentration then levels off and slowly increases (the calculated NO₂ concentration never decreases as the NO_x concentration increases). After the NO_x concentration reaches approximately 450 ppb, the NO₂ concentration linearly increases with increasing NO_x concentration which is a result of the minimum ambient ratio of 0.2.

¹⁰ http://www.epa.gov/ttn/scram/no2_isr_database.htm

¹¹ <http://dec.alaska.gov/air/ap/docs/NO2-NOx%20Instack%20Ratios%20from%20Source%20Tests%205-21-12.xlsx>

¹² http://www.valleyair.org/busind/pto/Tox_Resources/CAPCOANO2GuidanceDocument10-27-11.pdf

Figure 5 –ARM2 Calculated NO₂ Concentrations

3.2 Sensitivity of Ambient Ratio Equation to Monitoring Data Set

The sensitivity of the ARM2 equation to the data set used in the analysis was evaluated. The ARM2 equation was derived (as described in Section 3.1) using AQS data from different geographical regions, different land use categories (rural/suburban versus urban/city center as defined in the AQS monitor data base), and different time periods (2001-2003, 2004-2007, and 2008-2010). Data subsets were also investigated by selecting monitoring sites that were estimated to be within certain distances from sources with greater than 100 tpy of NO_x emissions (based on source data reported in EPA's 2008 National Emission Inventory data base at <http://www.epa.gov/ttn/chief/net/2008inventory.html>). For each data subset, a spreadsheet is presented in Appendix B that lists the number of observations in each NO_x concentration bin and the ARM2 calculated ambient ratio for each concentration bin. Again, the ambient ratios were constrained to a minimum of 0.2 and a maximum of 0.9.

For most of the data subsets, there was an elevated value for the 98th percentile ratio at the 500 ppb NO_x concentration bin when compared to the adjacent bins on either side. This is similar to the “upward spike” observed in Figure 3 for ambient ratios near the 500 ppb NO_x concentration level. It is likely that these ambient ratio “spikes” are caused by the 500 ppb full scale setting of most NO_x analyzers, with the NO_x measurements being limited at concentrations above 500 ppb. Therefore, the ambient ratios for the 500 ppb NO_x concentration bin were calculated as the average of the ratios for the two adjacent bins.

Table 1 summarizes the geographical region and land use category results. The first two columns in Table 1 list the number of observations and the “All AQS Sites” derived ambient ratios for the various NO_x concentration bins. The remaining pairs of columns for each of the data sets present first the number of observations in the bins, followed by the relative percent difference (RPD) for the calculated ambient ratio versus the “All AQS Sites” ambient ratio. In addition, at the bottom of the table the mean RPD is calculated over all NO_x concentration bins for each data set. The use of the RPD is a simple indicator for quantifying the differences in ambient ratios between the

various data sets relative to the “All AQS Sites” equation. The number of observations in each bin is presented to help weight the significance of any observed differences (large differences in sample size may contribute to observed variability amongst the data sets).

The data in Table 1 indicate that the ratios determined from various geographical and land use data sets are all very similar to the “All AQS Sites” ratios, with mean RPDs within approximately $\pm 10\%$. The only data sets with a positive mean RPD (i.e., data sets that result in higher ambient ratios than the full AQS data set) are the Urban data (with a +0.4% mean RPD) and the Midwest region data (with a +1.9% mean RPD). Some of the difference in the Midwest data set may be attributable to the differences in sample size. In summary, Table 1 indicates that the ambient ratios calculated using the “All AQS Sites” data set are generally higher and therefore more conservative than the ratios predicted by the various regional or land use derived equations. There is no need to develop separate ARM2 equations for specific geographical or land use categories.

Table 1 - Calculated Ambient Ratios by Geographical Regions and Land Use Categories

| Upper Nox BIN ppb | ALL Sites # Obs | ALL Sites Ambient Ratios | Rural & Suburban # Obs | Rural & Suburban Ambient Ratio RPD | Urban & City Center # Obs | Urban & City Center Ambient Ratio RPD | Northeast Region # Obs | Northeast Region Ambient Ratio RPD | Southeast Region # Obs | Southeast Region Ambient Ratio RPD | Midwest Region # Obs | Midwest Region Ambient Ratio RPD | Mountain Region # Obs | Mountain Region Ambient Ratio RPD | Southwest Region # Obs | Southwest Region Ambient Ratio RPD |
|-------------------|--------------------|-----------------------------|------------------------------|---|---------------------------------|---|------------------------------|---|------------------------------|---|----------------------------|---|-----------------------------|--|------------------------------|---|
| 30 | 2,330,786 | 0.90 | 1,295,731 | 0% | 1,020,825 | 0% | 643,420 | 0% | 179,401 | 0% | 283,397 | 0% | 84,000 | 0% | 324,695 | 0% |
| 40 | 1,310,767 | 0.90 | 703,035 | 0% | 601,939 | 0% | 375,722 | 0% | 82,728 | 0% | 156,803 | 0% | 51,663 | 0% | 161,405 | 0% |
| 50 | 817,380 | 0.90 | 426,294 | 0% | 388,295 | 0% | 236,081 | 0% | 45,511 | 0% | 95,480 | 0% | 35,255 | 0% | 92,570 | 0% |
| 60 | 546,757 | 0.90 | 280,390 | 0% | 264,973 | 0% | 157,607 | -3% | 27,839 | -3% | 62,403 | -3% | 25,925 | -6% | 58,213 | -3% |
| 70 | 385,505 | 0.85 | 194,496 | 0% | 190,229 | 0% | 110,023 | -5% | 18,693 | -5% | 43,440 | -4% | 19,801 | -8% | 39,442 | -6% |
| 80 | 282,446 | 0.80 | 142,124 | 0% | 139,879 | 0% | 79,065 | -6% | 13,197 | -8% | 30,934 | -4% | 15,568 | -9% | 28,437 | -8% |
| 90 | 214,245 | 0.75 | 107,859 | 0% | 106,102 | 0% | 58,733 | -7% | 9,866 | -10% | 22,693 | -3% | 12,591 | -10% | 21,170 | -9% |
| 100 | 164,999 | 0.70 | 82,582 | 0% | 82,251 | 0% | 44,084 | -8% | 7,427 | -11% | 16,907 | -3% | 10,148 | -10% | 16,697 | -11% |
| 110 | 130,316 | 0.65 | 64,674 | 0% | 65,527 | 0% | 34,197 | -9% | 5,635 | -13% | 12,983 | -2% | 8,258 | -11% | 12,993 | -12% |
| 120 | 104,093 | 0.61 | 52,123 | 0% | 51,895 | 0% | 26,539 | -10% | 4,433 | -14% | 10,210 | -2% | 6,866 | -10% | 10,358 | -13% |
| 130 | 84,056 | 0.57 | 41,692 | 0% | 42,318 | 0% | 21,049 | -11% | 3,688 | -16% | 7,780 | -1% | 5,671 | -10% | 8,458 | -14% |
| 140 | 68,350 | 0.54 | 34,307 | 0% | 34,009 | 0% | 16,665 | -11% | 2,971 | -17% | 6,191 | 0% | 4,531 | -9% | 6,830 | -15% |
| 150 | 56,801 | 0.50 | 28,242 | 0% | 28,542 | 0% | 13,535 | -12% | 2,566 | -17% | 4,955 | 1% | 3,961 | -8% | 5,833 | -16% |
| 160 | 46,956 | 0.47 | 23,148 | 0% | 23,785 | 0% | 10,935 | -13% | 2,110 | -18% | 3,948 | 1% | 3,313 | -7% | 4,844 | -17% |
| 170 | 39,220 | 0.45 | 19,288 | 0% | 19,921 | 0% | 9,027 | -13% | 1,762 | -18% | 3,332 | 2% | 2,781 | -6% | 4,107 | -17% |
| 180 | 32,769 | 0.42 | 15,995 | 0% | 16,761 | 0% | 7,477 | -13% | 1,475 | -19% | 2,692 | 2% | 2,278 | -5% | 3,519 | -17% |
| 190 | 27,696 | 0.40 | 13,403 | 0% | 14,290 | 0% | 6,202 | -13% | 1,350 | -19% | 2,158 | 3% | 1,953 | -4% | 2,952 | -18% |
| 200 | 23,436 | 0.38 | 11,235 | 0% | 12,192 | 0% | 5,107 | -13% | 1,156 | -18% | 1,816 | 3% | 1,630 | -3% | 2,524 | -18% |
| 220 | 36,618 | 0.35 | 17,754 | 0% | 18,854 | 0% | 8,093 | -13% | 1,775 | -18% | 2,811 | 3% | 2,584 | -1% | 3,780 | -17% |
| 240 | 26,373 | 0.32 | 12,477 | 0% | 13,893 | 1% | 5,695 | -12% | 1,331 | -17% | 2,010 | 2% | 1,784 | 0% | 2,748 | -17% |
| 260 | 19,349 | 0.31 | 9,007 | -1% | 10,338 | 1% | 4,234 | -10% | 1,026 | -16% | 1,387 | 1% | 1,270 | 1% | 2,048 | -16% |
| 280 | 14,204 | 0.29 | 6,700 | -1% | 7,503 | 1% | 3,116 | -9% | 813 | -16% | 952 | 1% | 849 | 2% | 1,487 | -15% |
| 300 | 10,468 | 0.28 | 4,853 | -1% | 5,615 | 1% | 2,312 | -7% | 583 | -16% | 780 | 0% | 631 | 2% | 1,124 | -14% |
| 320 | 7,816 | 0.27 | 3,571 | -1% | 4,245 | 2% | 1,772 | -6% | 512 | -16% | 536 | 0% | 422 | 2% | 874 | -13% |
| 340 | 5,774 | 0.25 | 2,538 | -1% | 3,236 | 2% | 1,322 | -5% | 356 | -16% | 387 | 1% | 308 | 2% | 653 | -12% |
| 360 | 4,293 | 0.24 | 1,883 | -1% | 2,409 | 2% | 960 | -4% | 318 | -16% | 310 | 3% | 199 | 2% | 494 | -11% |
| 380 | 3,287 | 0.23 | 1,430 | -1% | 1,857 | 2% | 772 | -4% | 275 | -14% | 224 | 5% | 125 | 2% | 374 | -11% |
| 400 | 2,524 | 0.22 | 1,029 | -1% | 1,494 | 1% | 606 | -4% | 228 | -10% | 200 | 8% | 74 | 2% | 271 | -10% |
| 420 | 1,880 | 0.21 | 769 | -1% | 1,111 | 1% | 444 | -4% | 156 | -6% | 148 | 12% | 63 | 2% | 200 | -6% |
| 440 | 1,401 | 0.20 | 578 | -1% | 823 | 0% | 332 | -1% | 130 | -1% | 111 | 14% | 24 | 1% | 155 | -1% |
| 460 | 1,079 | 0.20 | 447 | 0% | 632 | 0% | 282 | 0% | 99 | 0% | 100 | 13% | 25 | 0% | 122 | 0% |
| 480 | 810 | 0.20 | 336 | 0% | 474 | 0% | 187 | 0% | 80 | 0% | 91 | 9% | 12 | 0% | 85 | 0% |
| 500 | 658 | 0.20 | 283 | 0% | 375 | 0% | 173 | 0% | 61 | 0% | 99 | 4% | 5 | 0% | 53 | 0% |
| 520 | 413 | 0.20 | 173 | 0% | 240 | 0% | 111 | 0% | 44 | 0% | 25 | 0% | 2 | 0% | 41 | 0% |
| 540 | 310 | 0.20 | 122 | 0% | 188 | 0% | 82 | 0% | 32 | 0% | 10 | 0% | 4 | 0% | 22 | 0% |
| 560 | 239 | 0.20 | 91 | 0% | 148 | 0% | 74 | 0% | 27 | 0% | 16 | 0% | 1 | 0% | 29 | 0% |
| 580 | 159 | 0.20 | 57 | 0% | 102 | 0% | 41 | 0% | 14 | 0% | 10 | 0% | 2 | 0% | 14 | 0% |
| 600 | 115 | 0.20 | 47 | 0% | 68 | 0% | 36 | 0% | 13 | 0% | 7 | 0% | 1 | 0% | 5 | 0% |
| Total # Points | 6,804,348 | NA | 3,600,763 | NA | 3,177,338 | NA | 1,886,112 | NA | 419,681 | NA | 778,336 | NA | 304,578 | NA | 819,626 | NA |
| Mean % Difference | NA | NA | NA | -0.2% | NA | 0.4% | NA | -6.3% | NA | -10.2% | NA | 1.9% | NA | -2.8% | NA | -9.4% |

NOTE: Column 3 presents the ARM2 ambient ratio based on the “ALL AQS Sites” data set. The odd columns from 5 to 17 present the relative percent difference between the calculated ambient ratio for the specified data subsets versus the “ALL AQS Sites” ambient ratio for each NO_x concentration bin.

Table 2 presents the ambient ratio comparison data for different time periods extracted from the AQS data base. The mean RPD is higher (+7%) in the 2001-2003 time period versus the complete 2001-2010 data period, while the mean RPD for the 2004-2007 and 2008-2010 time periods are lower (-1%) versus the complete 2001-2010 data period. This could be related to natural variability in meteorological and ozone background conditions, or it may indicate that higher ambient ozone concentrations occurred throughout the US during 2000-2003 as compared to more recent years, which would have resulted in higher oxidation rates of NO emissions.

Table 3 presents the ambient ratio comparison data based on varying distances of the monitoring site from 100 tpy or greater NO_x emission sources, as determined by processing the 2008 NEI source emission data base. These data indicate that the calculated 98th percentile ambient ratios for monitoring sites within 1 km of an emission sources are generally lower than the “All AQS Sites” ambient ratios. The data subsets for monitoring sites located from 1 to 5 km distance, and greater than 5 km from NO_x sources, are similar to the All Sites ambient ratios. These results are consistent with the basic plume dispersion, entrainment, and chemistry conceptual mechanisms in the near field, and with the field observations reported in the Wang et al. and Janssen reports (i.e., at distances nearer to the emission source, the ambient ratios are lower than at further distances).

The overall conclusion of this sensitivity analysis is that the ARM2 conversion equation derived from the “All AQS Sites” data set generally results in higher ambient ratios than equations derived from the other data subsets. Therefore, the “All AQS Sites” equation can be conservatively used as the basis of the ARM2 method.

Table 2 - Ambient Ratio Comparison for Different Time Period Subsets of AQS Data

| Upper Nox BIN ppb | 2001-2010 # Obs | 2001-2010 Ambient Ratios | 2001-2003 # Obs | 2001-2003 Ambient Ratio RPD | Normalized 2004-2007 # Obs | 2004-2007 Ambient Ratio RPD | 2008-2010 # Obs | 2008-2010 Ambient Ratio RPD |
|-------------------|--------------------|-----------------------------|--------------------|-----------------------------------|----------------------------------|-----------------------------------|--------------------|-----------------------------------|
| 30 | 2,330,786 | 0.90 | 744,455 | 0% | 716,871 | 0% | 630,503 | 0% |
| 40 | 1,310,767 | 0.90 | 443,379 | 0% | 404,004 | 0% | 328,716 | 0% |
| 50 | 817,380 | 0.90 | 287,844 | 0% | 250,812 | 0% | 195,120 | 0% |
| 60 | 546,757 | 0.90 | 197,928 | 0% | 166,304 | -1% | 127,090 | -4% |
| 70 | 385,505 | 0.85 | 142,352 | 4% | 116,393 | -2% | 87,963 | -5% |
| 80 | 282,446 | 0.80 | 105,246 | 5% | 85,073 | -2% | 63,770 | -7% |
| 90 | 214,245 | 0.75 | 80,624 | 6% | 64,244 | -3% | 47,963 | -8% |
| 100 | 164,999 | 0.70 | 63,110 | 7% | 49,059 | -3% | 36,477 | -8% |
| 110 | 130,316 | 0.65 | 50,459 | 8% | 38,781 | -4% | 28,149 | -9% |
| 120 | 104,093 | 0.61 | 41,019 | 8% | 30,625 | -4% | 22,241 | -9% |
| 130 | 84,056 | 0.57 | 33,297 | 9% | 24,730 | -5% | 17,786 | -10% |
| 140 | 68,350 | 0.54 | 27,659 | 9% | 20,171 | -5% | 13,796 | -10% |
| 150 | 56,801 | 0.50 | 23,192 | 9% | 16,855 | -6% | 11,136 | -10% |
| 160 | 46,956 | 0.47 | 19,517 | 9% | 13,895 | -6% | 8,913 | -10% |
| 170 | 39,220 | 0.45 | 16,682 | 9% | 11,472 | -7% | 7,242 | -10% |
| 180 | 32,769 | 0.42 | 14,130 | 9% | 9,564 | -7% | 5,887 | -10% |
| 190 | 27,696 | 0.40 | 12,315 | 9% | 7,937 | -7% | 4,798 | -9% |
| 200 | 23,436 | 0.38 | 10,569 | 9% | 6,761 | -8% | 3,853 | -9% |
| 220 | 36,618 | 0.35 | 16,807 | 8% | 10,649 | -8% | 5,613 | -8% |
| 240 | 26,373 | 0.32 | 12,542 | 7% | 7,590 | -7% | 3,711 | -6% |
| 260 | 19,349 | 0.31 | 9,452 | 6% | 5,578 | -6% | 2,460 | -5% |
| 280 | 14,204 | 0.29 | 7,207 | 5% | 4,040 | -6% | 1,611 | -3% |
| 300 | 10,468 | 0.28 | 5,515 | 5% | 2,902 | -5% | 1,084 | -2% |
| 320 | 7,816 | 0.27 | 4,223 | 4% | 2,134 | -4% | 748 | 0% |
| 340 | 5,774 | 0.25 | 3,259 | 4% | 1,521 | -4% | 487 | 1% |
| 360 | 4,293 | 0.24 | 2,452 | 3% | 1,098 | -3% | 377 | 3% |
| 380 | 3,287 | 0.23 | 1,907 | 3% | 855 | -3% | 240 | 4% |
| 400 | 2,524 | 0.22 | 1,501 | 2% | 618 | -4% | 199 | 5% |
| 420 | 1,880 | 0.21 | 1,163 | 2% | 446 | -4% | 122 | 5% |
| 440 | 1,401 | 0.20 | 822 | 1% | 344 | -1% | 120 | 5% |
| 460 | 1,079 | 0.20 | 672 | 3% | 248 | 3% | 76 | 4% |
| 480 | 810 | 0.19 | 510 | 6% | 176 | 6% | 66 | 6% |
| 500 | 658 | 0.18 | 414 | 8% | 147 | 8% | 48 | 8% |
| 520 | 413 | 0.18 | 267 | 9% | 85 | 9% | 33 | 9% |
| 540 | 310 | 0.18 | 222 | 10% | 58 | 10% | 11 | 10% |
| 560 | 239 | 0.18 | 153 | 11% | 55 | 11% | 13 | 11% |
| 580 | 159 | 0.17 | 116 | 14% | 26 | 14% | 8 | 14% |
| 600 | 115 | 0.16 | 82 | 23% | 20 | 23% | 7 | 23% |
| Total # Points | 3,162,795 | NA | 1,195,229 | NA | 951,261 | NA | 699,218 | NA |
| Mean % Difference | NA | NA | NA | 6.8% | NA | -1.1% | NA | -1.2% |

NOTE: Column 3 presents the ARM2 ambient ratio based on the “ALL AQS Sites” data set for the complete time period 2001-2010. Columns 5, 7, and 9 present the relative percent difference between the ambient ratio calculated for the specified time period data sets versus the “ALL AQS Sites” 2001-2010 ambient ratios.

Table 3 - Ambient Ratio Comparison for Varying Distances from NO_x Emission Sources

| Upper Nox BIN ppb | All Sites # Obs | All Sites Ambient Ratios | Within 1 km of 100 tpy Sources # Obs | Within 1 km Ambient Ratio RPD | 1 to 5 km from 100 tpy Sources # Obs | 1 to 5 km Ambient Ratio RPD | Greater than 5 km from 100 tpy Sources # Obs | Greater than 5 km Ambient Ratio RPD |
|-------------------|--------------------|-----------------------------|--|-------------------------------------|---|-----------------------------------|--|--|
| 30 | 2,330,786 | 0.90 | 83,331 | 0% | 714,815 | 0% | 1,357,621 | 0% |
| 40 | 1,310,767 | 0.90 | 52,394 | 0% | 418,088 | 0% | 744,075 | 0% |
| 50 | 817,380 | 0.90 | 35,868 | 0% | 264,623 | 0% | 459,651 | 0% |
| 60 | 546,757 | 0.90 | 25,553 | 0% | 178,340 | -1% | 306,175 | 0% |
| 70 | 385,505 | 0.85 | 18,735 | -1% | 126,521 | -1% | 215,051 | 1% |
| 80 | 282,446 | 0.80 | 13,573 | 0% | 92,234 | -1% | 158,346 | 1% |
| 90 | 214,245 | 0.75 | 10,272 | 0% | 69,529 | -1% | 120,705 | 1% |
| 100 | 164,999 | 0.70 | 7,653 | -1% | 53,237 | -2% | 93,413 | 1% |
| 110 | 130,316 | 0.65 | 5,898 | -1% | 41,333 | -2% | 74,751 | 1% |
| 120 | 104,093 | 0.61 | 4,460 | -2% | 32,402 | -2% | 60,579 | 1% |
| 130 | 84,056 | 0.57 | 3,442 | -3% | 25,844 | -2% | 49,353 | 1% |
| 140 | 68,350 | 0.54 | 2,702 | -3% | 20,827 | -2% | 40,396 | 1% |
| 150 | 56,801 | 0.50 | 2,234 | -4% | 17,099 | -2% | 33,857 | 1% |
| 160 | 46,956 | 0.47 | 1,735 | -5% | 13,951 | -1% | 28,214 | 1% |
| 170 | 39,220 | 0.45 | 1,420 | -6% | 11,679 | -1% | 23,554 | 1% |
| 180 | 32,769 | 0.42 | 1,200 | -7% | 9,635 | -1% | 19,841 | 1% |
| 190 | 27,696 | 0.40 | 979 | -8% | 8,117 | -1% | 16,808 | 1% |
| 200 | 23,436 | 0.38 | 858 | -9% | 6,813 | -1% | 14,292 | 1% |
| 220 | 36,618 | 0.35 | 1,361 | -10% | 10,550 | 0% | 22,451 | 0% |
| 240 | 26,373 | 0.32 | 934 | -11% | 7,633 | 1% | 16,135 | 0% |
| 260 | 19,349 | 0.31 | 652 | -11% | 5,666 | 1% | 11,747 | 0% |
| 280 | 14,204 | 0.29 | 480 | -11% | 4,082 | 2% | 8,728 | 0% |
| 300 | 10,468 | 0.28 | 434 | -10% | 3,102 | 2% | 6,311 | 0% |
| 320 | 7,816 | 0.27 | 300 | -9% | 2,271 | 3% | 4,766 | 0% |
| 340 | 5,774 | 0.25 | 224 | -8% | 1,753 | 2% | 3,414 | -1% |
| 360 | 4,293 | 0.24 | 169 | -7% | 1,311 | 2% | 2,532 | -1% |
| 380 | 3,287 | 0.23 | 129 | -6% | 967 | 2% | 1,992 | -1% |
| 400 | 2,524 | 0.22 | 118 | -5% | 776 | 1% | 1,482 | -2% |
| 420 | 1,880 | 0.21 | 103 | -5% | 599 | 1% | 1,080 | -2% |
| 440 | 1,401 | 0.20 | 65 | -1% | 430 | 0% | 825 | -1% |
| 460 | 1,079 | 0.20 | 57 | 3% | 354 | 3% | 605 | 3% |
| 480 | 810 | 0.19 | 51 | 6% | 245 | 6% | 482 | 6% |
| 500 | 658 | 0.18 | 45 | 8% | 218 | 8% | 364 | 8% |
| 520 | 413 | 0.18 | 25 | 9% | 128 | 9% | 242 | 9% |
| 540 | 310 | 0.18 | 13 | 10% | 83 | 10% | 199 | 10% |
| 560 | 239 | 0.18 | 19 | 11% | 80 | 11% | 132 | 11% |
| 580 | 159 | 0.17 | 5 | 14% | 47 | 14% | 97 | 14% |
| 600 | 115 | 0.16 | 7 | 23% | 33 | 23% | 71 | 23% |
| Total # Points | 3,162,795 | NA | 141,773 | NA | 1,012,507 | NA | 1,798,641 | NA |
| Mean % Difference | NA | NA | NA | -1.7% | NA | 2.3% | NA | 2.6% |

NOTE: Column 3 presents the ARM2 ambient ratio based on the “ALL AQS Sites” data set. Columns 5, 7, and 9 present the relative percent difference between the ambient ratio calculated for the specified “distance from source” data sets versus the “ALL AQS Sites” ambient ratios.

3.3 Sensitivity of Ambient Ratio to Percentile Selection

The ARM2 method is based on the 98th percentile ambient ratios for the entire AQS data base. Note that the use of the 98th percentile is not related to the form of the 1-hr NO₂ NAAQS. EPA's OAQPS Air Quality Modeling Group requested a sensitivity analysis for the use of different percentile values for establishing the ARM2 ambient ratios. Table 4 presents the results of the sensitivity analysis on the use of the 99th and 95th percentiles versus the 98th percentile, all based on the "All AQS Sites" main data set. As expected, the 99th percentile ARM2 equation results in higher ARM2 calculated ambient ratios (mean RPD of +9%), while the 95th percentile ARM2 equation results in lower ARM2 calculated ambient ratios (mean RPD of -9%).

The selection of the percentile value used to represent the upper bound of the observed ambient ratios in each NO_x concentration bin is arbitrary. The 98th percentile rather than the maximum value was used because this value is more stable and not subject to extreme variations that could be associated with artifacts in the data, including calibrations, audits, span checks, power restarts, or other analyzer operations that were not screened from the AQS data base. Using the 98th percentile removes these potentially spurious data points, while still providing a robust upper bound estimate of the ambient ratio for each bin. The original ARM method used the 90th percentile to determine the conversion factor, and the ARM2 method uses a more conservative 98th percentile.

Based on consideration of the above factors, the 98th percentile value has been selected as a reasonable estimator of the upper bound ambient ratios, and is the basis for the performance evaluations of the ARM2 method.

Table 4 - Sensitivity of ARM2 Ambient Ratios to the Percentile Value

| Upper Nox BIN ppb | ALL Sites # Obs | Ratios for 98th Percentile | Ratios for 99th Percentile | RPD | Ratios for 95th Percentile | RPD |
|-------------------|--------------------|-------------------------------|-------------------------------|------|-------------------------------|-------|
| 30 | 2,330,786 | 0.90 | 0.90 | 0% | 0.90 | 0% |
| 40 | 1,310,767 | 0.90 | 0.90 | 0% | 0.90 | 0% |
| 50 | 817,380 | 0.90 | 0.90 | 0% | 0.90 | 0% |
| 60 | 546,757 | 0.90 | 0.90 | 0% | 0.84 | -6% |
| 70 | 385,505 | 0.85 | 0.90 | 5% | 0.78 | -9% |
| 80 | 282,446 | 0.80 | 0.85 | 7% | 0.72 | -10% |
| 90 | 214,245 | 0.75 | 0.80 | 8% | 0.66 | -11% |
| 100 | 164,999 | 0.70 | 0.76 | 9% | 0.61 | -12% |
| 110 | 130,316 | 0.65 | 0.72 | 10% | 0.57 | -13% |
| 120 | 104,093 | 0.61 | 0.68 | 11% | 0.53 | -14% |
| 130 | 84,056 | 0.57 | 0.64 | 11% | 0.49 | -14% |
| 140 | 68,350 | 0.54 | 0.60 | 12% | 0.46 | -15% |
| 150 | 56,801 | 0.50 | 0.56 | 12% | 0.43 | -15% |
| 160 | 46,956 | 0.47 | 0.53 | 12% | 0.40 | -15% |
| 170 | 39,220 | 0.45 | 0.50 | 13% | 0.38 | -15% |
| 180 | 32,769 | 0.42 | 0.48 | 13% | 0.36 | -15% |
| 190 | 27,696 | 0.40 | 0.45 | 13% | 0.34 | -15% |
| 200 | 23,436 | 0.38 | 0.43 | 13% | 0.33 | -14% |
| 220 | 36,618 | 0.35 | 0.39 | 12% | 0.30 | -14% |
| 240 | 26,373 | 0.32 | 0.36 | 12% | 0.28 | -13% |
| 260 | 19,349 | 0.31 | 0.34 | 12% | 0.27 | -13% |
| 280 | 14,204 | 0.29 | 0.32 | 12% | 0.25 | -13% |
| 300 | 10,468 | 0.28 | 0.31 | 12% | 0.24 | -12% |
| 320 | 7,816 | 0.27 | 0.30 | 12% | 0.23 | -12% |
| 340 | 5,774 | 0.25 | 0.29 | 12% | 0.22 | -12% |
| 360 | 4,293 | 0.24 | 0.27 | 12% | 0.22 | -12% |
| 380 | 3,287 | 0.23 | 0.26 | 13% | 0.21 | -12% |
| 400 | 2,524 | 0.22 | 0.25 | 13% | 0.20 | -10% |
| 420 | 1,880 | 0.21 | 0.24 | 14% | 0.20 | -6% |
| 440 | 1,401 | 0.20 | 0.23 | 14% | 0.20 | -1% |
| 460 | 1,079 | 0.20 | 0.22 | 11% | 0.20 | 0% |
| 480 | 810 | 0.20 | 0.22 | 8% | 0.20 | 0% |
| 500 | 658 | 0.20 | 0.21 | 5% | 0.20 | 0% |
| 520 | 413 | 0.20 | 0.21 | 3% | 0.20 | 0% |
| 540 | 310 | 0.20 | 0.20 | 1% | 0.20 | 0% |
| 560 | 239 | 0.20 | 0.20 | 0% | 0.20 | 0% |
| 580 | 159 | 0.20 | 0.20 | 0% | 0.20 | 0% |
| 600 | 115 | 0.20 | 0.20 | 0% | 0.20 | 0% |
| Mean % Difference | NA | NA | NA | 9.0% | NA | -9.0% |

3.4 ARM2 Implementation into AERMOD

The “All AQS Sites” ARM2 ambient ratio equation was coded directly into the AERMOD model. Because the monitoring data are expressed in ppb units but modeled concentrations are in $\mu\text{g}/\text{m}^3$ units, the ARM2 equation coefficients were recalculated for use with $\mu\text{g}/\text{m}^3$ NO_x concentration inputs. Refer to Section 4.2 of this report for additional details on the method for converting from ppb to $\mu\text{g}/\text{m}^3$ units. The final ARM2 ambient ratio equation in units of ug/m^3 is:

$$y = -1.1723\text{E-}17x^6 + 4.2795\text{E-}14x^5 - 5.8345\text{E-}11x^4 + 3.4555\text{E-}08x^3 - 5.6062\text{E-}06x^2 - 2.7383\text{E-}03x + 1.2441\text{E+}00,$$

Where x is the NO_x concentration in ug/m^3 , and y is the calculated ambient ratio.

The Fortran source code for the AERMOD dispersion model was edited to include the ARM2 method as a model option using the keyword “ARM2”. Under this option, AERMOD calculates the cumulative 1-hr NO_x concentration (i.e., based on the source group “ALL”) at each receptor on an hour-by-hour basis. Each hourly NO_x concentration for the “ALL” source group is input to the ARM2 equation to determine the ARM2 ambient ratio for that hour. The ARM2 ambient ratio for that hour is then multiplied by the NO_x concentration for any other source groups in the model run to determine that hour’s NO_2 concentration for each source group. The hourly NO_2 concentrations for each receptor and source group are then stored as usual into the appropriate AERMOD data array for subsequent output processing. The ARM2 maximum and minimum ratios are set by default to 0.9 and 0.2; however, the NO2EQUIL and NO2STACK keywords (typically used for the PVMRM method) can be used to change these default settings.

The ARM2 coding changes in AERMOD have been highlighted with comments that include the word “ARM2”. Since the ARM2 ambient ratio is determined from the cumulative modeled NO_x concentration, there must be an “ALL” source group in the AERMOD run (if an “ALL” source group is not included in a run, AERMOD halts with an error message). User instructions for the AERMOD-ARM2 option are presented in Appendix C.

4.0 ARM2 Performance Evaluation

4.1 Evaluation Data Sets

The performance of the ARM2 method was evaluated with three test data sets that have previously been used to evaluate PVMRM and OLM (Hanrahan¹³, Brode¹⁴, and Hanna¹⁵ et. al.). It is important to note that there are significant uncertainties and assumptions in the emission inventories for all of these data sets. The Empire Abo and Palaau data sets assumed a fixed emission rate over the entire time period evaluated, while the Wainwright data set contains hourly estimates of power station load which were used to scale vendor data emission tests to estimate hourly emissions. The evaluation data sets are described below:

Empire Abo, New Mexico Data Set – The operators of the Empire Abo Gas plant in southeastern New Mexico collected ozone, NO_x, and meteorological data at two sites for a two year period (June 1993 through June 1995). The objective of this program was to collect representative ozone monitoring data not influenced by NO scavenging for use in NO₂ modeling with the OLM method. The North monitoring site was located 1.6 km north of the plant with a direct transport wind direction from the plant to the monitor of 220 degrees, and the South monitoring site was located 2.5 km to the south of the facility with a direct transport wind direction of 340 degrees.

¹³ Hanrahan, P.L., 1999b. "The plume volume molar ratio method for determining NO₂/NO_x ratios in modeling. Part II: Evaluation Studies," J. Air & Waste Manage. Assoc., 49, 1332- 1338.

¹⁴ "Evaluation of Bias in AERMOD-PVMRM", Alaska DEC Contract No. 18-9010-12, MACTEC Report, June 2005, and "Additional Clarification Regarding Application of Appendix W Modeling Guidance for the 1-hour NO₂ NAAQS", dated March 11, 2011.

¹⁵ "PVMRM and OLM Evaluation using a New Data Set", Steven Hanna, Elizabeth Hendrick, Vincent Tino, and Bruce Egan, presented at the 10th EPA Air Modeling Conference on March 14, 2012.

The Empire Abo Gas plant consists of 15 compressor engines, five boilers, two heaters, a sulfur recovery unit flare, an inlet flare, an acid gas flare, and a diesel fire pump. Typical stack heights are approximately 30 feet, and there are squat buildings with approximate heights of 20 feet, so downwash is a factor at this facility. Of the 26 emission sources, there is no one dominant source of emissions but the compressor engines are the largest type of emission source. There are numerous other emission sources in the area from gas wells, compressors, and gas treatment plants, but these are more distant from the ambient monitors and were not included in the modeling input files.

EPA provided the Empire Abo AERMOD modeling files and ambient monitoring data for the period June 1993 through June 1994 that were used for the PVMRM and OLM 1-hr NO₂ performance evaluations reported in EPA's 2011 Additional Clarifications memo. The AERMOD input files included source and stack parameters, building parameters as calculated by the BPIP-PRIME program, and the model receptor location. The AERMET meteorological data was processed by EPA using AERMET version 11059. The ambient NO_x and ozone monitoring data were also provided. The AERMOD files include input files for both the North and South monitor locations. The meteorological data from the 10-meter tower located at the North monitor were used for both sets of model runs, as these data were representative for both the source and the two monitoring site locations.

Palaau, Hawaii Data Set – NO_x, ozone, and meteorological data were collected at a site near the Palaau Generating Station on the island of Molokai. The facility consisted of four diesel-engine generators and an oil-fired combustion turbine. The engine and CT stack heights range from 25 to 32 feet tall. The building heights affecting these stacks ranged from 25 feet to 70 feet high, so downwash is a factor at this facility. In contrast to the New Mexico monitors, this ambient monitor was located closer to the source (220 m distant), with a direct transport wind direction of 120 degrees. Data were collected from January 1, 1993 through December 31, 1993.

EPA provided the Palaau AERMOD modeling files and ambient monitoring data that were used for the PVMRM and OLM 1-hr NO₂ performance evaluations reported in EPA's 2011 Additional Clarifications memo. The AERMOD input files included source and stack parameters, building parameters as calculated by the BPIP-PRIME program, and the model receptor location. The AERMET meteorological data was processed using AERMET version 11059. The ambient NO_x and ozone monitoring data were also provided.

Wainwright, Alaska Data Set - This data set consists of hourly observations of ozone, NO_x, and meteorological data from September 16, 2009, through September 30, 2010, from an ambient monitoring station located near a power plant site in Wainwright, Alaska. The power plant consists of five diesel generators (rated from 425 to 950kW), each vented through its own stack, all located in a single building. The stack heights are about the same as the building height at 30 feet tall. There were few other sources of NO_x emissions in the area. The monitoring station is located about 500 m to the east-south-east of the plant (the direct transport wind direction is 280 degrees). By focusing this data set on those hours when the wind direction would transport plant emission to the monitor (defined as a 60 degree sector centered on 280 degrees), a total of 594 hours of data were available for the performance evaluation data set. Hourly surface meteorological data from the ambient monitoring station and data from the nearby NWS ASOS station (including 1-minute ASOS wind data) were used with AERMET version 11059 to create the meteorological data input files for AERMOD.

The Wainwright data set contains hourly estimates of power station load which were used to scale vendor data emission tests to estimate hourly emissions. Modeling results and monitoring data for the Wainwright test data set were provided in a spreadsheet format by Epsilon Associates, Inc., who performed the AERMOD (version 11103) modeling for the "PVMRM and OLM Evaluation using a New Data Set" study presented at the EPA 10th Modeling Conference in March 2012. The total modeled NO_x concentrations reported in the spreadsheet were input to the ARM2 equation to calculate ARM2 NO₂ concentrations, which were then compared to the Tier 3 concentrations and ambient monitoring data compiled by Epsilon.

4.2 Updates to Previously Used In-stack Ratios and Ozone Data

The in-stack NO_2/NO_x ratio values previously used in the Empire Abo and Palaau evaluations were updated to reflect current guidance. The value originally used by Hanrahan and Brode was 0.1 for all modeled sources, even though the majority of the larger emission sources in these data sets are internal combustion engines. As discussed in Section 3.1 of this report, EPA is currently developing a database of in-stack NO_2/NO_x ratios, and in the interim some state agencies have developed in-stack ratio data and guidance. The in-stack ratio generally recommended for boilers, combustion turbines, and IC engines ranges from approximately 0.1 to 0.2. Therefore, the in-stack ratio used for the Empire Abo and Palaau Tier 3 evaluations performed for this study was revised from the original 0.1 value to 0.2 (the Wainwright modeling evaluations also used an in-stack ratio of 0.2). This is much less conservative than EPA's recommended default in-stack ratio of 0.5, and if the default in-stack ratio were to be used in this study the NO_2/NO_x ratios predicted by the PVMRM and OLM methods would be substantially higher than what is reported herein.

The ozone data that is input for the Tier 3 analyses must be representative of background ozone concentrations, without "ozone scavenging" from local NO_x emission sources that bias the ozone data low. For the Empire Abo data set, ozone data were collected at two stations. For periods when NO_x plumes are impacting one of the monitor stations, the plumes are also causing "scavenging" or "consumption" of ozone at that site that lowers the ambient ozone concentrations. Therefore, to minimize the effects of ozone scavenging, the higher of the North or South ozone hourly averages was used to create the "maximum hourly ozone" data file for the Empire Abo data sets that was used for subsequent Tier 3 modeling. In addition, the relative few hours of missing ozone data were filled in with the maximum hourly observation of 74.7 ppb, using the OZONEVAL setting in AERMOD.

These two updates reflect current in-stack ratio guidance and account for ozone “scavenging”, and will likely result in higher predicted NO₂/NO_x ratios for the Tier 3 methods than have been reported in the 2011 Additional Clarifications memo.

4.3 Converting Monitoring Data from ppb to µg/m³ Units

The NO_x ambient monitoring data is typically reported in units of ppb by volume, while the air quality dispersion models output NO and NO₂ concentrations in units of µg/m³. Therefore, to compare modeled concentrations to monitoring data, it is necessary to convert the monitored data from ppb to µg/m³ units. The New Mexico Environmental Department (NMED) Air Dispersion Modeling Guidelines provide a good discussion of conversion equations and the need to account for site pressure and temperature. These guidelines note that the convention for the NAAQS is defined at 40 CFR 50.3 “Reference conditions”, which states that air quality measured in units of mass per unit volume shall be corrected to STP conditions of 25 degrees C and 760 millimeters of mercury. In addition, the California Air Pollution Control Officers Association (CAPCOA) has developed the guidance document “Modeling Compliance of the Federal 1-Hour NO₂ NAAQS”, and section 4 discusses how background NO₂ data should be converted from ppb to µg/m³ units. Like the NMED guidance, CAPCOA recommends that the conversion be performed assuming STP conditions, rather than converting to µg/m³ in “actual temperature and pressure units” based on monitoring site conditions. Therefore, the following equation was used to convert the monitored NO/NO₂/NO_x concentrations from ppb to µg/m³ at STP:

$$C = (\text{ppb}/1000) \times \text{MW} \times 40.8727,$$

where MW is the molecular weight of the pollutant in grams/mole.

For NO_x, the two species are NO and NO₂, which have different molecular weights (30 and 46 grams/mole, respectively). However, it must be noted that the convention for the NO_x mass emission rate input to dispersion models is “as NO₂” (i.e., using the molecular weight of NO₂), and the model predicted NO_x concentrations do not weight the relative percentages and molecular

weights of the NO and NO₂ species. Therefore, the ambient NO, NO₂, and NO_x concentrations should be converted using the molecular weight for NO₂ (the same as how the model treats the various species). The STP conversion equation for both NO₂ and NO_x then is simplified to $\mu\text{g}/\text{m}^3 = \text{ppb} * 1.88$.

This ambient data unit conversion method differs from the “actual $\mu\text{g}/\text{m}^3$ ” method previously used in both EPA’s 1-hr NO₂ performance evaluation presented in the 2011 Additional Clarifications memo, and in the “Updated Tier 2 Ambient Ratio Method (ARM) for 1-hr NO₂ NAAQS Analyses” presentation at EPA’s 10th Modeling Conference. Those analyses used site temperature and pressure in the ppb to $\mu\text{g}/\text{m}^3$ conversion equation. For the Empire Abo site located at approximately 3500 foot elevation, the “standard $\mu\text{g}/\text{m}^3$ ” monitored concentrations are about 10% higher than the “actual $\mu\text{g}/\text{m}^3$ ” concentrations, which has implications on the previously reported model performance at Empire Abo. For the Palaau site near sea level, there is little difference between the “standard $\mu\text{g}/\text{m}^3$ ” and “actual $\mu\text{g}/\text{m}^3$ ”, and the Wainwright analysis has always used “standard $\mu\text{g}/\text{m}^3$ ” monitored concentrations.

4.4 Evaluation Procedures

When evaluating the various NO₂ model conversion options, it is important to recall that the modules for predicting the NO₂/NO_x ratio (ARM2, PVMRM, and OLM) are only one part of the modeling system. The agreement between the prediction of NO₂ versus actual observations is a result of both AERMOD’s ability to estimate the NO_x concentration, as well as the ability to estimate the NO₂/NO_x ambient ratio. For example, PVMRM may be accurately predicting the NO₂/NO_x ratio, but the overall AERMOD NO₂ predictions might be poor because of AERMOD performance. Conversely, PVMRM may be poorly predicting the ambient ratio, but compensating errors in AERMOD’s performance may result in good overall agreement with NO₂ observations. Therefore, the evaluation procedures look at both the performance of the methods for predicting the ambient ratio, as well as the overall ability of the modeling system (AERMOD NO_x predictions and ambient ratio predictions).

The source inputs, building downwash parameters, and meteorological data for each evaluation data set were input to AERMOD (version 12345) to calculate hourly NO_x concentrations at a single receptor located at the respective ambient monitor location. In addition, the AERMOD-ARM2, AERMOD-PVMRM, and AERMOD-OLM methods were also used to calculate hourly NO₂ concentrations (all reported OLM concentrations are based on the Group All option). The hourly model predicted NO₂ and NO_x concentrations were input to a spreadsheet, along with the concurrent observations of hourly NO₂, NO_x, wind speed, and wind direction.

The data set was first sorted by wind direction, and hours were selected when the wind direction was within a 60 degree sector centered on the direct transport wind direction from the modeled source to the monitoring station. This focused the analysis on direct impacts from the source on the ambient monitor, and helped reduced the potential effects of other background sources on the monitoring data. A minimum or “threshold” NO_x concentration of 20 µg/m³ (approximately 10 ppb) was then applied to both the monitoring and modeling data. This was done to remove the measurement variability at low monitored concentrations, and to reduce the potential effects of other background sources at low monitored concentrations.

Scatter plots were prepared of the model predicted versus observed NO₂/NO_x ratios, paired in time (the data is already paired in space since a single receptor at the monitor location is used in the modeling). These plots were reviewed to evaluate the amount of variability in the model predicted ratios, as well as any overall bias in the predictions.

Scatter plots were also prepared for model predictions of the NO₂/NO_x ratio as a function of model predicted NO_x concentration. The observed NO₂/NO_x ratio as a function of observed NO_x concentration was also plotted on these graphs (note in these plots, the modeled and monitored data points are not paired in time). These plots help evaluate the agreement between model predicted and monitored ratios as a function of NO_x concentration.

Quantile-Quantile (Q-Q) plots of model predicted NO₂ concentrations versus observations were prepared to evaluate the overall ability of the modeling methods to match the frequency distribution of the NO₂ observations, especially in the high range concentrations near the NAAQS.

The model predicted hourly NO₂ and NO_x concentrations for each of the test data sets were also compared to observations using the Robust Highest Concentration (RHC). The RHC represents a smoothed estimate of the highest concentrations, based on a tail exponential fit to the upper end of the concentration distribution. This reduces the effect of extreme values on the comparison of model predicted concentrations to ambient monitoring concentrations. The RHC is calculated as follows:

$$\text{RHC} = \chi\{n\} + (\chi - \chi\{n\}) \ln((3n-1)/2)$$

where $n = 26$,

χ is the average of the $n - 1$ largest values, and

$\chi\{n\}$ is the n th largest value.

In the Q-Q and RHC performance evaluations, the combined performance of AERMOD and the conversion methods is evaluated against NO₂ monitoring data. To help determine the performance of AERMOD itself, and to identify potential compensating errors between the dispersion model and the NO₂ conversion methodology, RHC statistics are also presented for the AERMOD predicted NO_x concentration versus monitored NO_x observations (without any conversion ratios applied).

4.5 Empire Abo Evaluation Results

For the North monitoring site data set, there were 2137 hours that met the “direct transport” wind direction criteria. There were 734 of those hours that also had monitored NO_x concentrations greater than the 20 µg/m³ threshold, and 329 hours with both monitored and modeled NO_x concentrations greater than the threshold. For the 329 hour data set, the maximum monitored NO_x concentration was 561 µg/m³, and the maximum monitored NO₂ concentration was 125 µg/m³.

For the South monitoring site data set, there were 1206 hours that met the “direct transport” wind direction criteria. There were 376 of those hours that also had monitored NO_x concentrations greater than the 20 µg/m³ threshold, and 128 hours with both monitored and modeled NO_x concentrations greater than the threshold. For the 128 hour data set, the maximum monitored NO_x concentration was 388 µg/m³, and the maximum monitored NO₂ concentration was 59 µg/m³.

The scatter plots of predicted versus observed NO₂/NO_x ratios, paired on an hour-by-hour basis, are presented in Figures 6 and 7 for the Empire Abo North and South data sets. These plots indicate that there is a wide range in predicted and observed ratios and a large amount of scatter for all of the methods evaluated (ARM2, PVMRM, and OLM). The PVMRM data points, and OLM data points for the South data set, are roughly evenly distributed on either side of the 1:1 line. The OLM data points for the North data set, and the ARM2 data points from both data sets, are predominately distributed above the 1:1 line. This indicates that OLM and ARM2 have a greater tendency towards overprediction of the ambient ratio than PVMRM. However, the large amount of scatter in these paired data plots suggests that none of the methods have much skill in predicting the ratios on a paired in space and time basis.

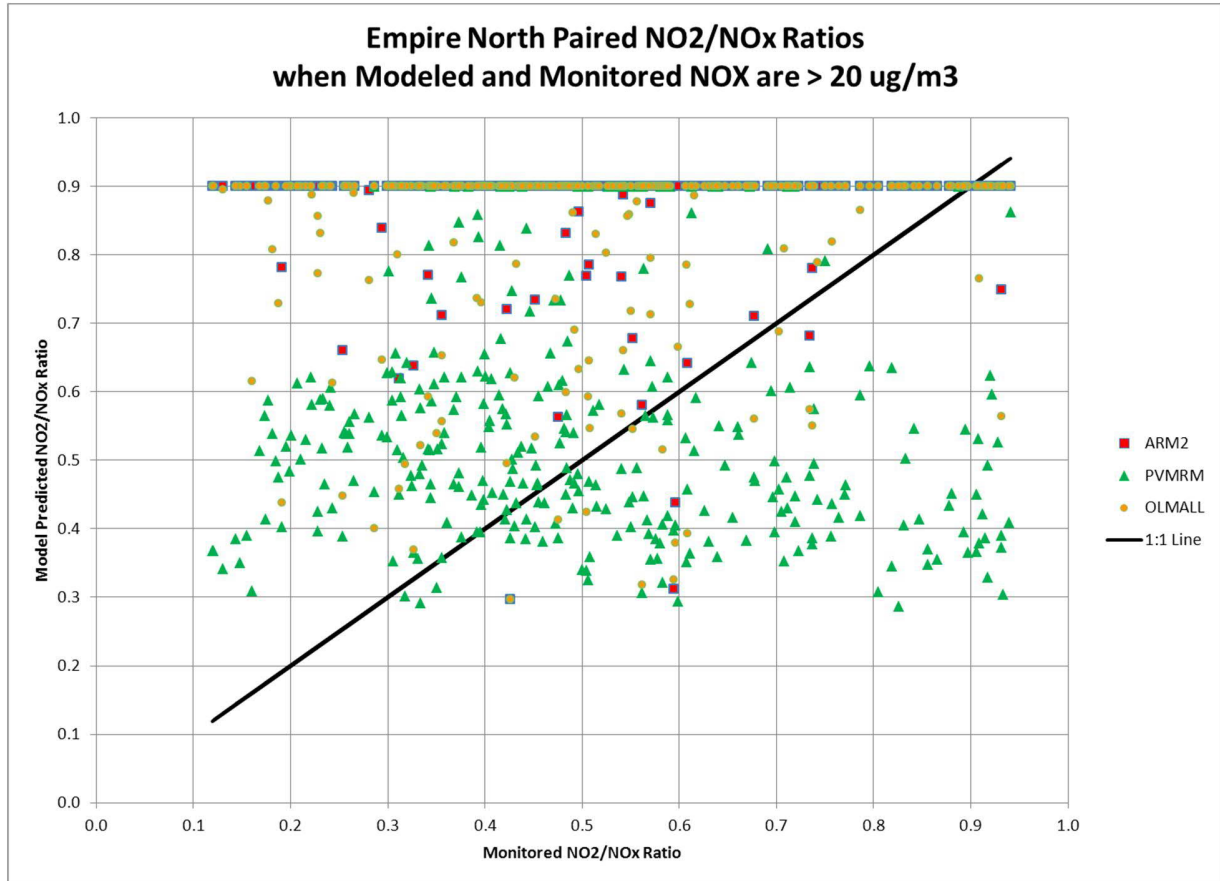
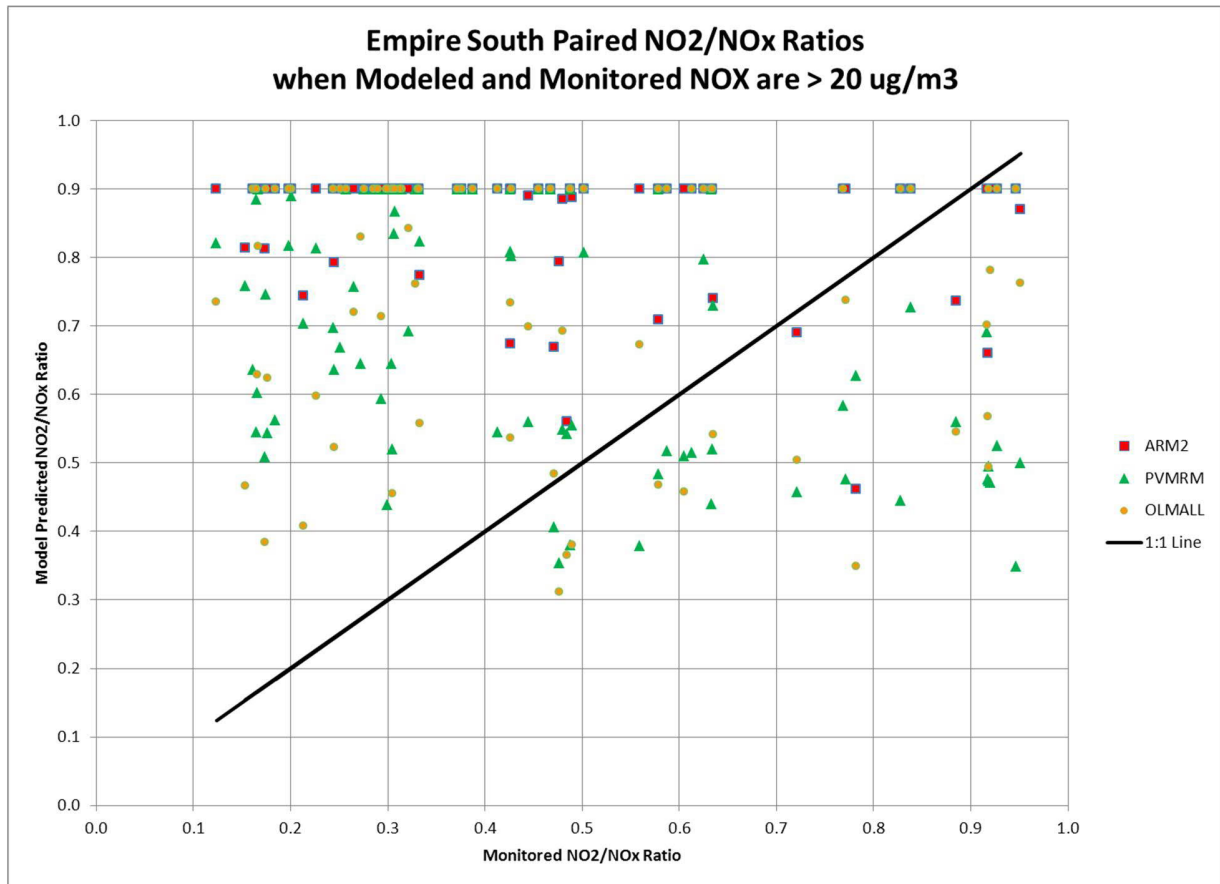
Figure 6 – Scatter Plot of NO₂/NO_x Ratio for Empire Abo North Data Set

Figure 7 – Scatter Plot of NO₂/NO_x Ratio for Empire Abo South Data Set

Plots of the predicted NO_2/NO_x ratio as a function of AERMOD predicted NO_x concentration were prepared. The observed NO_2/NO_x ratios plotted as a function of observed NO_x concentrations are also included in these plots (the monitoring data are not paired in time with the modeling predictions). Figures 8 and 9 present the plots for the Empire Abo North and South data sets, respectively. The North site plot indicates that the PVMRM predicted ratios generally match the observed ratios the best, and the OLM predicted ratios are typically higher than the observed ratios. The ARM2 predicted ratios follow the “ARM2 curve” equation, and generally are the highest estimate of the NO_2/NO_x ratio. However, at the highest NO_x concentrations, the ratios predicted by the three methods exhibit different behavior, with the PVMRM predicting the highest ratios followed by OLM and ARM2. Importantly, all three methods are overpredicting the observed NO_2/NO_x ratios at the higher NO_x concentrations. The South site plot exhibits these same general relationships. The significant finding of this portion of the analysis is that for the elevated NO_x concentrations, the monitoring data indicate the ambient ratios are less than 0.2, while the three methods predict ambient ratios ranging from approximately 0.3 to 0.6. Thus, at the upper end of the NO_x frequency distribution, the methods are not able to accurately estimate NO_2 conversion.

Figures 10 and 11 present the hourly unpaired Q-Q plots for NO_2 for the Empire North and South monitor data sets, respectively. These data were for those hours that met the “direct transport” wind direction criteria and had monitored or modeled concentrations greater than $1 \mu\text{g}/\text{m}^3$. These results show the conservatism of the full conversion option, with a significant bias to over-predict hourly NO_2 concentrations. The AERMOD-ARM2 curve indicates significant over-prediction at lower concentrations, but the Q-Q plot then flattens out and results in good performance at the extreme high concentrations (this flattened shape is a result of the power function shape of the ARM2 ambient ratio curve). The AERMOD-OLM method exhibits better performance than the other Tier 3 method, AERMOD-PVMRM. The AERMOD-ARM2 combination performs as well as, if not better, than the Tier 3 methods for the Empire Abo data sets. All methods are over-predicting at the highest NO_2 concentrations when compared to monitoring data.

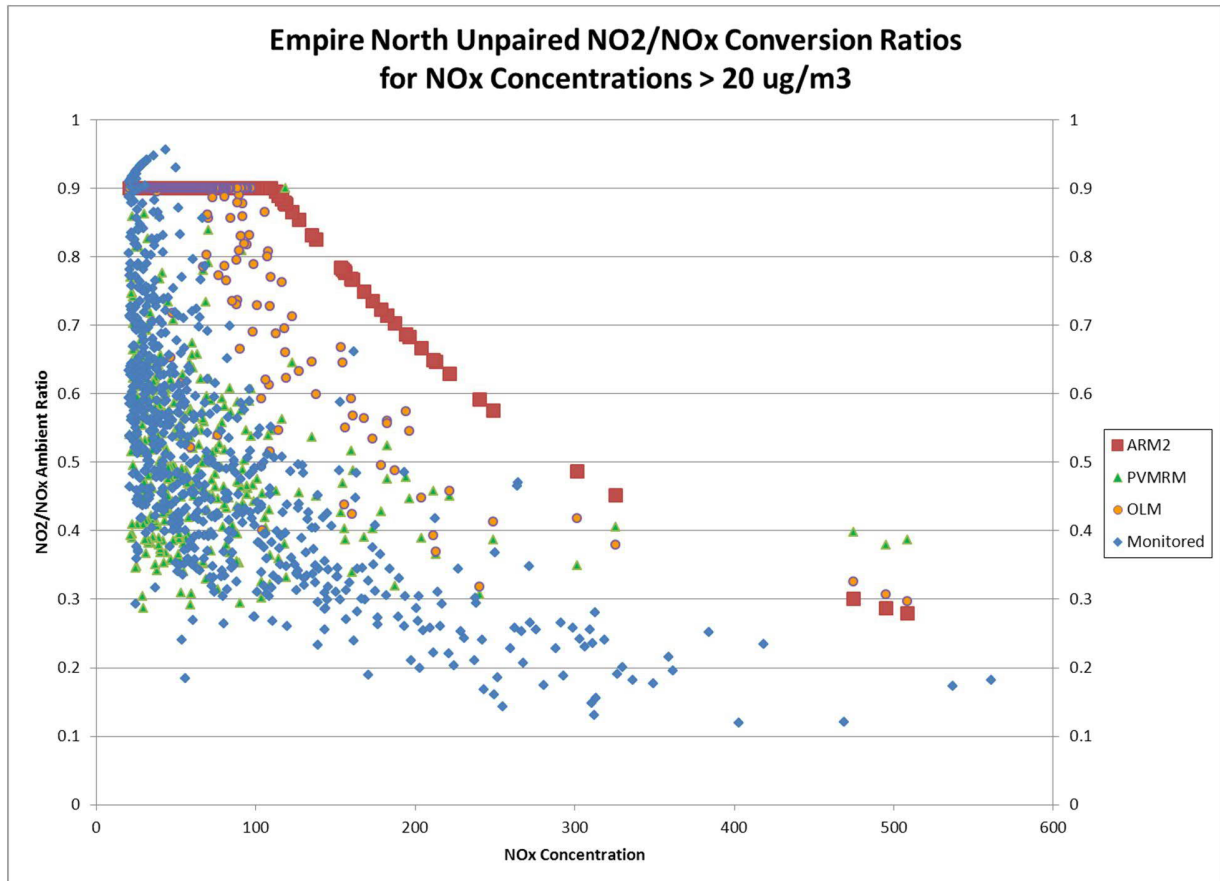
Figure 8 –Plot of NO₂/NO_x Ratio vs NO_x for Empire Abo North Data Set

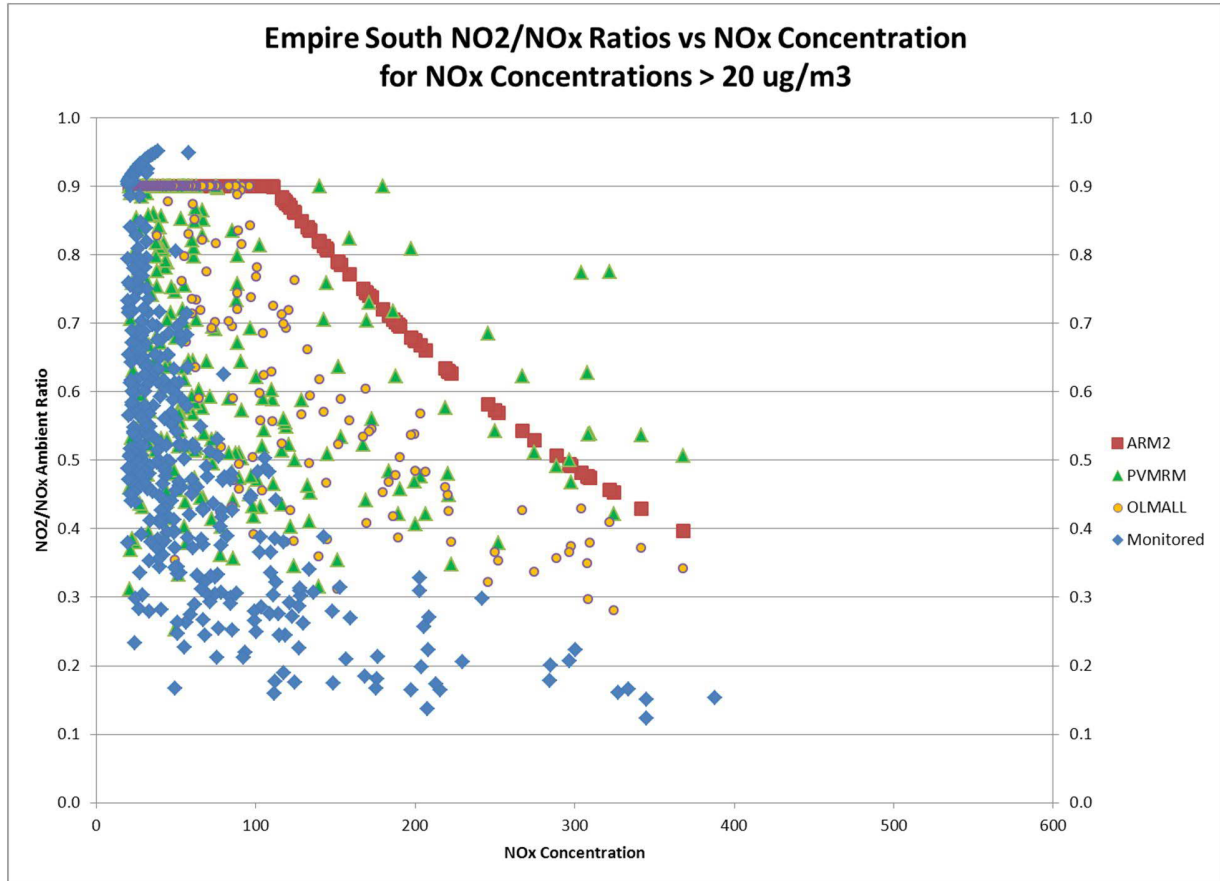
Figure 9 –Plot of NO₂/NO_x Ratio vs NO_x for Empire Abo South Data Set

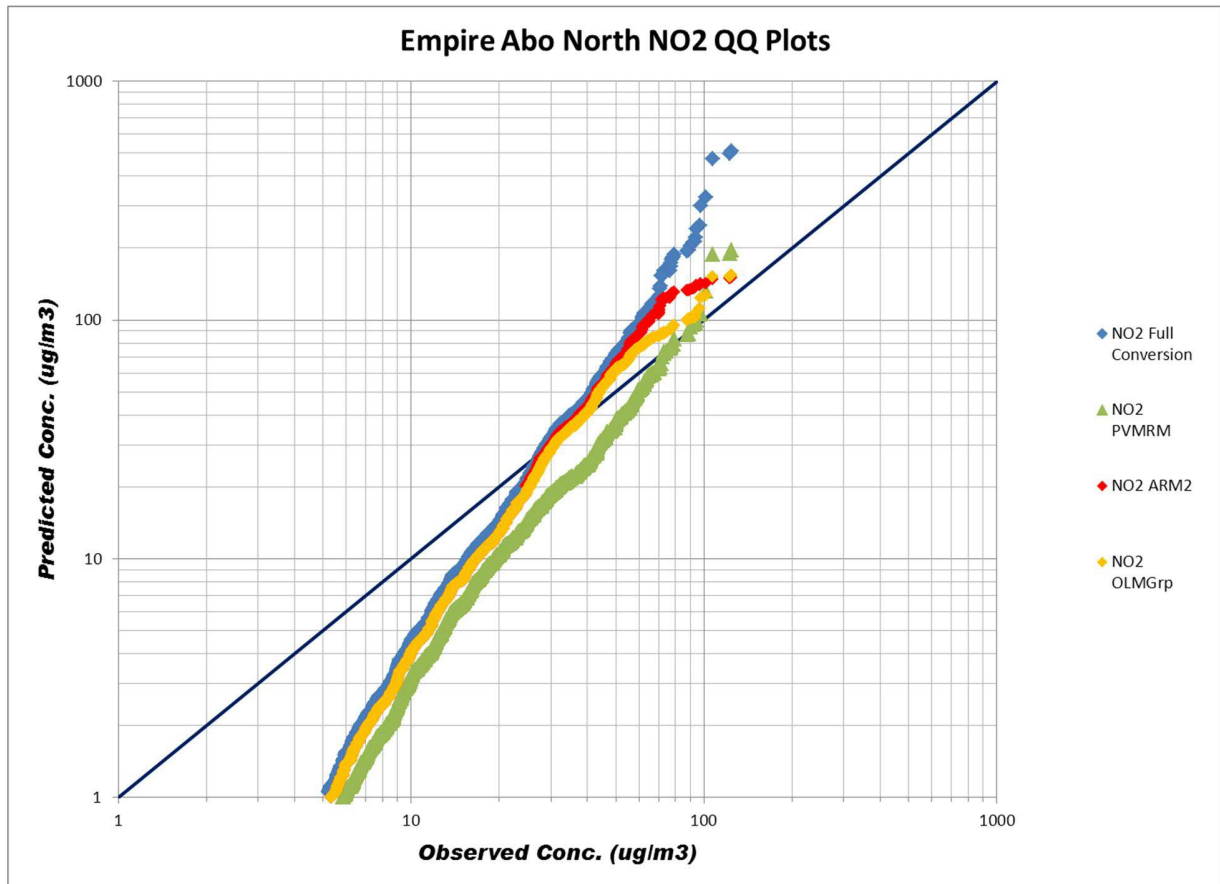
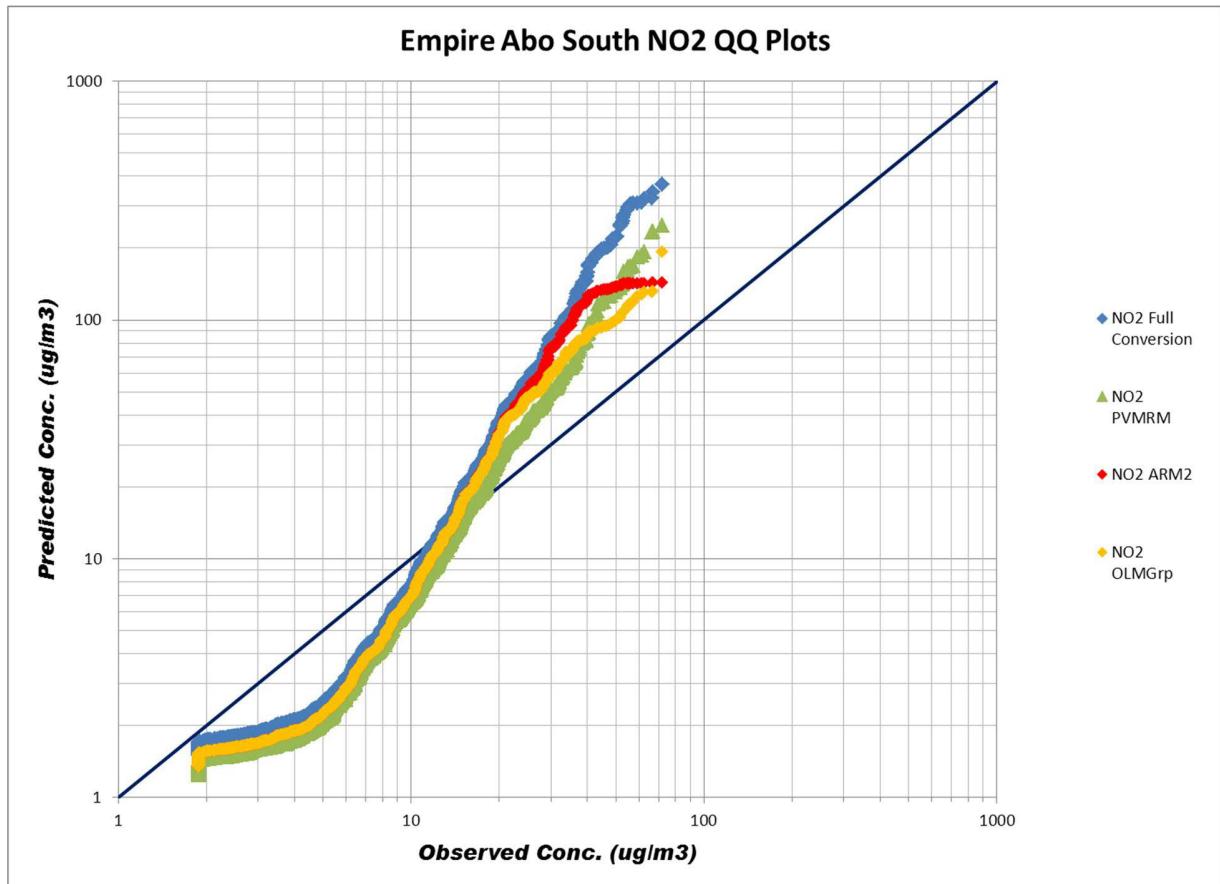
Figure 10 – Q-Q NO₂ Plots for the Empire North Data Set

Figure 11 – Q-Q NO₂ Plots for the Empire South Data Set

4.6 Palaau Evaluation Results

There were 699 hours that met the “direct transport” wind direction criteria. There were 466 of those hours that also had monitored NO_x concentrations greater than the $20 \mu\text{g}/\text{m}^3$ threshold, and 441 hours with both monitored and modeled NO_x concentrations greater than the $20 \mu\text{g}/\text{m}^3$ threshold. For the 441 hour data set, the maximum monitored NO_x concentration was $642 \mu\text{g}/\text{m}^3$, and the maximum monitored NO_2 concentration was $85 \mu\text{g}/\text{m}^3$.

The scatter plot of predicted versus observed NO_2/NO_x ratios, paired on an hour-by-hour basis, is presented in Figure 12. This plot indicates that the observed ratios are clustered between 0.1 and 0.3. The PVMRM predicted ratios are clustered between 0.2 and 0.3 and are distributed above the 1:1 line. The OLM predicted ratios are found between 0.2 and 0.9, and the ARM2 predicted ratios are found between approximately 0.4 and 0.9. A large amount of scatter is observed for the OLM and ARM2 methods, but the scatter for PVMRM is appreciably less. These plots suggest that PVMRM has better skill than OLM or ARM2 for predicting the ratios on a paired in space and time basis for this data set, and that OLM and ARM2 have a greater tendency towards overprediction of the ambient ratio than does PVMRM.

Figure 13 is the scatter plot of the predicted NO_2/NO_x ratio as a function of AERMOD predicted NO_x concentration. The observed NO_2/NO_x ratios plotted as a function of observed NO_x concentrations are also included (the monitoring data are not paired in time with the modeling predictions). This plot indicates that the PVMRM predicted ratios generally match the observed ratios the best, followed by OLM predictions, and the ARM2 predicted ratios are generally the highest estimate of the NO_2/NO_x ratio. At the extreme high end of the observed NO_x concentrations, the predicted ratios from the three methods cluster together in the range 0.25 to 0.35, while the observed ambient ratios are all less than 0.2. Therefore, at the upper end of the NO_x frequency distribution, all the methods are overestimating the NO_2 conversion.

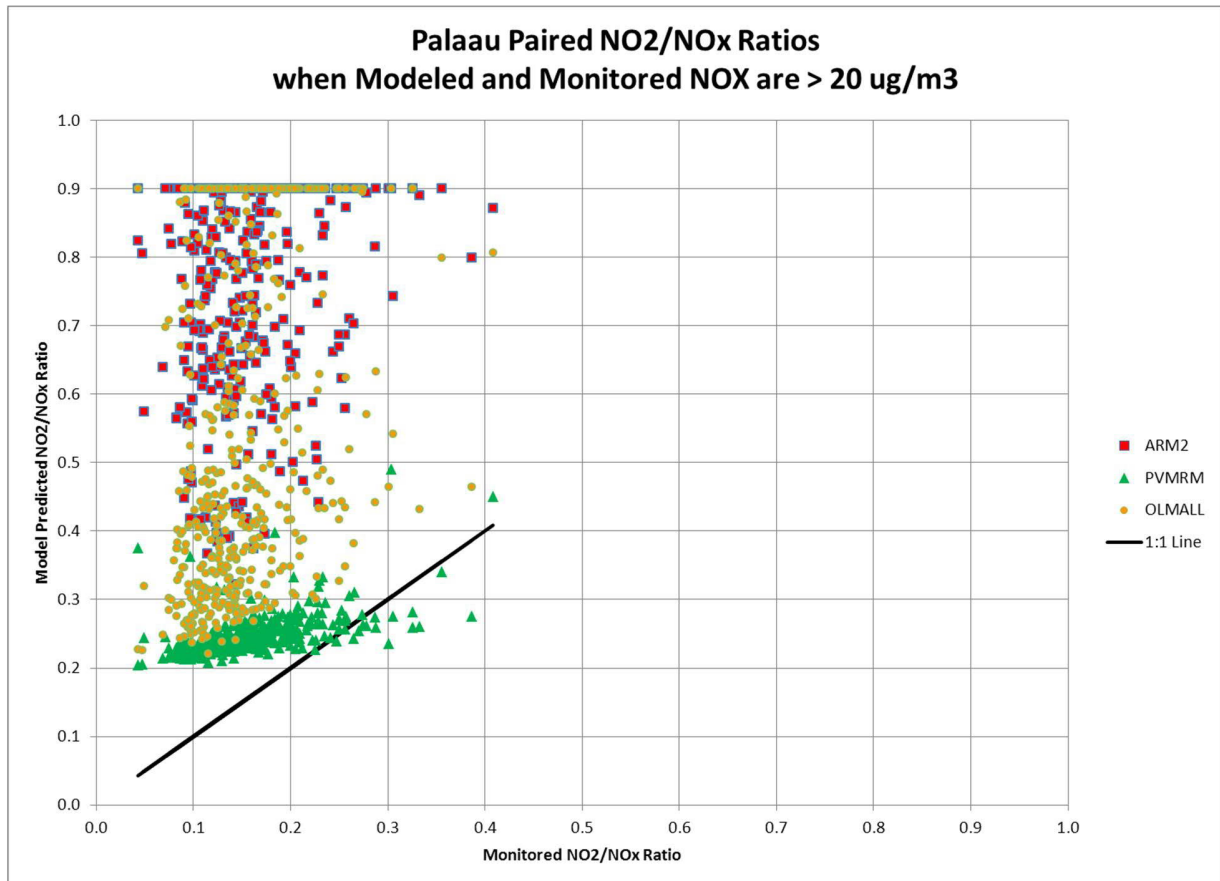
Figure 12 – Scatter Plot of NO₂/NO_x Ratio for Palaau Data Set

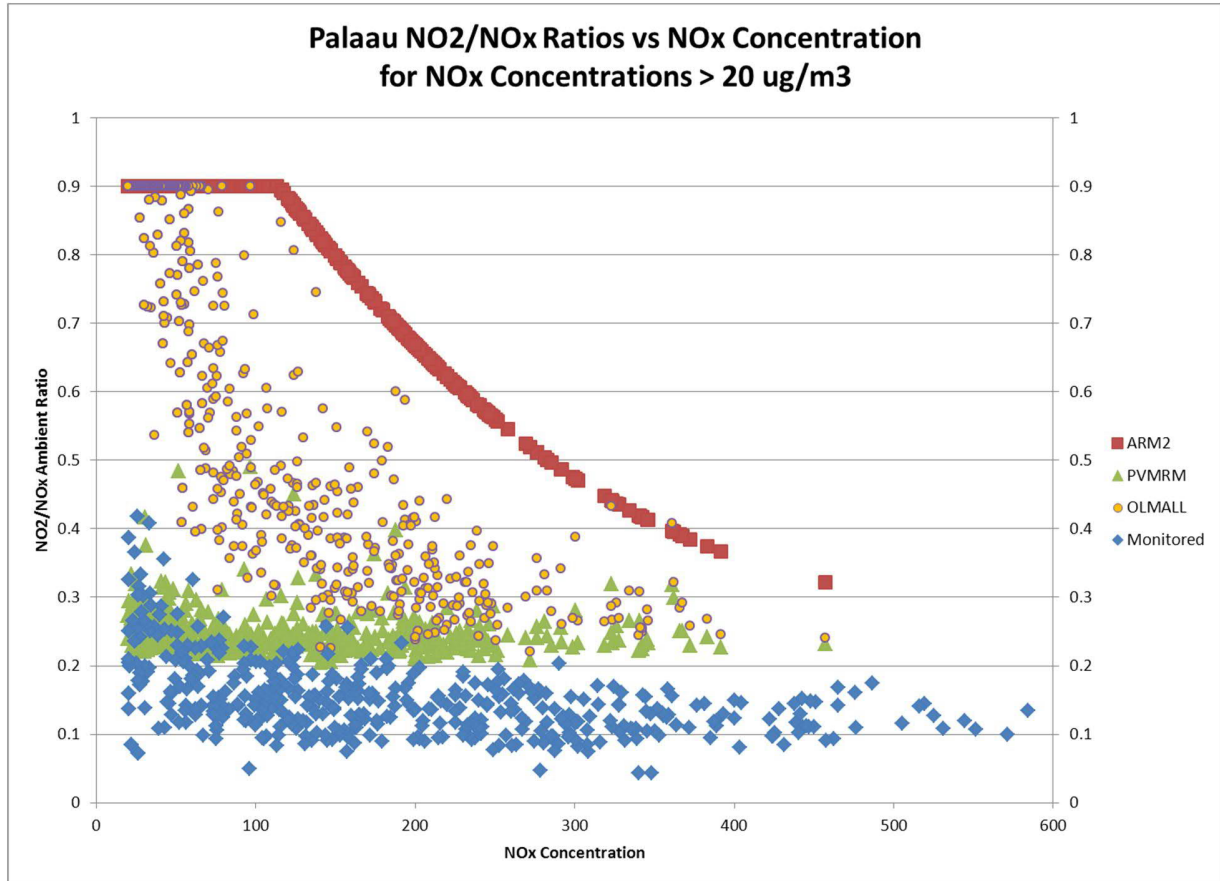
Figure 13 –Plot of NO₂/NO_x Ratio vs NO_x for Palaau Data Set

Figure 14 is the hourly Q-Q plot for NO_2 . These data were for those hours that met the “direct transport” wind direction criteria and had monitored or modeled concentrations greater than $1 \mu\text{g}/\text{m}^3$. All the methods exhibit a similar over-prediction at the highest NO_2 concentrations (an approximate factor of 2). As was the case with the Empire Abo data sets, the AERMOD-ARM2 plot shows significant over-prediction at lower concentrations, but the curve then flattens out and predicts high concentrations at similar levels as AERMOD-PVMRM.

4.7 Wainwright Evaluation Results

There were 458 hours that met the “direct transport” wind direction criteria (and had non-zero concentrations for both modeled and monitored NO_x concentrations). There were 185 of those hours that also had monitored NO_x concentrations greater than the $20 \mu\text{g}/\text{m}^3$ threshold, and 97 hours with both monitored and modeled NO_x concentrations greater than the $20 \mu\text{g}/\text{m}^3$ threshold. For the 97 hour data set, the maximum monitored NO_x concentration was $298 \mu\text{g}/\text{m}^3$, and the maximum monitored NO_2 concentration was $66 \mu\text{g}/\text{m}^3$.

The paired scatter plot of predicted versus observed NO_2/NO_x ratios is presented in Figure 15. This plot indicates that the observed ratios are clustered between 0.2 and 0.4. The PVMRM predicted ratios are clustered between 0.3 and 0.5, the OLM predicted ratios are found between 0.2 and 0.9, and the ARM2 predicted ratios are found between approximately 0.4 and 0.9. The OLM and ARM2 data points are predominately distributed above the 1:1 line. A large amount of scatter is observed for the OLM and ARM2 methods, but the scatter for PVMRM is appreciably less. These plots suggest that, for this data set, PVMRM has better skill than OLM or ARM2 for predicting the ratios on a paired in space and time basis, and that OLM and ARM2 have a greater tendency towards overprediction of the ambient ratio than PVMRM.

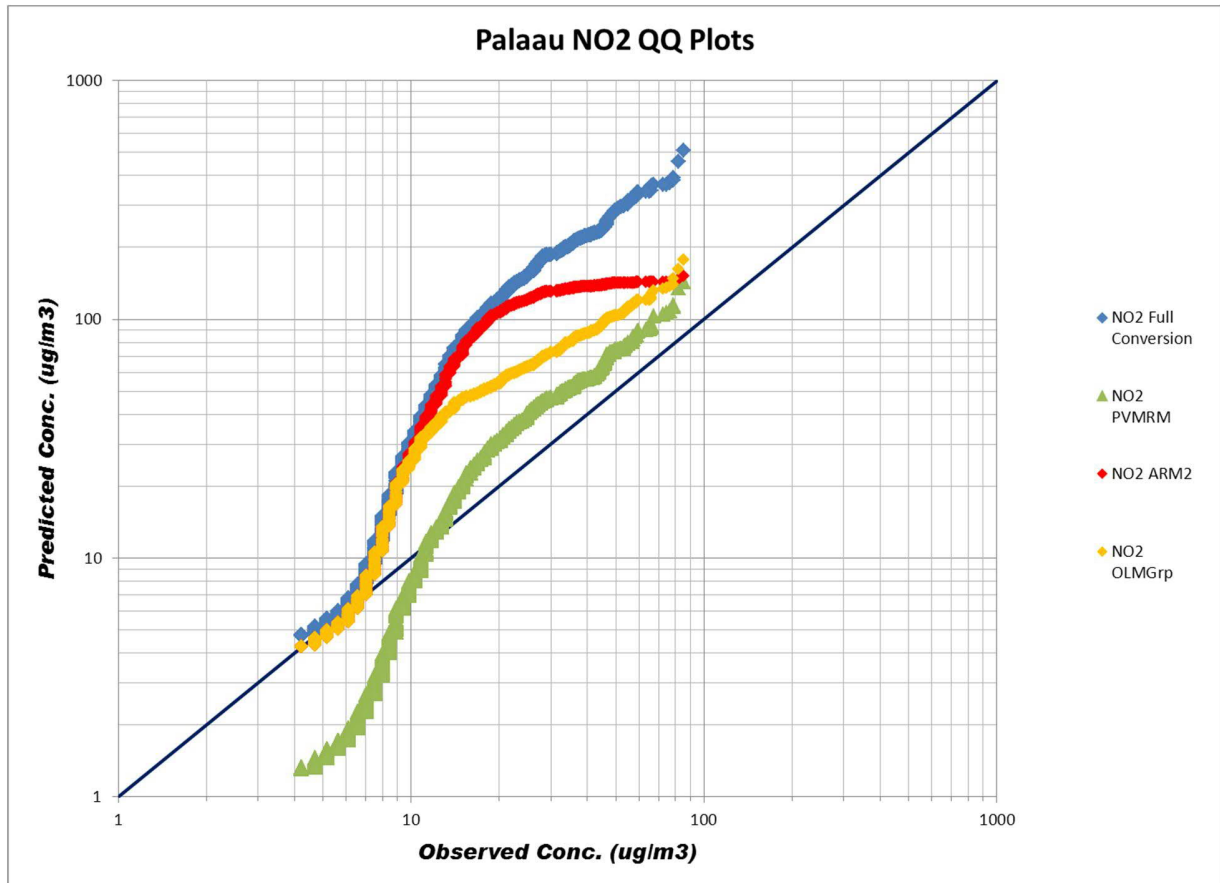
Figure 14 – Q-Q NO₂ Plot for the Palaau Data Set

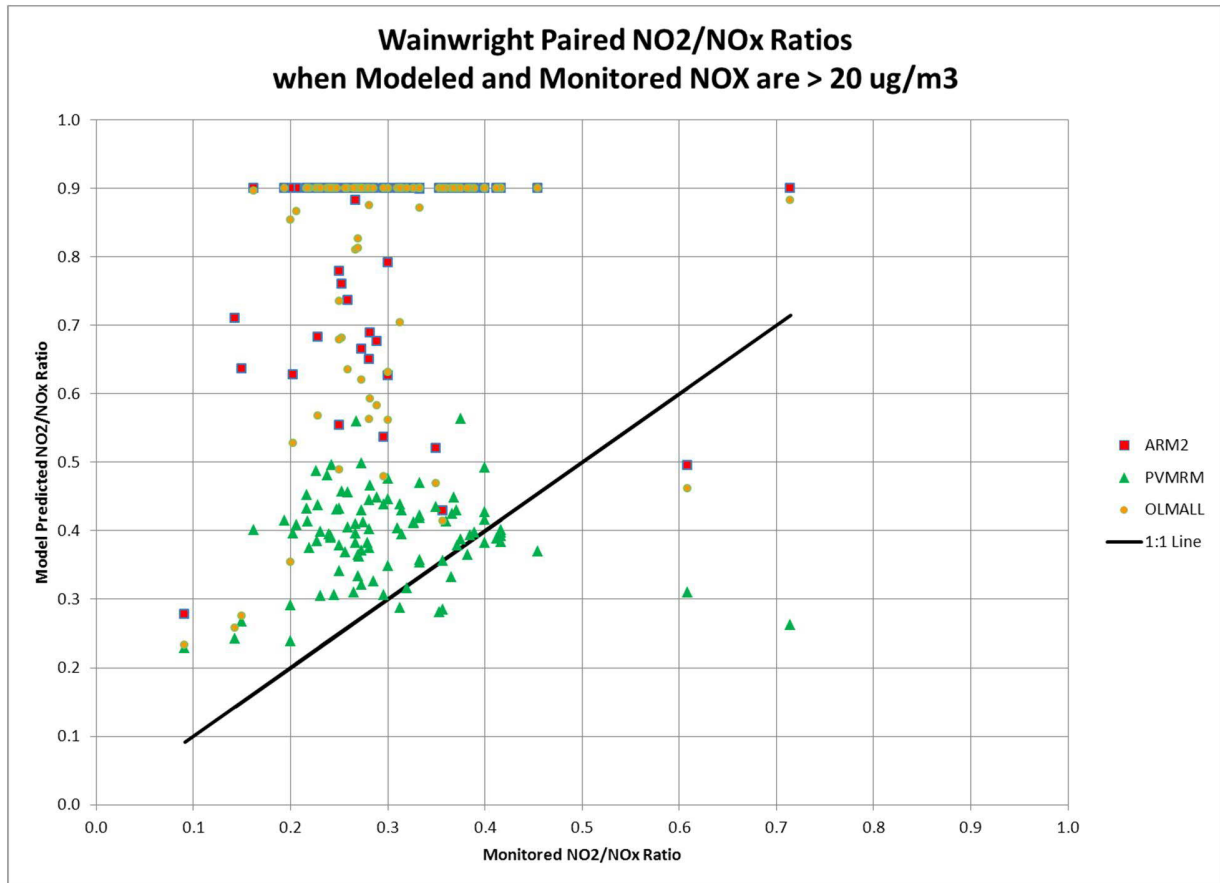
Figure 15 – Scatter Plot of NO₂/NO_x Ratio for Wainwright Data Set

Figure 16 is the scatter plot of the predicted NO_2/NO_x ratio as a function of AERMOD predicted NO_x concentration. The observed NO_2/NO_x ratios plotted as a function of observed NO_x concentrations are also included (the monitoring data are not paired in time with the modeling predictions). This plot indicates that the PVMRM predicted ratios generally match the observed ratios the best, while the OLM and ARM2 predicted ratios are typically higher than the observed ratios. However, at the extreme high end of the NO_x concentration range, the predicted ratios from the three methods cluster closer together.

Figure 17 presents the hourly Q-Q plot for NO_2 for the Wainwright data set. Similar to the previous plots, when combined with the AERMOD predictions of NO_x , the Tier 1 option of full conversion has a significant bias to overpredict NO_2 . The AERMOD-ARM2 and AERMOD-OLM methods show significant over-prediction at lower concentrations, but then predict the highest concentrations at very similar levels as AERMOD-PVMRM. All of the methods exhibit a similar over-prediction at the highest NO_2 concentrations (an approximate factor of 2).

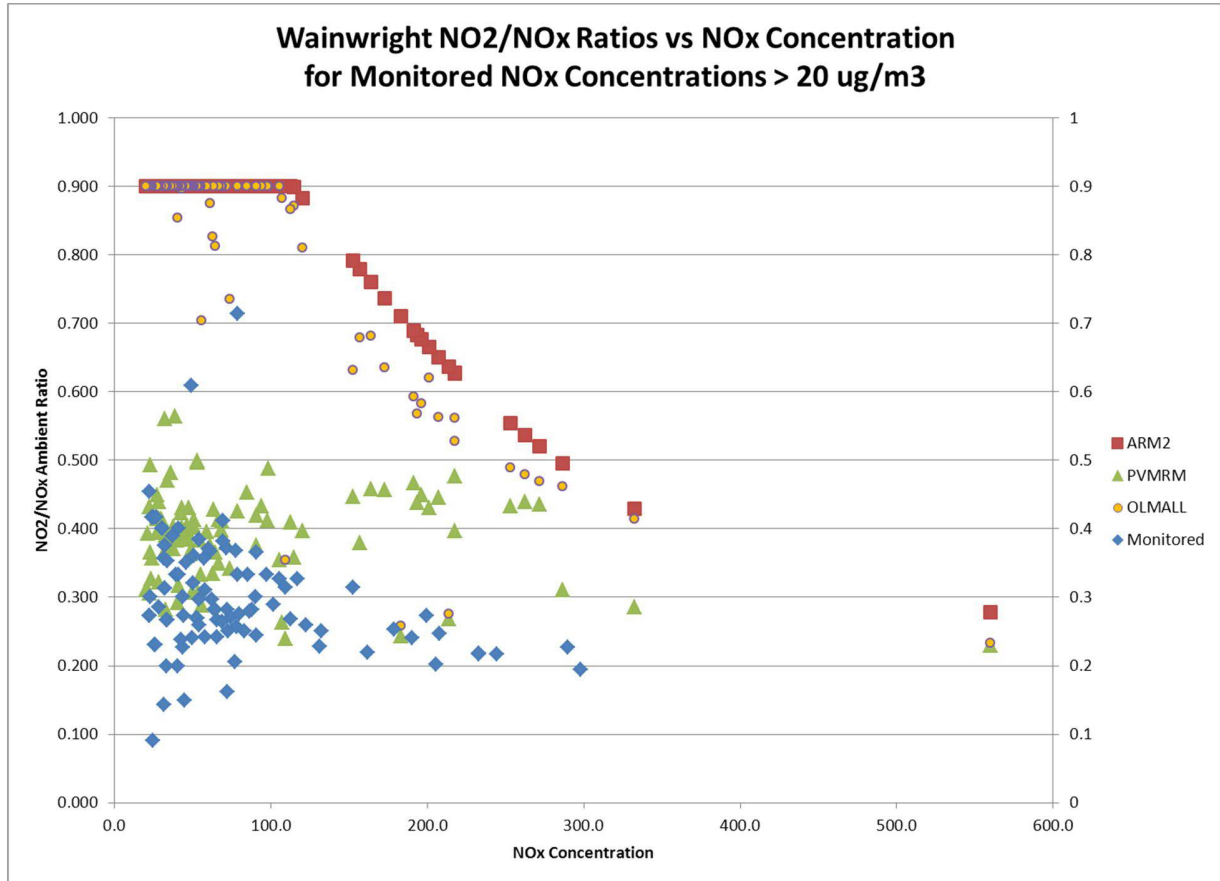
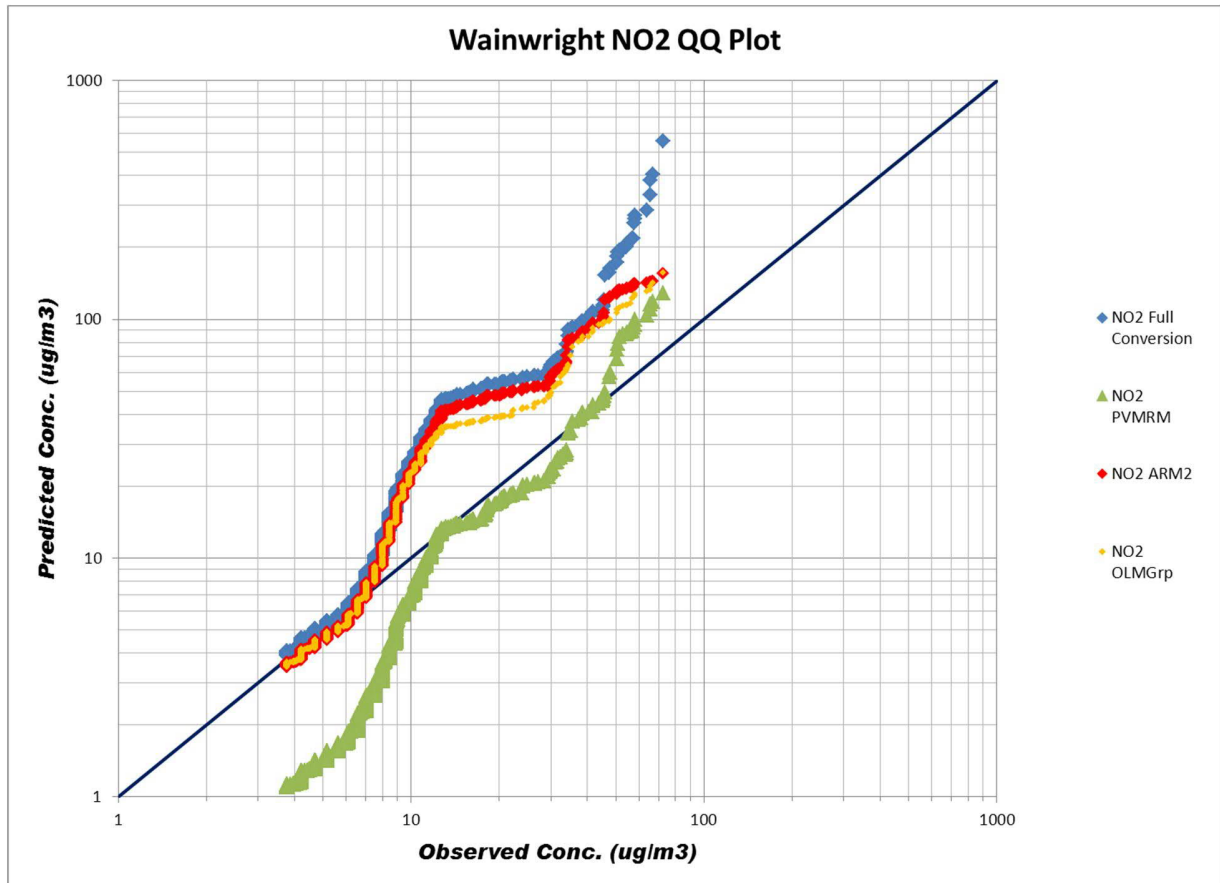
Figure 16 –Plot of NO₂/NO_x Ratio vs NO_x for Wainwright Data Set

Figure 17 – Q-Q Plot for NO₂ for Wainwright Data Set

4.8 RHC and Performance Evaluation Summary

The RHC represents a smoothed estimate of the highest concentrations, based on a tail exponential fit to the upper end of the concentration distribution. The RHC summary is presented in Table 5 for all of the evaluation data sets, along with the geometric means of the predicted RHC concentration divided by the observed RHC concentration across all data sets. These results are consistent with the relative performance illustrated in the Q-Q plots for NO₂. AERMOD predictions of NO_x combined with the full conversion option significantly over-predict NO₂ concentrations by an approximate factor of 5. Based on the RHC geometric means, the NO₂ performance of the AERMOD-ARM2, AERMOD-OLM, and AERMOD-PVMRM methods are all very similar at the elevated concentrations that are represented by the RHC, with the 3 methods over-predicting on average by factors of 1.8, 1.7, and 2.0 respectively. AERMOD-ARM2 performs better than AERMOD-PVMRM but not as well as AERMOD-OLM over all of these data sets.

Table 5 - Summary of RHCs for Evaluation Data Sets (µg/m³)

| | Observed NO _x | Observed NO ₂ | AERMOD NO _x (FULL NO ₂) | AERMOD- PVMRM NO ₂ | AERMOD- OLM NO ₂ | AERMOD- ARM2 NO ₂ |
|--|-----------------------------|-----------------------------|--|-------------------------------------|-----------------------------------|------------------------------------|
| New Mexico Abo North | 532.5 | 128.9 | 436.8 | 184.2 | 149.3 | 163.8 |
| New Mexico Abo South | 462.3 | 79.9 | 446.2 | 264.7 | 163.9 | 156.2 |
| Hawaii Palaau | 650.9 | 93.5 | 465.1 | 138.7 | 166.3 | 145.7 |
| Wainwright | 393.6 | 82.0 | 528.1 | 177.0 | 169.4 | 215.2 |
| Geometric Mean for Pred/Obs NO ₂ RHC | N/A | N/A | 4.96 | 1.97 | 1.72 | 1.78 |
| Geometric Mean for Pred/Obs NO _x RHC | N/A | N/A | 0.93 | N/A | N/A | N/A |

The geometric mean ratios for AERMOD-PVMRM and AERMOD-OLM are higher than what was previously reported in the 2011 Additional Clarifications memo. This is likely caused by the updates to the in-stack ratios (increased from 0.1 to 0.2), and for the Empire Abo data sets the use of “un-scavenged” ozone data. Since these changes are intended to better reflect current 1-hr NO₂ modeling guidance, these updated performance results better reflect the accuracy of the Tier 3 screening methods using current guidance.

To help determine the performance of AERMOD itself and identify potential compensating errors between the dispersion model and the NO₂ conversion methodology, RHC statistics for NO_x concentrations are also presented in Table 5. The geometric mean of the AERMOD predicted RHC NO_x concentration (equivalent to the full conversion option) divided by the Observed RHC NO_x concentration is 0.939. This indicates that on average AERMOD is under predicting the total NO_x concentration by approximate 6%, while all of the “AERMOD plus Tier 2 or 3” combinations are over-predicting the NO₂ concentration by 70% to 100%. This suggests that for these data sets, the NO₂ conversion methodologies are the most important factor in the observed over-prediction of NO₂ concentrations.

5.0 ARM2 Sensitivity Analyses

The ARM2 sensitivity analyses compare the predicted NO₂ concentrations using AERMOD and the ARM2, PVMRM, OLM, and full conversion methods across a wide range of meteorology and emissions source characteristics. The scenarios modeled include updated versions of those in the MACTEC report “Sensitivity Analysis of PVMRM and OLM in AERMOD”, 2004 for the diesel generator, gas turbine, and “35-meter stack” single source scenarios, and the cumulative source scenario. In addition, some scenarios were analyzed that are representative of large diesel generators, a refinery, a gas pipeline compressor station, natural gas production fields and processing plants, and a large boiler.

5.1 Updates to 2004 Sensitivity Scenarios

The 2004 MACTEC sensitivity analysis modeled a variety of single source scenarios, including a generic 35m buoyant point source, a diesel generator source, and a gas turbine source. A range of buoyancy and momentum fluxes are represented by these sources, with the diesel generator representing a relatively low buoyancy/low momentum release, the 35m stack representing medium buoyancy and momentum, and the gas turbine representing relatively high buoyancy and momentum. The emission rates for the diesel generator and gas turbine sources are considered typical uncontrolled rates for those source types. The 35m stack was modeled with two different emission rates to test the sensitivity of the ARM2, PVMRM, and OLM algorithms to the NO_x emission rate. In addition to these single source scenarios, a hypothetical multiple-source scenario was included in the 2004 MACTEC sensitivity analysis to test the plume merging algorithm of PVMRM. This multiple-source scenario included a total of 65 point sources with 1,598 receptors (note that the area sources were removed from the original input files for this updated analysis). For additional information on these scenarios, refer to http://www.epa.gov/ttn/scram/7thconf/aermod/pvmrm_sens.pdf.

EPA provided the AERMOD input files, meteorological data, and ozone data from these previous sensitivity analyses. The diesel generator, gas turbine, and “35-meter source” scenarios have been

rerun using the current version of AERMOD (12345). Note that the meteorological data files were simply edited to reflect an AERMET version of 11059 so that AERMOD version 12345 could be run with these meteorological files. The in-stack NO_2/NO_x ratio of 0.1 in the original analysis was updated to a value of 0.2 to better reflect current guidance. While the MACTEC study evaluated the maximum 1-hr and annual concentrations, this updated sensitivity analysis focused on the 1-hr NO_2 design concentrations directly reported by AERMOD (the design concentration is defined as the multiyear average of the 98th percentile of the annual distribution of daily maximum 1-hour values).

Table 6 and Figure 18 present the results for the single source scenarios. In general, the modeled concentrations and distance to maximum impacts are similar to the 2004 results. A comparison of the 35m stack with 1g/s versus 50g/s emission rate shows the sensitivity of the predicted NO_2/NO_x ratio to the level of NO_x emissions. For the 1g/s source, and for other scenarios with lower predicted NO_x concentrations, the ratio of predicted NO_2 concentration versus full conversion NO_x concentration was close to 1.0 for all methods (for the PVMRM and ARM2 methods the maximum ratio of 0.90 is reached). For other scenarios with larger predicted NO_x concentrations (i.e., when the full conversion NO_x concentrations are greater than $100 \mu\text{g}/\text{m}^3$) such as the downwash and complex terrain cases, ARM2 predicts the highest ratios, followed by OLM and then PVMRM. This is true for scenarios with both close-in impact locations with little time for entrainment of ozone (for example, the diesel generator with downwash in rural location), as well as more distant impact locations that allow more time for ozone entrainment and reaction (for example, the gas turbine in complex terrain). The average ratios of predicted NO_2 concentration versus full conversion concentration for all scenarios with predicted NO_x concentrations equal to or greater than $100 \mu\text{g}/\text{m}^3$ are 0.67 for ARM2, 0.54 for OLM, and 0.35 for PVMRM.

Table 6 - Comparison of 1-hr NO₂ Design Concentrations for Single Source Scenarios

| Source Scenario | Conversion Option | NO ₂ Design Conc. (µg/m ³) | Distance to Maximum (m) | Ratio of NO ₂ vs. Full Conversion |
|-------------------------------------|-------------------|---|-------------------------|--|
| 35m Stack, 1g/s Rural, No Downwash | FULL | 2.7 | 500 | 1.00 |
| | OLM | 2.7 | 500 | 1.00 |
| | PVMM | 2.4 | 500 | 0.90 |
| | ARM2 | 2.4 | 500 | 0.90 |
| 35m Stack, 50g/s Rural, No Downwash | FULL | 133.0 | 500 | 1.00 |
| | OLM | 73.2 | 750 | 0.55 |
| | PVMM | 33.2 | 3000 | 0.25 |
| | ARM2 | 112.4 | 500 | 0.84 |
| Diesel Generator Rural, No Downwash | FULL | 56.5 | 300 | 1.00 |
| | OLM | 47.5 | 300 | 0.84 |
| | PVMM | 21.8 | 1500 | 0.38 |
| | ARM2 | 50.9 | 300 | 0.90 |
| Diesel Generator Rural, Downwash | FULL | 386.5 | 50 | 1.00 |
| | OLM | 126.5 | 50 | 0.33 |
| | PVMM | 88.0 | 50 | 0.23 |
| | ARM2 | 143.5 | 50 | 0.37 |
| Diesel Generator Urban, No Downwash | FULL | 76.1 | 200 | 1.00 |
| | OLM | 55.0 | 200 | 0.72 |
| | PVMM | 23.0 | 3000 | 0.30 |
| | ARM2 | 68.5 | 200 | 0.90 |
| Diesel Generator Flat Terrain | FULL | 64.0 | 200 | 1.00 |
| | OLM | 61.1 | 200 | 0.96 |
| | PVMM | 26.7 | 1000 | 0.42 |
| | ARM2 | 57.6 | 200 | 0.90 |
| Diesel Generator Complex Terrain | FULL | 99.6 | 1500 | 1.00 |
| | OLM | 85.1 | 1500 | 0.85 |
| | PVMM | 52.8 | 3000 | 0.53 |
| | ARM2 | 89.7 | 1500 | 0.90 |
| Gas Turbine Rural, No Downwash | FULL | 15.8 | 750 | 1.00 |
| | OLM | 15.8 | 750 | 1.00 |
| | PVMM | 10.8 | 3000 | 0.68 |
| | ARM2 | 14.3 | 750 | 0.90 |
| Gas Turbine Rural, Downwash | FULL | 348.1 | 100 | 1.00 |
| | OLM | 98.0 | 100 | 0.28 |
| | PVMM | 68.9 | 200 | 0.20 |
| | ARM2 | 142.8 | 100 | 0.41 |
| Gas Turbine Urban, No Downwash | FULL | 54.9 | 1500 | 1.00 |
| | OLM | 50.5 | 1500 | 0.92 |
| | PVMM | 37.1 | 1500 | 0.68 |
| | ARM2 | 49.5 | 750 | 0.90 |
| Gas Turbine Flat Terrain | FULL | 20.4 | 750 | 1.00 |
| | OLM | 20.4 | 750 | 1.00 |
| | PVMM | 14.8 | 750 | 0.72 |
| | ARM2 | 18.4 | 750 | 0.90 |
| Gas Turbine Complex Terrain | FULL | 146.7 | 2000 | 1.00 |
| | OLM | 97.3 | 2000 | 0.66 |
| | PVMM | 76.9 | 2000 | 0.52 |
| | ARM2 | 118.3 | 2000 | 0.81 |

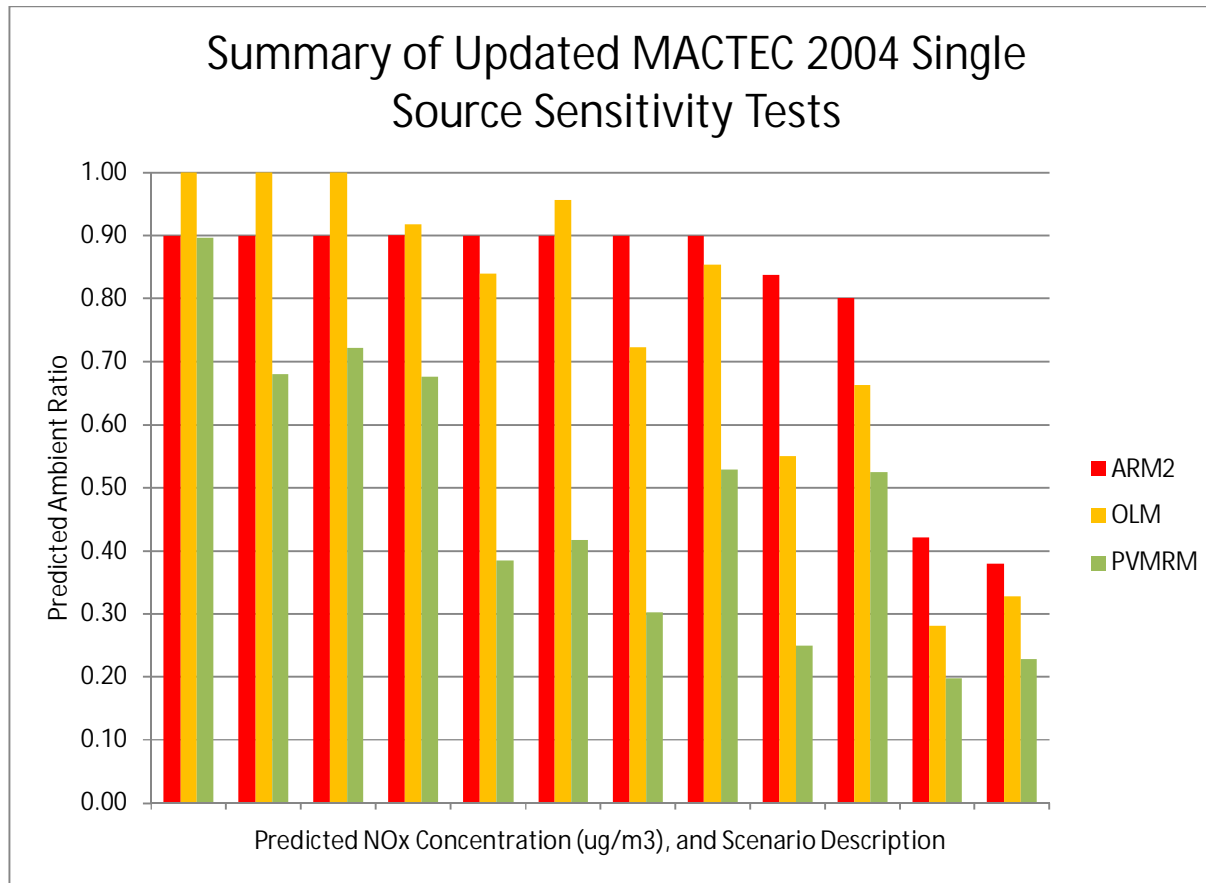
Figure 18 – Results from Single Source Sensitivity Analyses

Table 7 presents the results of the multiple-source scenario, showing the 1-hour NO₂ design concentration and the ratio of predicted NO₂ concentration versus full conversion concentration for each of the conversion options. The modeled NO_x concentrations for this scenario are high. The results indicate that the predicted NO₂/NO_x ratios for all three methods are similar, with the OLM method being the lowest, and PVMRM and ARM2 being very similar. The distance to maximum modeled impact for this multi-source scenario is 500 meters for ARM2 and OLM, and 1950 meters for PVMRM.

Table 7 - Comparison of 1-hr NO₂ Design Concentrations for Multi-Source Scenario

| Conversion Option | Maximum NO ₂ Concentration (µg/m ³) | Distance to Maximum (m) | Predicted NO ₂ /NO _x Ratio |
|-------------------|---|----------------------------|---|
| Full Conversion | 1774 | 502 | 1 |
| OLM-GroupAll | 238 | 502 | 0.13 |
| PVMRM | 322 | 1953 | 0.18 |
| ARM2 | 355 | 502 | 0.20 |

5.2 Analysis of Additional Scenarios

Some additional multi-source scenarios were analyzed that are based on actual configurations of large diesel IC generators, a refinery, a gas pipeline compressor station, natural gas production fields and processing plants, and a large boiler in complex terrain. The first five of these additional cases were originally presented at the EPA 10th Modeling Conference in the paper “Updated Tier 2 Ambient Ratio Method (ARM) for 1-hr NO₂ NAAQS Analyses”, March 12, 2012. They have been updated with the AERMOD version 12345, and the version of AERMOD that includes the ARM2 code changes. For each of these cases except the large boiler, the assumed in-stack ratio was 0.2. For the large boiler, the assumed in-stack ratio was 0.1 (a more representative value for large boilers).

A brief description of each of these additional scenarios follows:

- Diesel IC Generators – This scenario includes five IC engine generators rated at approximately 2500 kW, with NO_x emission rates of 2.6 g/hp-hr. The stack heights are approximately 100 feet and there is a large 200 foot tall building nearby, so downwash is a factor for this scenario.
- Refinery – This scenario includes 59 emission units with stack heights ranging from 80 to 350 feet tall, representing typical refinery distillation, cracking, treating, and reforming emission sources including a fluid catalytic cracking (FCC) unit. Building structures are not included in this scenario.
- Gas Pipeline Compressor station – This scenario includes a 10 Mw combustion turbine with a 35 foot tall stack and a 30 foot tall nearby building, and a 500 hp IC engine generator with a 15 foot tall stack and a 15 foot tall nearby building.
- Gas Plant and Production Scenario A – This scenario includes 84 sources that represent typical natural gas plant and gas field production sources, including IC engines (for pumps and compressors), heaters, and dehydration units. The stack heights range from 8 to 100 feet, and no buildings or structures were included.
- Gas Field Production Scenario B – This scenario examines the close-in impacts from two IC engines, one with an 8 foot stack and the second with a 25 foot stack. There is no building downwash, but the receptor is located approximately 20 meters distant from the emission units.
- Large Boiler with Complex Terrain – This scenario examines the impacts from a large 3,000 MMBtu/hr boiler with NO_x emissions equal to 0.2 lb/MMBtu, and a 300 foot tall stack. A 150 foot tall adjacent building was included in the analysis. This emission source was located at three distances (1, 3, and 10 km) from an elevated ridge approximately 300 to 500 feet tall.

Table 8 presents the results from these additional scenario analyses, including the ratio of the predicted 1-hr NO₂ design concentration versus the predicted full conversion 1-hr NO_x design concentration for each method. Four of the scenarios have predicted NO_x concentrations that are moderately high (ranging from 190 to 461 µg/m³), and the other four scenarios have NO_x concentrations that are very high (ranging from 1080 to 2823 µg/m³).

For the first two scenarios (generators and refinery examples), the OLM and ARM2 predicted concentrations are very similar, while the PVMRM predicted concentrations are slightly lower. For the third scenario (compressor station), the PVMRM predicted concentration is significantly higher than either OLM or ARM2. A supplemental model run was made for this scenario using OLM on a “source-by-source” conversion basis; this should establish an upper limit for ozone limited conversion. The OLM “source-by-source” concentration was still substantially lower than the PVMRM predicted concentration. Therefore, the PVMRM result for the compressor station scenario appears unrealistically high. This scenario included downwash, and the maximum impact receptor for all of the methods was located at the nearby fence line, within the wake zone of the building. Under these conditions, PVMRM may be having difficulty calculating the plume volume that ambient ozone is entrained into.

Table 8 - 1-hr NO₂ Design Concentrations for Additional Sensitivity Scenarios

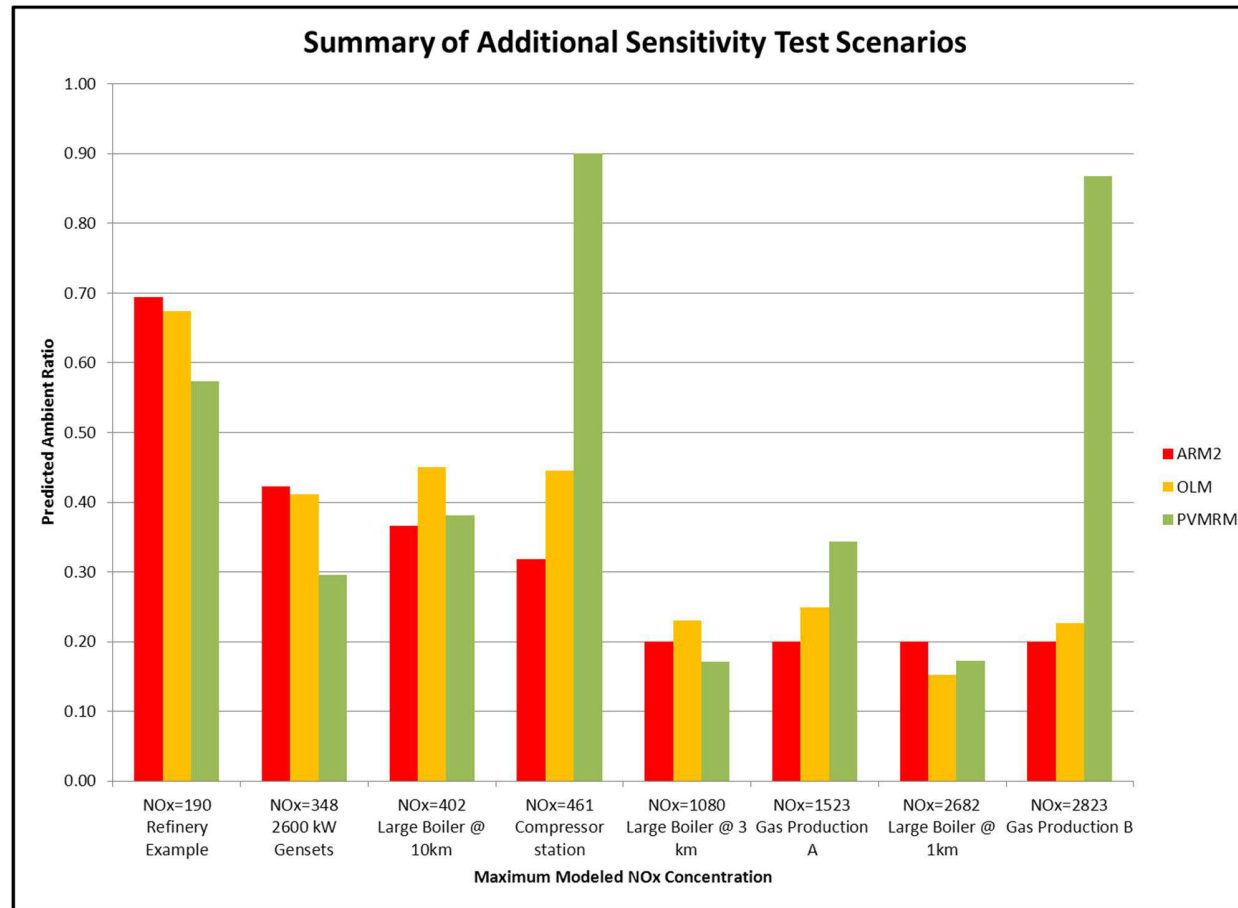
| Description of Scenario | Full NO _x µg/m ³ | OLM µg/m ³ | PVMRM µg/m ³ | ARM2 µg/m ³ | OLM Ratio | PVMRM Ratio | ARM2 Ratio |
|-------------------------|---|--------------------------|----------------------------|---------------------------|--------------|----------------|---------------|
| 2600 kW Gensets | 348 | 143 | 103 | 145 | 0.41 | 0.30 | 0.42 |
| Refinery Example | 190 | 128 | 109 | 129 | 0.67 | 0.57 | 0.68 |
| Compressor station | 461 | 205 | 415 | 147 | 0.44 | 0.90 | 0.32 |
| Gas Production A | 1523 | 380 | 524 | 305 | 0.25 | 0.34 | 0.20 |
| Gas Production B | 2823 | 640 | 2450 | 565 | 0.23 | 0.87 | 0.20 |
| Large Boiler @ 1km | 2682 | 352 | 352 | 422 | 0.13 | 0.13 | 0.16 |
| Large Boiler @ 3 km | 957 | 237 | 174 | 191 | 0.25 | 0.18 | 0.20 |
| Large Boiler @ 10km | 334 | 174 | 129 | 143 | 0.52 | 0.39 | 0.43 |

For the gas production field scenarios, the OLM and ARM2 concentrations are similar, while the PVMRM concentrations and ratios are substantially higher. In particular, the PVMRM ratio for the gas production B scenario is very high, similar to the situation discussed above for the compressor station scenario. Like that scenario, the receptor with the maximum predicted impact for the gas production B scenario is located close to the source (approximately 20 meters distant), and the stack heights are low (8 to 25 feet). OLM on a source-by-source method was run for the gas production B scenario to determine an upper limit for the conversion, and the result was the same as the OLM Group All concentration (the gas production B scenario does consist of only two sources, so this is not surprising). Therefore, once again the PVMRM predicted concentration and ratio appear to be unrealistically high for the gas production B scenario.

The large boiler examples evaluate the relative performance of the various methods at three distances. As the distance increases, the NO_x concentrations decrease with the additional time for dispersion, and the NO_2/NO_x ratios should increase with additional time for entrainment and reaction. The predicted ambient ratios do increase with each distance value for all three methods. For all of the boiler cases, the ARM2 predicted ratios are higher or equal to the lower of the two Tier 3 methods.

Figure 19 presents the ratio of the predicted 1-hr NO_2 design concentration versus the predicted full conversion 1-hr NO_x design concentration for these additional scenarios, sorted by the predicted NO_x concentration. This plot shows the general tendency of all the methods to predict lower ambient ratios as the modeled NO_x concentration increases. The two cases with unrealistic PVMRM predicted concentrations are readily evident. For all other cases, the ARM2 method predicts higher or similar concentrations and ratios than the lower of the two Tier 3 methods.

Figure 19 – Results from Additional Scenario Sensitivity Analyses



6.0 Conclusions and Recommendations

6.1 Summary of Key Findings

The key findings from the ambient monitoring data analyses, performance evaluations, and sensitivity analyses are:

1. Plots of ambient NO_2/NO_x ratios as a function of NO_x concentration from various ambient monitoring data sets all show the same relationship of decreasing ambient ratios with increasing NO_x concentrations. At NO_x concentrations above approximately 300 ppb, the observed ambient ratios tend to cluster in a range of approximately 0.1 to 0.2.
2. The ARM2 conversion method was developed using 10 years of ambient monitoring data from 580 monitoring sites in EPA's AQS data base. ARM2 uses an empirically derived relationship between an upper limit of the observed NO_2/NO_x ambient ratio versus the ambient NO_x concentration.
3. ARM2 predicted ambient ratios derived from various geographical, land use, and time period data sets were evaluated and found to be similar to, and typically lower than, the ratios predicted by the "All AQS Sites" equation. This indicates that the ARM2 equation derived from the "All AQS Sites" data set is representative of a wide range of geographical or land use categories, and can be used as the basis of the ARM2 method.
4. The ARM2 conversion method has been programmed into the latest version of AERMOD (version 12345).
5. The performance of the ARM2 method has been compared to monitoring observations and predictions from the Tier 3 screening methods, using the same evaluation data sets that have previously been used to test the Tier 3 methods. Plots of ambient ratios versus NO_x concentration indicate that all three methods overpredict the observed ambient ratios. Scatter plots of predicted versus observed NO_2/NO_x ratios indicate that all methods have little skill in predicting the ambient ratio on a "paired in space and time" basis. The Q-Q

NO₂ plots indicate that all methods overpredict the highest NO₂ concentrations by factors ranging from approximately 1.2 to 2.

6. The Robust Highest Concentration (RHC) summary of the evaluation results is a measure of model performance based on the top 26 highest modeled and observed NO₂ concentrations (which range from 55 to 196 µg/m³). The RHC summary indicates that the AERMOD-ARM2, AERMOD-OLM, and AERMOD-PVMRM methods perform very similar, with all 3 methods over-predicting the highest NO₂ concentrations by factors of 1.8, 1.7, and 2.0 respectively. The ARM2 RHC result falls between the OLM and PVMRM results.
7. Sensitivity analyses were performed for the ARM2, PVMRM, OLM, and full conversion methods across a range of meteorology and source characteristics. The scenarios modeled included updated versions of those described in the MACTEC report “Sensitivity Analysis of PVMRM and OLM in AERMOD”, 2004 for the diesel generator, gas turbine, and 35-meter stack “single source” scenarios, and the cumulative source scenario. For the single source sensitivity scenarios with low predicted NO_x concentrations (below 20 µg/m³), all of the conversion methods predicted NO₂/NO_x ratios near 0.9. When the predicted NO_x concentrations were higher (greater than 300 µg/m³), all the methods predict NO₂/NO_x ratios in the range of 0.2 to 0.4, with ARM2 conservatively predicting the highest NO₂/NO_x ratios of any of the methods. For the multi-source scenario, the model predicted NO_x concentration is 1,774 µg/m³, and the predicted NO₂/NO_x ratios are 0.13 for OLM, 0.18 for PVMRM, and 0.2 for ARM2.
8. Some additional multi-source scenarios were also analyzed. These scenarios were based on real-world configurations of large diesel IC generators, a refinery, a gas pipeline compressor station, natural gas production fields and processing plants, and a large boiler in complex terrain. The ARM2 method predicts higher or similar ratios when compared to the Tier 3 methods. Two of the cases were further evaluated because of suspiciously high PVMRM predicted ratios, and based on comparisons of PVMRM versus OLM “source-by-source” predicted ratios, it was concluded that PVMRM is predicting unrealistically high concentrations and ratios for those cases. In summary, the additional

multi-source sensitivity analysis indicates that the ARM2 method predicts higher or similar ratios when compared to the two Tier 3 methods.

6.2 Recommendation

The performance evaluations and sensitivity analyses presented in this report document that the performance of the ARM2, PVMRM, and OLM methods is very similar. The relative performance ranking can vary depending upon the data set, and there can be cases when PVMRM is predicting unrealistically high NO₂/NO_x ratios.

The ARM2 method has some advantages over the Tier 3 methods. ARM2 does not require case-by-case review and approval of special input data such as in-stack NO₂/NO_x ratios and ambient ozone data. There is no concern over possible ozone scavenging in data used to estimate NO₂ conversion. The model execution times are faster than the Tier 3 methods. ARM2 can reduce the resources expended by both the facility and the reviewing agency to perform and review 1-hr NO₂ NAAQS modeling analyses.

Based on the analyses presented herein, it is recommended that ARM2 be approved as a refinement to the current fixed-ratio ARM method for performing 1-hr NO₂ analyses. Given its model performance as detailed in this report, ARM2 could also be considered a Tier 3 screening method equivalent to the other two Tier 3 screening methods. Consistent with other recent updates to AERMOD, EPA could implement the ARM2 refinement by posting a memorandum on the SCRAM website authorizing its use for AERMOD modeling analyses, and providing an updated version of AERMOD that includes the ARM2 model option.

APPENDIX A

List of AQS NO_x monitoring sites and available data periods by site

List of AQS NOx Monitoring Stations and Number of Valid Observations by Year

| State | County | | | | | | | | | | | | |
|-------|--------|---------|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Code | Code | Site ID | Total Obs | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 1 | 33 | 1002 | 8,285 | | 6,298 | 1,987 | | | | | | | |
| 1 | 71 | 20 | 8,162 | | | | 7,451 | 711 | | | | | |
| 1 | 83 | 4 | 8,261 | | | 6,174 | 2,087 | | | | | | |
| 4 | 7 | 10 | 18,046 | | 2,529 | 3,668 | 3,933 | 4,107 | 3,809 | | | | |
| 4 | 12 | 8000 | 8,167 | | | | | 3,091 | 5,076 | | | | |
| 4 | 13 | 9993 | 18,830 | 3,398 | 5,172 | 5,155 | 5,105 | | | | | | |
| 4 | 13 | 9997 | 82,126 | 6,817 | 8,597 | 8,471 | 8,653 | 8,698 | 8,701 | 8,554 | 8,217 | 8,307 | 7,111 |
| 4 | 19 | 1011 | 2,186 | | | | | | | | | | 2,186 |
| 4 | 19 | 1028 | 2,070 | | | | | | | | | | 2,070 |
| 4 | 21 | 8001 | 25,458 | 3,612 | 4,848 | 4,941 | 4,306 | 3,761 | 3,990 | | | | |
| 4 | 27 | 6 | 17,377 | | | | | 6,549 | 8,483 | 2,345 | | | |
| 5 | 35 | 5 | 25,759 | | | | | | | | 8,464 | 8,712 | 8,583 |
| 5 | 119 | 7 | 73,755 | | 7,852 | 8,131 | 7,815 | 7,921 | 8,337 | 8,307 | 8,234 | 8,484 | 8,674 |
| 5 | 119 | 1002 | 7,042 | 7,042 | | | | | | | | | |
| 5 | 119 | 1005 | 457 | | 457 | | | | | | | | |
| 6 | 1 | 7 | 33,027 | | | | | | | 8,309 | 8,352 | 8,213 | 8,153 |
| 6 | 1 | 9 | 25,825 | | | | | | | 1,378 | 8,096 | 8,158 | 8,193 |
| 6 | 1 | 10 | 12,056 | 1,214 | 8,091 | 2,751 | | | | | | | |
| 6 | 1 | 11 | 15,004 | | | | | | | | | 6,920 | 8,084 |
| 6 | 1 | 1001 | 31,615 | | | | | | | 8,265 | 8,356 | 8,138 | 6,856 |
| 6 | 1 | 2004 | 25,392 | | | | | | | 537 | 8,321 | 8,315 | 8,219 |
| 6 | 7 | 2 | 82,478 | 8,319 | 8,163 | 8,342 | 8,353 | 8,324 | 8,349 | 7,608 | 8,305 | 8,342 | 8,373 |
| 6 | 13 | 2 | 33,053 | | | | | | | 8,315 | 8,316 | 8,304 | 8,118 |
| 6 | 13 | 10 | 10,989 | 1,203 | 7,357 | 2,429 | | | | | | | |
| 6 | 13 | 1002 | 33,060 | | | | | | | 8,238 | 8,333 | 8,270 | 8,219 |
| 6 | 13 | 1004 | 23,069 | | | | | | | 8,253 | 7,929 | 1,788 | 5,099 |
| 6 | 13 | 3001 | 16,384 | | | | | | | 8,147 | 8,237 | | |
| 6 | 17 | 11 | 27,573 | 8,370 | 8,249 | 8,237 | 2,717 | | | | | | |
| 6 | 17 | 12 | 19,775 | 3,719 | 5,045 | 5,933 | 5,078 | | | | | | |
| 6 | 19 | 7 | 77,732 | 8,145 | 7,043 | 8,039 | 8,317 | 8,264 | 8,035 | 7,963 | 8,289 | 7,771 | 5,866 |
| 6 | 19 | 8 | 82,276 | 8,301 | 8,267 | 8,286 | 8,321 | 7,900 | 8,231 | 8,297 | 8,110 | 8,247 | 8,316 |
| 6 | 19 | 242 | 76,219 | 7,844 | 8,144 | 8,238 | 7,960 | 8,221 | 7,916 | 7,993 | 6,506 | 7,206 | 6,191 |
| 6 | 19 | 243 | 5,131 | | | 5,131 | | | | | | | |
| 6 | 19 | 244 | 4,929 | | | 4,929 | | | | | | | |
| 6 | 19 | 4001 | 78,208 | 8,000 | 8,263 | 8,066 | 8,291 | 8,294 | 7,913 | 8,257 | 8,005 | 6,976 | 6,143 |
| 6 | 19 | 5001 | 79,051 | 8,016 | 8,105 | 7,728 | 8,204 | 8,190 | 8,203 | 8,012 | 7,961 | 6,685 | 7,947 |
| 6 | 25 | 5 | 81,336 | 6,956 | 8,263 | 8,253 | 8,323 | 8,258 | 8,223 | 8,270 | 8,298 | 8,207 | 8,285 |
| 6 | 25 | 6 | 69,556 | 6,646 | 6,723 | 7,705 | 8,093 | 6,646 | 8,180 | 6,998 | 8,187 | 7,720 | 2,658 |
| 6 | 25 | 1003 | 57,303 | 1,603 | 4,247 | 6,721 | 6,659 | 8,005 | 7,191 | 7,395 | 5,873 | 6,843 | 2,766 |
| 6 | 29 | 7 | 80,709 | 7,758 | 8,368 | 8,177 | 7,981 | 8,293 | 8,282 | 8,188 | 7,946 | 7,985 | 7,731 |
| 6 | 29 | 10 | 69,727 | 7,732 | 7,944 | 7,150 | 5,938 | 8,295 | 8,294 | 8,159 | 8,146 | 7,970 | 99 |
| 6 | 29 | 11 | 33,645 | 8,233 | 7,152 | 7,742 | 7,435 | 3,083 | | | | | |
| 6 | 29 | 14 | 81,972 | 8,027 | 8,240 | 8,230 | 7,619 | 8,355 | 8,248 | 8,243 | 8,315 | 8,345 | 8,350 |
| 6 | 29 | 232 | 40,648 | 8,289 | 8,047 | 8,248 | 7,759 | 8,305 | | | | | |
| 6 | 29 | 5001 | 79,141 | 8,301 | 8,376 | 8,304 | 8,222 | 8,141 | 8,166 | 8,222 | 7,115 | 7,691 | 6,603 |
| 6 | 29 | 6001 | 82,308 | 8,223 | 8,252 | 8,195 | 8,245 | 8,294 | 8,262 | 8,284 | 8,297 | 8,307 | 7,949 |
| 6 | 31 | 1004 | 54,688 | 5,930 | 8,241 | 7,916 | 8,287 | 8,285 | 8,242 | 6,839 | | | 948 |
| 6 | 37 | 2 | 80,870 | 8,337 | 8,021 | 7,997 | 7,766 | 8,285 | 8,275 | 8,276 | 8,248 | 7,729 | 7,936 |
| 6 | 37 | 16 | 79,234 | 8,297 | 8,076 | 8,165 | 8,013 | 8,172 | 8,120 | 8,260 | 8,283 | 6,215 | 7,633 |
| 6 | 37 | 30 | 10,018 | 6,786 | 3,232 | | | | | | | | |
| 6 | 37 | 31 | 9,108 | 4,240 | 4,868 | | | | | | | | |
| 6 | 37 | 113 | 81,092 | 8,395 | 8,285 | 8,118 | 7,643 | 8,284 | 8,070 | 7,940 | 8,229 | 8,034 | 8,094 |
| 6 | 37 | 1002 | 80,389 | 8,390 | 8,164 | 7,922 | 8,023 | 8,337 | 8,340 | 8,265 | 8,262 | 7,692 | 6,994 |
| 6 | 37 | 1103 | 80,154 | 8,334 | 8,330 | 8,285 | 7,498 | 8,236 | 8,216 | 8,167 | 7,820 | 7,116 | 8,152 |
| 6 | 37 | 1201 | 81,400 | 8,047 | 8,334 | 8,298 | 8,086 | 7,979 | 8,005 | 7,914 | 8,318 | 8,256 | 8,163 |
| 6 | 37 | 1301 | 64,605 | 8,392 | 8,305 | 8,131 | 8,191 | 8,194 | 8,243 | 8,265 | 6,884 | | |
| 6 | 37 | 1302 | 16,402 | | | | | | | | 1,123 | 7,333 | 7,946 |
| 6 | 37 | 1601 | 35,691 | 8,388 | 8,268 | 8,250 | 8,142 | 2,643 | | | | | |
| 6 | 37 | 1602 | 32,186 | | | | | | 3,861 | 7,925 | 7,493 | 6,937 | 5,970 |
| 6 | 37 | 1701 | 82,294 | 8,365 | 8,312 | 8,342 | 8,271 | 8,349 | 8,144 | 8,275 | 8,341 | 8,133 | 7,762 |
| 6 | 37 | 2005 | 81,423 | 8,351 | 8,376 | 8,146 | 8,153 | 8,239 | 8,369 | 8,335 | 8,289 | 8,304 | 6,861 |
| 6 | 37 | 4002 | 80,025 | 8,384 | 8,050 | 7,775 | 8,114 | 8,329 | 8,100 | 7,983 | 8,279 | 7,224 | 7,787 |
| 6 | 37 | 5001 | 26,215 | 8,257 | 7,684 | 8,247 | 2,027 | | | | | | |
| 6 | 37 | 5005 | 45,428 | | | | 4,712 | 6,204 | 7,824 | 7,397 | 7,198 | 5,678 | 6,415 |
| 6 | 37 | 6002 | 2,783 | 2,783 | | | | | | | | | |
| 6 | 37 | 6012 | 77,052 | 5,181 | 7,736 | 8,317 | 8,131 | 7,890 | 8,217 | 7,816 | 8,221 | 8,134 | 7,409 |
| 6 | 37 | 9002 | 6,492 | 6,492 | | | | | | | | | |
| 6 | 37 | 9033 | 74,427 | 1,394 | 8,151 | 7,606 | 8,137 | 8,265 | 8,127 | 8,227 | 8,327 | 8,215 | 7,978 |

List of AQS NOx Monitoring Stations and Number of Valid Observations by Year

| State | County | | | | | | | | | | | | | |
|-------|--------|---------|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|--|
| Code | Code | Site ID | Total Obs | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | |
| 6 | 39 | 4 | 80,501 | 8,144 | 8,135 | 8,093 | 7,553 | 8,110 | 8,302 | 8,130 | 7,870 | 8,283 | 7,881 | |
| 6 | 41 | 1 | 33,188 | | | | | | | 8,350 | 8,306 | 8,328 | 8,204 | |
| 6 | 43 | 3 | 4,388 | | | | | | 2,228 | 2,160 | | | | |
| 6 | 43 | 33 | 3,813 | | 2,660 | 1,153 | | | | | | | | |
| 6 | 45 | 8 | 1,496 | | | | | | | | | 1,496 | | |
| 6 | 47 | 3 | 78,228 | 5,472 | 8,198 | 8,236 | 8,248 | 8,280 | 8,109 | 8,189 | 7,872 | 7,711 | 7,913 | |
| 6 | 53 | 1003 | 80,845 | 7,927 | 7,257 | 7,654 | 7,763 | 8,547 | 8,278 | 8,460 | 8,326 | 8,230 | 8,403 | |
| 6 | 55 | 3 | 33,169 | | | | | | | 8,253 | 8,355 | 8,269 | 8,292 | |
| 6 | 57 | 5 | 17,379 | | | | | | | | 6,248 | 4,606 | 6,525 | |
| 6 | 59 | 1 | 3,497 | 3,497 | | | | | | | | | | |
| 6 | 59 | 7 | 75,483 | 2,711 | 8,244 | 8,147 | 8,162 | 8,277 | 7,786 | 8,179 | 8,345 | 8,218 | 7,414 | |
| 6 | 59 | 1003 | 80,837 | 8,391 | 8,356 | 8,286 | 7,812 | 7,786 | 8,201 | 8,196 | 8,262 | 8,237 | 7,310 | |
| 6 | 59 | 5001 | 78,177 | 8,342 | 7,517 | 8,303 | 7,736 | 8,189 | 8,210 | 8,329 | 8,187 | 8,042 | 5,322 | |
| 6 | 61 | 6 | 82,349 | 8,129 | 8,198 | 7,972 | 8,365 | 8,270 | 8,372 | 8,210 | 8,262 | 8,293 | 8,278 | |
| 6 | 61 | 7 | 1,393 | | | | 1,393 | | | | | | | |
| 6 | 65 | 4 | 18,616 | | | | | | | | 7,344 | 5,714 | 5,558 | |
| 6 | 65 | 9 | 18,578 | | | | | | | | 3,571 | 7,994 | 7,013 | |
| 6 | 65 | 12 | 76,214 | 7,962 | 6,992 | 7,741 | 7,451 | 6,882 | 8,057 | 8,192 | 7,757 | 7,714 | 7,466 | |
| 6 | 65 | 1003 | 17,103 | | | | | | | | 1,591 | 8,264 | 7,248 | |
| 6 | 65 | 5001 | 80,024 | 8,394 | 8,063 | 7,903 | 7,857 | 8,008 | 7,958 | 8,064 | 8,155 | 7,724 | 7,898 | |
| 6 | 65 | 8001 | 77,716 | 8,325 | 7,456 | 8,181 | 7,317 | 8,257 | 8,356 | 8,275 | 8,142 | 7,713 | 5,694 | |
| 6 | 65 | 8005 | 23,042 | | | | | | | | 8,305 | 7,062 | 7,675 | |
| 6 | 65 | 9001 | 78,923 | 8,244 | 8,114 | 7,425 | 7,425 | 8,018 | 7,896 | 7,949 | 8,084 | 8,224 | 7,544 | |
| 6 | 67 | 2 | 72,946 | 7,947 | 8,293 | 8,291 | 8,047 | 6,759 | 4,890 | 6,472 | 5,953 | 8,151 | 8,143 | |
| 6 | 67 | 6 | 73,005 | 7,759 | 7,845 | 5,590 | 2,894 | 8,246 | 8,223 | 7,966 | 8,118 | 8,330 | 8,034 | |
| 6 | 67 | 10 | 80,924 | 7,918 | 8,120 | 7,103 | 8,141 | 8,373 | 8,345 | 8,134 | 8,148 | 8,346 | 8,296 | |
| 6 | 67 | 11 | 74,667 | 8,003 | 6,824 | 5,176 | 8,283 | 7,907 | 6,926 | 8,057 | 8,292 | 7,561 | 7,638 | |
| 6 | 67 | 12 | 80,151 | 7,432 | 7,985 | 8,103 | 8,255 | 8,107 | 8,296 | 7,960 | 8,103 | 8,132 | 7,778 | |
| 6 | 67 | 13 | 59,179 | 7,149 | 7,771 | 6,863 | 8,242 | 8,057 | 8,293 | 8,005 | 4,799 | | | |
| 6 | 67 | 14 | 17,041 | | | | | | | | 1,083 | 7,990 | 7,968 | |
| 6 | 71 | 1 | 80,568 | 7,826 | 7,849 | 8,065 | 8,201 | 8,182 | 7,828 | 8,003 | 8,320 | 8,265 | 8,029 | |
| 6 | 71 | 306 | 81,883 | 7,863 | 8,159 | 8,129 | 8,310 | 8,242 | 8,238 | 8,348 | 8,328 | 8,288 | 7,978 | |
| 6 | 71 | 1004 | 80,064 | 8,379 | 8,277 | 8,300 | 8,319 | 8,293 | 7,700 | 7,410 | 8,245 | 7,222 | 7,919 | |
| 6 | 71 | 1234 | 72,409 | 6,891 | 7,665 | 6,823 | 7,914 | 6,951 | 7,610 | 6,939 | 7,097 | 7,125 | 7,394 | |
| 6 | 71 | 2002 | 77,474 | 8,389 | 7,051 | 7,966 | 7,044 | 8,196 | 8,220 | 7,765 | 7,431 | 8,097 | 7,315 | |
| 6 | 71 | 9004 | 80,180 | 8,341 | 7,995 | 8,214 | 8,000 | 8,366 | 8,250 | 8,022 | 8,316 | 8,234 | 6,442 | |
| 6 | 73 | 1 | 81,990 | 8,195 | 8,192 | 8,205 | 8,241 | 8,098 | 8,143 | 8,081 | 8,295 | 8,235 | 8,305 | |
| 6 | 73 | 3 | 80,351 | 8,254 | 7,934 | 8,051 | 8,307 | 8,255 | 7,772 | 7,581 | 7,940 | 8,240 | 8,017 | |
| 6 | 73 | 5 | 8,411 | 8,411 | | | | | | | | | | |
| 6 | 73 | 6 | 81,301 | 8,048 | 7,977 | 8,140 | 8,309 | 8,283 | 8,093 | 8,047 | 8,061 | 8,090 | 8,253 | |
| 6 | 73 | 1002 | 82,915 | 8,490 | 8,642 | 8,486 | 8,644 | 8,611 | 8,207 | 7,885 | 7,554 | 8,311 | 8,085 | |
| 6 | 73 | 1006 | 80,065 | 8,308 | 8,141 | 7,626 | 8,265 | 8,077 | 8,237 | 8,189 | 8,340 | 6,785 | 8,097 | |
| 6 | 73 | 1007 | 35,815 | 7,246 | 8,241 | 7,990 | 8,167 | 4,171 | | | | | | |
| 6 | 73 | 1008 | 81,981 | 8,600 | 8,431 | 8,516 | 8,538 | 8,164 | 8,152 | 8,098 | 8,021 | 7,540 | 7,921 | |
| 6 | 73 | 1009 | 1,353 | 1,353 | | | | | | | | | | |
| 6 | 73 | 1010 | 43,563 | | | | | 3,817 | 8,104 | 8,193 | 7,796 | 8,134 | 7,519 | |
| 6 | 73 | 2007 | 79,622 | 7,965 | 7,887 | 8,065 | 8,246 | 7,555 | 7,922 | 8,221 | 7,872 | 7,735 | 8,154 | |
| 6 | 75 | 5 | 32,193 | | | | | | | 8,166 | 7,907 | 8,129 | 7,991 | |
| 6 | 77 | 1002 | 82,853 | 8,361 | 8,335 | 8,353 | 8,272 | 8,218 | 8,344 | 8,206 | 8,338 | 8,122 | 8,304 | |
| 6 | 77 | 3003 | 28,932 | 8,241 | 8,222 | 7,995 | 4,474 | | | | | | | |
| 6 | 77 | 3005 | 40,233 | | | | | | 8,241 | 7,922 | 8,285 | 7,940 | 7,845 | |
| 6 | 79 | 2001 | 30,921 | 6,979 | 6,027 | 6,426 | 7,034 | 4,455 | | | | | | |
| 6 | 79 | 2002 | 37,181 | 8,111 | 8,155 | 6,761 | 8,191 | 5,963 | | | | | | |
| 6 | 79 | 2006 | 9,263 | | | | | 2,336 | 6,927 | | | | | |
| 6 | 79 | 3001 | 70,159 | 2,775 | 8,202 | 8,123 | 8,014 | 7,799 | 7,584 | 7,479 | 7,394 | 5,829 | 6,960 | |
| 6 | 79 | 4002 | 74,032 | 6,755 | 8,088 | 8,120 | 8,131 | 8,002 | 6,946 | 6,922 | 7,667 | 7,039 | 6,362 | |
| 6 | 79 | 8001 | 78,257 | 7,654 | 8,009 | 8,227 | 7,922 | 7,393 | 8,070 | 8,243 | 8,141 | 7,221 | 7,377 | |
| 6 | 81 | 1001 | 32,621 | | | | | | | 8,219 | 8,210 | 8,169 | 8,023 | |
| 6 | 83 | 8 | 78,059 | 8,225 | 8,263 | 8,248 | 8,220 | 8,118 | 8,266 | 7,745 | 8,263 | 6,549 | 6,162 | |
| 6 | 83 | 11 | 51,904 | | | 3,158 | 8,153 | 8,334 | 4,826 | 2,804 | 8,160 | 8,206 | 8,263 | |
| 6 | 83 | 1008 | 59,881 | | 3,787 | 8,117 | 7,482 | 8,282 | 4,808 | 2,752 | 8,128 | 8,346 | 8,179 | |
| 6 | 83 | 1013 | 63,510 | 6,976 | 6,221 | 6,156 | 8,263 | 7,076 | 5,752 | 5,510 | 4,064 | 7,277 | 6,215 | |
| 6 | 83 | 1014 | 67,452 | 7,764 | 8,035 | 7,913 | 7,917 | 6,882 | 7,066 | 6,545 | 6,830 | 5,822 | 2,678 | |
| 6 | 83 | 1018 | 80,185 | 8,158 | 8,277 | 8,211 | 8,187 | 8,142 | 8,071 | 8,013 | 7,936 | 7,678 | 7,512 | |
| 6 | 83 | 1021 | 63,115 | 7,363 | 4,540 | 4,531 | 3,677 | 7,931 | 8,114 | 8,018 | 6,892 | 6,824 | 5,225 | |
| 6 | 83 | 1025 | 78,307 | 8,102 | 8,163 | 8,120 | 8,142 | 7,976 | 7,807 | 8,178 | 7,949 | 7,004 | 6,866 | |
| 6 | 83 | 2004 | 69,839 | 7,884 | 4,275 | 7,869 | 7,802 | 8,080 | 7,919 | 7,387 | 4,910 | 6,856 | 6,857 | |
| 6 | 83 | 2011 | 80,800 | 8,214 | 8,224 | 7,521 | 8,287 | 8,180 | 8,244 | 7,986 | 7,871 | 8,263 | 8,010 | |

List of AQS NOx Monitoring Stations and Number of Valid Observations by Year

| State | County | | | | | | | | | | | | |
|-------|--------|---------|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Code | Code | Site ID | Total Obs | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 6 | 83 | 4003 | 42,798 | 1,745 | 3,605 | 5,478 | 7,041 | 5,971 | 5,014 | 4,668 | 4,040 | 2,934 | 2,302 |
| 6 | 85 | 5 | 32,697 | | | | | | | 8,324 | 8,351 | 7,845 | 8,177 |
| 6 | 87 | 3 | 78,266 | 8,275 | 8,377 | 6,468 | 7,005 | 8,187 | 8,283 | 8,561 | 8,402 | 7,846 | 6,862 |
| 6 | 95 | 4 | 32,930 | | | | | | | 8,284 | 8,264 | 8,111 | 8,271 |
| 6 | 95 | 6 | 14,338 | | | | | | | 6,211 | 8,127 | | |
| 6 | 97 | 3 | 33,115 | | | | | | | 8,336 | 8,364 | 8,133 | 8,282 |
| 6 | 99 | 5 | 41,496 | 8,194 | 8,382 | 8,257 | 8,295 | 8,368 | | | | | |
| 6 | 99 | 6 | 79,937 | 7,199 | 8,265 | 8,202 | 8,288 | 7,318 | 8,227 | 8,272 | 8,096 | 8,079 | 7,991 |
| 6 | 101 | 3 | 79,681 | 7,590 | 8,242 | 8,226 | 7,978 | 8,092 | 7,751 | 8,128 | 8,091 | 7,487 | 8,096 |
| 6 | 107 | 2002 | 82,846 | 8,352 | 8,324 | 8,300 | 8,326 | 8,068 | 8,255 | 8,278 | 8,374 | 8,315 | 8,254 |
| 6 | 111 | 5 | 11,518 | 6,622 | 4,896 | | | | | | | | |
| 6 | 111 | 7 | 29,167 | 8,069 | 8,331 | 8,100 | 4,667 | | | | | | |
| 6 | 111 | 1004 | 28,984 | 7,755 | 8,239 | 8,270 | 4,720 | | | | | | |
| 6 | 111 | 2002 | 79,481 | 8,025 | 8,010 | 8,000 | 7,576 | 7,749 | 7,879 | 7,755 | 8,095 | 8,250 | 8,142 |
| 6 | 111 | 2003 | 27,909 | 7,483 | 7,862 | 7,981 | 4,583 | | | | | | |
| 6 | 111 | 3001 | 79,110 | 7,687 | 7,409 | 7,671 | 7,909 | 7,966 | 8,102 | 8,003 | 7,974 | 8,164 | 8,225 |
| 6 | 113 | 4 | 80,169 | 6,871 | 8,241 | 8,264 | 8,321 | 8,363 | 8,349 | 8,328 | 8,109 | 8,299 | 7,024 |
| 8 | 41 | 6001 | 6,377 | 6,377 | | | | | | | | | |
| 8 | 41 | 6004 | 6,184 | 6,184 | | | | | | | | | |
| 8 | 41 | 6011 | 6,380 | 6,380 | | | | | | | | | |
| 8 | 41 | 6018 | 6,336 | 6,336 | | | | | | | | | |
| 8 | 67 | 1004 | 49,337 | | | | 5,652 | 7,994 | 8,286 | 7,218 | 6,981 | 7,204 | 6,002 |
| 8 | 67 | 7001 | 49,277 | 3,267 | 657 | 5,738 | 758 | 1,529 | 8,515 | 7,222 | 8,352 | 5,059 | 8,180 |
| 8 | 67 | 7003 | 63,260 | 2,048 | 7,996 | 7,291 | 7,614 | 4,541 | 7,818 | 7,888 | 4,795 | 6,535 | 6,734 |
| 9 | 1 | 9003 | 81,542 | 8,231 | 8,506 | 8,602 | 8,182 | 8,179 | 8,046 | 7,815 | 8,187 | 7,679 | 8,115 |
| 9 | 3 | 1003 | 83,849 | 8,648 | 8,280 | 8,452 | 8,569 | 8,371 | 8,312 | 8,560 | 8,320 | 8,307 | 8,030 |
| 9 | 5 | 4 | 31,040 | | | | | | 2,074 | 8,127 | 8,107 | 7,573 | 5,159 |
| 9 | 9 | 27 | 57,439 | | | | 7,557 | 8,444 | 8,474 | 8,620 | 8,046 | 8,036 | 8,262 |
| 9 | 9 | 1123 | 26,723 | 8,609 | 8,603 | 8,551 | 960 | | | | | | |
| 9 | 9 | 9005 | 13,052 | 4,321 | 4,210 | 4,521 | | | | | | | |
| 9 | 13 | 1001 | 4,022 | 4,022 | | | | | | | | | |
| 10 | 3 | 2004 | 64,453 | 23 | 7,774 | 7,835 | 6,213 | 7,892 | 7,428 | 5,462 | 8,062 | 5,751 | 8,013 |
| 11 | 1 | 25 | 85,348 | 8,512 | 8,220 | 8,661 | 8,711 | 8,483 | 8,647 | 8,608 | 8,585 | 8,400 | 8,521 |
| 11 | 1 | 41 | 85,522 | 8,390 | 8,657 | 8,484 | 8,532 | 8,470 | 8,708 | 8,611 | 8,534 | 8,572 | 8,564 |
| 11 | 1 | 43 | 85,478 | 8,278 | 8,658 | 8,498 | 8,710 | 8,496 | 8,653 | 8,638 | 8,324 | 8,577 | 8,646 |
| 12 | 11 | 31 | 31,906 | | | | | | 697 | 7,507 | 7,150 | 8,394 | 8,158 |
| 12 | 11 | 8002 | 21,863 | | | | | | 490 | 5,280 | 4,695 | 5,676 | 5,722 |
| 12 | 31 | 32 | 27,378 | | | | | | 696 | 7,962 | 2,066 | 8,211 | 8,443 |
| 12 | 33 | 4 | 26,965 | | | | | | 659 | 8,327 | 8,249 | 8,341 | 1,389 |
| 12 | 57 | 81 | 13,502 | | | | | | 729 | 8,491 | 2,114 | 2,168 | |
| 12 | 57 | 1065 | 20,407 | | | | | | 742 | 8,450 | 2,168 | 2,177 | 6,870 |
| 12 | 57 | 3002 | 13,520 | | | | | | 723 | 7,364 | 5,433 | | |
| 12 | 81 | 4012 | 10,000 | | | | | | 711 | 7,661 | 1,628 | | |
| 12 | 86 | 27 | 26,992 | | | | | | 604 | 7,011 | 6,320 | 6,418 | 6,639 |
| 12 | 86 | 4002 | 33,572 | | | | | | 736 | 7,951 | 8,427 | 8,190 | 8,268 |
| 12 | 95 | 2002 | 20,927 | | | | | | 729 | 7,912 | 2,089 | 2,062 | 8,135 |
| 12 | 99 | 20 | 16,191 | | | | | | | | 1,532 | 7,396 | 7,263 |
| 12 | 99 | 1004 | 12,273 | | | | | | 734 | 8,657 | 2,882 | | |
| 12 | 103 | 18 | 20,077 | | | | | | 731 | 8,127 | 2,050 | 2,066 | 7,103 |
| 12 | 115 | 1006 | 1,750 | | | | | | | | | | 1,750 |
| 13 | 89 | 2 | 81,508 | 8,485 | 8,227 | 6,487 | 8,613 | 8,588 | 8,587 | 7,907 | 7,990 | 8,110 | 8,514 |
| 13 | 89 | 3001 | 49,202 | 8,596 | 8,600 | 6,719 | 8,476 | 8,589 | 8,222 | | | | |
| 13 | 121 | 48 | 69,278 | 8,537 | 8,575 | 8,433 | 8,407 | 8,179 | 8,049 | 8,103 | 8,224 | 2,771 | |
| 13 | 223 | 3 | 80,494 | 8,538 | 6,134 | 8,347 | 8,050 | 8,375 | 8,044 | 8,510 | 7,807 | 8,259 | 8,430 |
| 13 | 247 | 1 | 84,026 | 8,605 | 8,415 | 8,476 | 8,578 | 8,434 | 8,150 | 8,331 | 8,519 | 8,051 | 8,467 |
| 16 | 1 | 10 | 13,275 | | | | | | | | | 5,498 | 7,777 |
| 16 | 1 | 16 | 740 | 740 | | | | | | | | | |
| 16 | 1 | 19 | 4,345 | | | | | | | 1,114 | | | 3,231 |
| 16 | 21 | 3 | 10,031 | | 1,446 | 8,585 | | | | | | | |
| 16 | 39 | 2 | 16,442 | | 6,646 | 8,376 | 1,420 | | | | | | |
| 16 | 55 | 3 | 25,762 | | | | | 8,180 | 7,853 | 3,825 | 2,658 | 1,365 | 1,881 |
| 17 | 31 | 63 | 84,356 | 8,643 | 8,656 | 8,486 | 8,486 | 8,383 | 8,425 | 7,906 | 8,595 | 8,347 | 8,429 |
| 17 | 31 | 72 | 33,422 | 3,211 | 3,162 | 3,449 | 3,189 | 3,889 | 3,524 | 3,602 | 2,904 | 3,163 | 3,329 |
| 17 | 31 | 75 | 8,360 | 8,360 | | | | | | | | | |
| 17 | 31 | 76 | 77,073 | | 8,640 | 8,574 | 8,415 | 8,608 | 8,686 | 8,510 | 8,714 | 8,648 | 8,278 |
| 17 | 31 | 3103 | 84,120 | 8,421 | 8,413 | 8,440 | 8,251 | 8,530 | 8,612 | 8,185 | 8,491 | 8,381 | 8,396 |
| 17 | 31 | 4002 | 85,664 | 8,676 | 8,234 | 8,635 | 8,655 | 8,602 | 8,691 | 8,632 | 8,430 | 8,650 | 8,459 |
| 17 | 31 | 4201 | 55,096 | 7,953 | 3,575 | 5,307 | 3,553 | 3,850 | 4,143 | 4,152 | 5,772 | 8,577 | 8,214 |

List of AQS NOx Monitoring Stations and Number of Valid Observations by Year

| State | County | | | | | | | | | | | | |
|-------|--------|---------|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Code | Code | Site ID | Total Obs | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 17 | 31 | 8003 | 17,306 | 8,655 | 8,651 | | | | | | | | |
| 17 | 97 | 1007 | 4,909 | 2,586 | 2,323 | | | | | | | | |
| 17 | 163 | 10 | 84,570 | 8,661 | 8,257 | 8,566 | 8,667 | 8,364 | 8,059 | 8,575 | 8,081 | 8,703 | 8,637 |
| 17 | 197 | 1011 | 24,775 | 3,110 | 3,751 | 3,383 | 3,193 | 4,260 | 3,884 | 3,194 | | | |
| 18 | 51 | 10 | 32,440 | 8,319 | 8,187 | 7,807 | 8,127 | | | | | | |
| 18 | 63 | 1 | 28,741 | | | | 7,987 | 8,062 | 8,441 | 4,251 | | | |
| 18 | 63 | 2 | 39,172 | | | | 6,324 | 7,346 | 7,758 | 7,747 | 7,899 | 2,098 | |
| 18 | 63 | 3 | 25,742 | | | | 7,196 | 7,269 | 7,597 | 3,680 | | | |
| 18 | 89 | 22 | 74,335 | 7,430 | 7,166 | 8,254 | 7,280 | 8,533 | 7,489 | 5,022 | 7,475 | 7,836 | 7,850 |
| 18 | 97 | 73 | 81,182 | 8,274 | 7,086 | 7,939 | 7,474 | 8,583 | 8,403 | 8,474 | 8,534 | 8,228 | 8,187 |
| 18 | 141 | 15 | 35,876 | | | | | | 4,918 | 8,467 | 8,196 | 7,835 | 6,460 |
| 18 | 141 | 1008 | 44,199 | 8,698 | 8,585 | 8,533 | 8,187 | 6,637 | 3,559 | | | | |
| 18 | 147 | 8 | 5,978 | 5,978 | | | | | | | | | |
| 18 | 163 | 12 | 68,818 | 8,697 | 8,701 | 8,690 | 6,661 | 8,711 | 8,668 | 6,372 | 8,397 | 3,921 | |
| 18 | 163 | 21 | 10,798 | | | | | | | | | 3,582 | 7,216 |
| 19 | 113 | 33 | 16,329 | 2,045 | 5,020 | 4,598 | 4,666 | | | | | | |
| 19 | 153 | 30 | 33,514 | | | | | | 676 | 7,725 | 8,667 | 8,586 | 7,860 |
| 19 | 153 | 58 | 43,845 | 5,320 | 7,384 | 8,555 | 8,324 | 7,419 | 6,843 | | | | |
| 19 | 163 | 14 | 38,454 | 6,048 | 8,137 | 8,400 | 8,670 | 7,199 | | | | | |
| 19 | 163 | 15 | 41,754 | | | | | | 8,107 | 8,614 | 8,412 | 8,553 | 8,068 |
| 20 | 107 | 2 | 47,280 | 7,400 | 7,891 | 6,828 | 3,320 | 5,368 | 4,437 | 4,994 | 2,710 | 2,550 | 1,782 |
| 20 | 173 | 10 | 72,461 | | 4,863 | 8,681 | 8,405 | 8,294 | 8,008 | 8,314 | 8,629 | 8,673 | 8,594 |
| 20 | 191 | 2 | 75,760 | 6,608 | 6,466 | 8,350 | 6,058 | 8,178 | 8,071 | 7,262 | 8,624 | 8,568 | 7,575 |
| 20 | 209 | 21 | 83,554 | 8,111 | 8,581 | 8,184 | 8,144 | 8,551 | 8,647 | 8,316 | 8,104 | 8,686 | 8,230 |
| 21 | 111 | 67 | 6,695 | | | | | | | | | | 6,695 |
| 21 | 177 | 1004 | 8,373 | 5,609 | 2,764 | | | | | | | | |
| 21 | 221 | 8001 | 8,240 | | | | | | | 6,911 | 1,329 | | |
| 22 | 5 | 4 | 48,571 | | | | | 7,276 | 8,270 | 8,326 | 8,158 | 8,399 | 8,142 |
| 22 | 19 | 8 | 46,269 | | | | | 7,173 | 8,284 | 7,630 | 7,621 | 8,190 | 7,371 |
| 22 | 19 | 9 | 32,038 | 8,423 | 7,844 | 7,327 | 8,444 | | | | | | |
| 22 | 33 | 3 | 82,866 | 8,396 | 8,331 | 8,256 | 8,201 | 8,304 | 8,252 | 8,317 | 8,181 | 8,311 | 8,317 |
| 22 | 33 | 9 | 83,828 | 8,309 | 8,524 | 8,486 | 8,474 | 8,418 | 8,413 | 8,459 | 8,210 | 8,277 | 8,258 |
| 22 | 33 | 13 | 77,479 | 8,075 | 6,126 | 7,732 | 8,211 | 8,033 | 7,792 | 8,117 | 7,800 | 8,136 | 7,457 |
| 22 | 33 | 1001 | 81,071 | 7,738 | 8,008 | 8,261 | 8,371 | 8,314 | 8,458 | 8,386 | 8,191 | 8,191 | 7,153 |
| 22 | 47 | 7 | 76,639 | 7,906 | 7,861 | 7,886 | 7,242 | 8,183 | 8,099 | 7,760 | 7,113 | 8,166 | 6,423 |
| 22 | 47 | 9 | 78,433 | 6,968 | 7,707 | 7,964 | 7,653 | 8,092 | 8,487 | 7,865 | 7,904 | 8,411 | 7,382 |
| 22 | 47 | 12 | 77,588 | 7,448 | 7,110 | 7,832 | 7,723 | 7,544 | 7,649 | 8,110 | 7,798 | 8,277 | 8,097 |
| 22 | 51 | 1001 | 79,489 | 8,292 | 8,169 | 8,362 | 8,048 | 7,867 | 8,214 | 8,008 | 6,926 | 7,858 | 7,745 |
| 22 | 63 | 2 | 77,817 | 7,858 | 7,626 | 7,593 | 7,831 | 6,936 | 7,804 | 8,441 | 7,744 | 8,381 | 7,603 |
| 22 | 71 | 12 | 37,129 | 8,095 | 7,050 | 8,199 | 8,321 | 5,464 | | | | | |
| 22 | 121 | 1 | 81,716 | 8,309 | 7,833 | 8,299 | 8,032 | 8,077 | 8,351 | 8,432 | 7,937 | 8,077 | 8,369 |
| 23 | 3 | 1100 | 25,505 | | | | | | 5,053 | 5,945 | 6,048 | 4,884 | 3,575 |
| 23 | 5 | 27 | 47,206 | 4,921 | 8,641 | 8,488 | 8,544 | 8,450 | 8,162 | | | | |
| 23 | 5 | 29 | 23,872 | | | | | | | | 7,580 | 8,048 | 8,244 |
| 23 | 9 | 102 | 5,510 | | | | 1,250 | 1,575 | 1,256 | 1,429 | | | |
| 23 | 31 | 3002 | 31,322 | 6,742 | 7,276 | 7,578 | 5,621 | 4,105 | | | | | |
| 24 | 3 | 19 | 9,448 | 4,838 | 3,874 | 736 | | | | | | | |
| 24 | 5 | 3001 | 76,488 | 8,107 | 8,209 | 8,250 | 8,048 | 7,530 | 4,617 | 7,341 | 8,419 | 7,781 | 8,186 |
| 24 | 33 | 30 | 22,285 | | | | | 4,150 | 6,499 | 8,400 | 3,236 | | |
| 24 | 510 | 40 | 79,911 | 7,933 | 8,182 | 8,099 | 8,035 | 8,265 | 8,248 | 6,255 | 8,339 | 8,471 | 8,084 |
| 24 | 510 | 50 | 1,263 | 1,263 | | | | | | | | | |
| 25 | 1 | 2 | 19,329 | 4,484 | 1,266 | 1,244 | 4,199 | 3,944 | 4,192 | | | | |
| 25 | 5 | 1002 | 3,269 | | 3,269 | | | | | | | | |
| 25 | 5 | 1005 | 5,979 | 5,979 | | | | | | | | | |
| 25 | 9 | 5 | 3,563 | | 3,563 | | | | | | | | |
| 25 | 9 | 2006 | 80,336 | 8,017 | 7,873 | 8,232 | 7,650 | 7,989 | 8,233 | 7,959 | 7,949 | 8,084 | 8,350 |
| 25 | 9 | 4004 | 34,807 | 7,516 | 6,062 | 3,716 | 3,720 | 3,627 | 3,305 | 2,780 | 2,872 | 1,209 | |
| 25 | 9 | 5005 | 53,835 | | | | 4,820 | 7,865 | 8,063 | 8,303 | 8,457 | 8,428 | 7,899 |
| 25 | 13 | 3 | 10,288 | 8,162 | 2,126 | | | | | | | | |
| 25 | 13 | 8 | 80,831 | 8,232 | 8,367 | 8,069 | 7,523 | 7,410 | 8,277 | 8,249 | 8,206 | 8,265 | 8,233 |
| 25 | 13 | 16 | 83,234 | 8,387 | 8,378 | 8,184 | 8,422 | 8,104 | 8,375 | 8,363 | 8,414 | 8,355 | 8,252 |
| 25 | 15 | 4002 | 75,045 | 7,985 | 6,664 | 8,397 | 6,936 | 8,225 | 7,825 | 5,938 | 7,685 | 7,516 | 7,874 |
| 25 | 21 | 3003 | 34,300 | | 4,121 | 3,334 | 3,392 | 2,893 | 3,604 | 4,141 | 4,162 | 3,974 | 4,679 |
| 25 | 25 | 2 | 77,852 | 8,155 | 6,770 | 6,392 | 8,199 | 8,333 | 8,250 | 8,214 | 8,372 | 6,938 | 8,229 |
| 25 | 25 | 21 | 15,048 | 7,799 | 7,249 | | | | | | | | |
| 25 | 25 | 40 | 75,497 | 8,544 | 8,505 | 8,348 | 8,543 | 8,168 | 5,550 | 2,259 | 8,544 | 8,512 | 8,524 |
| 25 | 25 | 41 | 43,330 | 7,593 | 5,701 | 4,142 | 3,674 | 4,051 | 3,854 | 1,172 | 4,170 | 4,185 | 4,788 |
| 25 | 25 | 42 | 81,337 | 8,012 | 7,431 | 8,341 | 7,631 | 8,342 | 8,307 | 8,423 | 8,402 | 8,234 | 8,214 |

List of AQS NOx Monitoring Stations and Number of Valid Observations by Year

| State | County | | | | | | | | | | | | |
|-------|--------|---------|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Code | Code | Site ID | Total Obs | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 25 | 27 | 20 | 22,367 | 8,355 | 8,266 | 5,746 | | | | | | | |
| 25 | 27 | 23 | 56,863 | | | | 7,779 | 8,295 | 8,145 | 8,329 | 8,119 | 8,270 | 7,926 |
| 26 | 65 | 12 | 10,776 | 6,551 | | | | | | | | | 4,225 |
| 26 | 81 | 20 | 41,555 | | 7,123 | 8,032 | 7,952 | 7,931 | 8,567 | 1,950 | | | |
| 26 | 113 | 1 | 1,266 | | | | | | | | | | 1,266 |
| 26 | 163 | 16 | 43,731 | 1,679 | 8,470 | 8,012 | 6,777 | 8,344 | 8,627 | 1,822 | | | |
| 26 | 163 | 19 | 76,690 | 2,525 | 8,024 | 8,431 | 7,735 | 7,455 | 8,607 | 8,462 | 8,506 | 8,657 | 8,288 |
| 27 | 3 | 1002 | 63,681 | | | 7,796 | 8,296 | 8,162 | 8,313 | 8,132 | 7,649 | 7,169 | 8,164 |
| 27 | 17 | 7416 | 20,815 | | 2,485 | 6,879 | 6,275 | | | | | | 5,176 |
| 27 | 37 | 20 | 74,725 | 8,527 | 8,010 | 8,097 | 6,956 | 6,526 | 8,343 | 5,234 | 7,031 | 8,194 | 7,807 |
| 27 | 37 | 423 | 74,590 | 8,646 | 8,641 | 8,609 | 7,870 | 7,931 | 7,610 | 6,852 | 6,131 | 6,427 | 5,873 |
| 27 | 49 | 5302 | 1,682 | | | 1,134 | 548 | | | | | | |
| 27 | 53 | 957 | 16,618 | 7,990 | 8,628 | | | | | | | | |
| 27 | 123 | 864 | 16,271 | 8,239 | 8,032 | | | | | | | | |
| 28 | 3 | 4 | 8,379 | 6,042 | 2,337 | | | | | | | | |
| 28 | 33 | 2 | 11,687 | 7,657 | 4,030 | | | | | | | | |
| 28 | 45 | 1 | 26,137 | 6,202 | 5,934 | 6,939 | 7,062 | | | | | | |
| 28 | 45 | 3 | 1,858 | | | | | 1,858 | | | | | |
| 28 | 59 | 6 | 46,602 | | | | 5,937 | 7,098 | 5,592 | 5,935 | 7,296 | 6,665 | 8,079 |
| 28 | 93 | 1 | 8,142 | | | | 6,788 | 1,354 | | | | | |
| 29 | 39 | 1 | 7,889 | 2,934 | 3,324 | 1,631 | | | | | | | |
| 29 | 47 | 5 | 79,685 | 8,041 | 7,801 | 7,889 | 7,931 | 7,996 | 8,409 | 8,569 | 8,197 | 8,462 | 6,390 |
| 29 | 47 | 6 | 13,489 | | | 7,362 | 6,127 | | | | | | |
| 29 | 77 | 36 | 31,999 | | | | | | | 8,414 | 8,657 | 8,571 | 6,357 |
| 29 | 95 | 34 | 63,962 | | | 4,723 | 7,752 | 8,471 | 8,639 | 8,687 | 8,733 | 8,653 | 8,304 |
| 29 | 129 | 1 | 6,478 | 6,478 | | | | | | | | | |
| 29 | 165 | 23 | 15,422 | | | 8,134 | 7,288 | | | | | | |
| 29 | 183 | 1002 | 79,648 | 7,974 | 7,760 | 8,089 | 7,547 | 8,603 | 8,549 | 8,579 | 7,932 | 8,354 | 6,261 |
| 29 | 186 | 5 | 65,244 | 3,308 | 5,424 | 4,485 | 7,145 | 8,549 | 8,126 | 6,889 | 7,957 | 7,492 | 5,869 |
| 29 | 189 | 4 | 49,414 | | | | | 8,318 | 8,608 | 8,666 | 8,720 | 8,644 | 6,458 |
| 29 | 189 | 6 | 7,115 | | | | | 7,115 | | | | | |
| 29 | 189 | 14 | 44,481 | | | | | 5,645 | 8,626 | 8,116 | 7,597 | 8,012 | 6,485 |
| 29 | 189 | 3001 | 48,987 | | | | | 8,659 | 8,370 | 8,323 | 8,666 | 8,586 | 6,383 |
| 29 | 189 | 5001 | 2,056 | | | | | 2,056 | | | | | |
| 29 | 211 | 1 | 2,545 | 2,545 | | | | | | | | | |
| 29 | 510 | 86 | 1,167 | | | | | | | | | | 1,167 |
| 30 | 31 | 17 | 4,171 | | | | | | | | | | 4,171 |
| 30 | 63 | 36 | 16,053 | 4,005 | 8,191 | 3,857 | | | | | | | |
| 30 | 65 | 4 | 9,719 | | 4,728 | 4,991 | | | | | | | |
| 30 | 83 | 1 | 1,822 | | | | | | | | 371 | 1,109 | 342 |
| 30 | 111 | 86 | 13,955 | | | | | 3,138 | 5,732 | 5,085 | | | |
| 32 | 31 | 16 | 46,666 | | | | | 5,767 | 6,485 | 8,580 | 8,667 | 8,530 | 8,637 |
| 32 | 31 | 2002 | 9,551 | 8,244 | 1,307 | | | | | | | | |
| 33 | 11 | 1011 | 20,352 | | | | 3,883 | 2,781 | 2,942 | 3,394 | 3,709 | 1,799 | 1,844 |
| 33 | 11 | 5001 | 4,297 | | | | | | | 1,581 | 1,254 | 538 | 924 |
| 33 | 15 | 13 | 15,143 | 7,601 | 5,823 | 1,719 | | | | | | | |
| 33 | 15 | 14 | 20,815 | | | 4,658 | 7,894 | 7,627 | | | 636 | | |
| 34 | 3 | 6 | 25,531 | | | | | | | 531 | 8,294 | 8,116 | 8,590 |
| 34 | 7 | 3 | 62,322 | 8,285 | 8,631 | 6,026 | 8,140 | 7,813 | 8,513 | 8,529 | 6,385 | | |
| 34 | 21 | 5 | 85,539 | 8,516 | 8,690 | 8,664 | 8,609 | 8,577 | 8,447 | 8,245 | 8,622 | 8,545 | 8,624 |
| 34 | 23 | 11 | 85,816 | 8,452 | 8,550 | 8,581 | 8,633 | 8,618 | 8,497 | 8,615 | 8,649 | 8,691 | 8,530 |
| 35 | 1 | 23 | 67,865 | | 2,063 | 8,165 | 8,277 | 8,449 | 8,080 | 8,301 | 8,270 | 8,233 | 8,027 |
| 35 | 1 | 24 | 28,136 | | | | 1,421 | 7,765 | 7,849 | 7,588 | 3,513 | | |
| 35 | 1 | 5010 | 596 | | | | 596 | | | | | | |
| 35 | 13 | 21 | 41,707 | | | | | | 8,409 | 8,504 | 8,416 | 8,130 | 8,248 |
| 35 | 13 | 22 | 33,628 | | | | | | 6,779 | 7,158 | 6,691 | 6,464 | 6,536 |
| 35 | 15 | 1004 | 22,509 | | | | | | 4,290 | 8,386 | 8,129 | 1,704 | |
| 35 | 15 | 1005 | 30,353 | | | | | | 3,793 | 7,084 | 6,909 | 5,995 | 6,572 |
| 35 | 25 | 8 | 37,004 | | | | | | 4,390 | 8,323 | 8,219 | 8,089 | 7,983 |
| 35 | 29 | 3 | 33,398 | | | | | | 3,747 | 7,647 | 7,275 | 7,341 | 7,388 |
| 35 | 43 | 1003 | 25,623 | | | | | | 4,293 | 8,570 | 7,947 | 4,813 | |
| 35 | 45 | 9 | 35,720 | | | | | | 8,654 | 8,624 | 8,650 | 3,159 | 6,633 |
| 35 | 45 | 18 | 40,870 | | | | | | 8,321 | 8,581 | 8,351 | 7,646 | 7,971 |
| 35 | 45 | 1005 | 37,500 | | | | | | 7,840 | 8,393 | 6,523 | 6,945 | 7,799 |
| 36 | 5 | 83 | 49,512 | 6,830 | 8,271 | 8,597 | 8,675 | 7,755 | 8,504 | 880 | | | |
| 36 | 5 | 110 | 72,407 | 4,945 | 7,169 | 6,875 | 8,557 | 7,801 | 8,542 | 8,420 | 8,343 | 7,941 | 3,814 |
| 36 | 5 | 133 | 33,464 | | | | | | | 7,552 | 8,579 | 8,623 | 8,710 |
| 36 | 29 | 2 | 76,276 | 7,617 | 8,093 | 8,341 | 7,896 | 7,796 | 8,458 | 8,560 | 5,289 | 6,203 | 8,023 |

List of AQS NOx Monitoring Stations and Number of Valid Observations by Year

| State | County | | | | | | | | | | | | |
|-------|--------|---------|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Code | Code | Site ID | Total Obs | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 36 | 29 | 5 | 84,079 | 8,455 | 8,563 | 8,428 | 8,602 | 8,535 | 8,337 | 8,531 | 8,289 | 8,126 | 8,213 |
| 36 | 33 | 7003 | 8,499 | | | | | | | | | 3,488 | 5,011 |
| 36 | 59 | 5 | 79,927 | 8,284 | 8,487 | 7,483 | 6,458 | 8,281 | 8,405 | 7,773 | 8,466 | 8,583 | 7,707 |
| 36 | 61 | 10 | 3,414 | 3,414 | | | | | | | | | |
| 36 | 61 | 56 | 55,420 | 6,683 | 8,337 | 5,118 | 5,507 | 8,525 | 8,521 | 8,551 | 4,178 | | |
| 36 | 81 | 97 | 8,363 | 8,363 | | | | | | | | | |
| 36 | 81 | 98 | 40,109 | 7,897 | 8,167 | 7,674 | 7,758 | 8,509 | 104 | | | | |
| 36 | 81 | 124 | 76,324 | 3,710 | 7,715 | 7,894 | 7,759 | 7,778 | 8,240 | 8,089 | 8,583 | 8,189 | 8,367 |
| 37 | 67 | 22 | 72,061 | 8,530 | 8,358 | 8,477 | 8,409 | 6,501 | 6,207 | 7,263 | 6,960 | 3,704 | 7,652 |
| 37 | 119 | 41 | 83,052 | 7,804 | 8,220 | 8,637 | 8,528 | 8,585 | 7,038 | 8,524 | 8,586 | 8,539 | 8,591 |
| 38 | 13 | 4 | 21,713 | | | | | | | 5,192 | 5,247 | 5,480 | 5,794 |
| 38 | 15 | 3 | 33,373 | | | | | | | 8,580 | 8,427 | 8,441 | 7,925 |
| 38 | 17 | 1004 | 29,764 | | | | | | | 7,816 | 7,420 | 7,233 | 7,295 |
| 38 | 25 | 3 | 20,399 | | | | | | | 6,034 | 6,554 | 6,080 | 1,731 |
| 38 | 53 | 2 | 15,193 | | | | | | | 2,793 | 3,907 | 3,887 | 4,606 |
| 38 | 57 | 4 | 29,316 | | | | | | | 7,266 | 7,526 | 7,380 | 7,144 |
| 38 | 57 | 102 | 22,133 | | | | | | | 7,146 | 7,146 | 7,407 | 7,580 |
| 38 | 57 | 124 | 19,713 | | | | | | | | 6,862 | 6,206 | 6,645 |
| 38 | 65 | 2 | 28,333 | | | | | | | 6,829 | 7,080 | 7,030 | 7,394 |
| 39 | 9 | 4 | 31,284 | | | | | | | 7,964 | 7,891 | 7,508 | 7,921 |
| 39 | 35 | 60 | 11,528 | | | | | | | | | 3,486 | 8,042 |
| 39 | 103 | 4 | 6,275 | | | | | | | | | 2,643 | 3,632 |
| 40 | 1 | 9009 | 17,332 | | | | | | | | 3,495 | 7,081 | 6,756 |
| 40 | 21 | 9002 | 63,658 | 7,805 | 6,760 | 4,196 | 7,953 | 8,451 | 7,989 | 8,186 | 8,100 | 4,218 | |
| 40 | 27 | 49 | 12,515 | | 8,307 | 4,208 | | | | | | | |
| 40 | 71 | 9003 | 29,138 | 7,939 | 8,507 | 7,934 | 4,758 | | | | | | |
| 40 | 97 | 9014 | 28,915 | | | | 3,814 | 5,739 | 7,042 | 2,582 | 6,138 | 3,600 | |
| 40 | 101 | 167 | 11,784 | | 8,278 | 3,506 | | | | | | | |
| 40 | 109 | 33 | 75,342 | | 8,697 | 8,099 | 8,610 | 8,493 | 8,432 | 8,355 | 8,264 | 8,347 | 8,045 |
| 40 | 109 | 1037 | 74,096 | | 7,969 | 7,677 | 8,564 | 8,601 | 8,119 | 8,408 | 8,425 | 8,434 | 7,899 |
| 40 | 115 | 9004 | 34,517 | 8,394 | 7,975 | 5,944 | 7,954 | 4,250 | | | | | |
| 40 | 135 | 9015 | 21,749 | | | | | | 6,196 | 5,714 | 5,377 | 4,462 | |
| 40 | 143 | 174 | 10,752 | | 8,009 | 2,743 | | | | | | | |
| 40 | 143 | 1127 | 64,708 | | 8,317 | 8,330 | 4,201 | 8,320 | 8,533 | 7,794 | 3,576 | 7,676 | 7,961 |
| 41 | 51 | 80 | 48,213 | | 3,558 | 3,626 | 3,068 | 3,590 | 3,638 | 5,619 | 8,303 | 8,222 | 8,589 |
| 41 | 59 | 1003 | 8,585 | | | | | | | 7,045 | 1,540 | | |
| 42 | 1 | 1 | 42,398 | 3,043 | 3,637 | 2,951 | 4,897 | 4,817 | 5,099 | 5,063 | 3,444 | 4,805 | 4,642 |
| 42 | 3 | 8 | 83,673 | 8,553 | 8,173 | 8,552 | 8,618 | 8,603 | 8,549 | 8,092 | 8,562 | 8,698 | 7,273 |
| 42 | 3 | 10 | 83,085 | 8,323 | 8,733 | 8,481 | 8,647 | 8,497 | 8,620 | 8,088 | 7,804 | 8,103 | 7,789 |
| 42 | 3 | 31 | 4,140 | 4,140 | | | | | | | | | |
| 42 | 3 | 1005 | 79,360 | 3,672 | 8,442 | 8,633 | 8,416 | 8,372 | 7,666 | 8,092 | 8,709 | 8,714 | 8,644 |
| 42 | 7 | 14 | 84,148 | 8,577 | 8,641 | 8,708 | 8,678 | 8,683 | 8,415 | 8,654 | 8,202 | 8,346 | 7,244 |
| 42 | 11 | 9 | 46,152 | 8,670 | 8,664 | 8,645 | 8,617 | 8,675 | 2,881 | | | | |
| 42 | 11 | 11 | 29,270 | | | | | | | 4,381 | 8,430 | 8,135 | 8,324 |
| 42 | 13 | 801 | 68,895 | 8,616 | 8,563 | 8,478 | 8,680 | 8,655 | 8,622 | 8,699 | 8,582 | | |
| 42 | 17 | 12 | 82,953 | 8,633 | 8,441 | 8,265 | 8,526 | 8,537 | 8,300 | 8,503 | 7,234 | 8,006 | 8,508 |
| 42 | 21 | 11 | 85,965 | 8,507 | 8,332 | 8,607 | 8,671 | 8,677 | 8,675 | 8,712 | 8,583 | 8,697 | 8,504 |
| 42 | 27 | 100 | 72,403 | | 6,833 | 7,858 | 8,623 | 8,343 | 8,665 | 8,132 | 7,665 | 8,460 | 7,824 |
| 42 | 43 | 401 | 85,819 | 8,657 | 8,407 | 8,547 | 8,695 | 8,655 | 8,509 | 8,658 | 8,513 | 8,538 | 8,640 |
| 42 | 45 | 2 | 81,863 | 8,510 | 7,960 | 8,202 | 8,452 | 7,125 | 8,107 | 8,104 | 8,158 | 8,604 | 8,641 |
| 42 | 49 | 3 | 82,078 | 8,173 | 8,459 | 8,280 | 8,363 | 8,527 | 8,513 | 8,208 | 8,250 | 8,384 | 6,921 |
| 42 | 63 | 4 | 35,222 | | | | 1,390 | 8,299 | 8,294 | 8,582 | 8,657 | | |
| 42 | 69 | 2006 | 83,384 | 8,673 | 8,461 | 8,574 | 8,291 | 8,219 | 7,731 | 7,790 | 8,549 | 8,580 | 8,516 |
| 42 | 71 | 7 | 84,941 | 8,386 | 8,650 | 8,640 | 8,430 | 8,667 | 8,646 | 8,568 | 8,446 | 8,415 | 8,093 |
| 42 | 73 | 15 | 69,159 | 8,670 | 8,686 | 8,675 | 8,711 | 8,650 | 8,663 | 8,566 | 8,538 | | |
| 42 | 77 | 4 | 68,252 | 8,563 | 8,687 | 8,629 | 8,669 | 8,690 | 8,490 | 7,941 | 8,583 | | |
| 42 | 79 | 1101 | 66,399 | 8,538 | 8,371 | 8,583 | 8,436 | 7,811 | 8,113 | 7,939 | 8,608 | | |
| 42 | 91 | 13 | 65,657 | 8,305 | 7,857 | 7,805 | 8,460 | 8,522 | 8,177 | 7,914 | 8,617 | | |
| 42 | 95 | 25 | 84,845 | 8,676 | 8,589 | 8,558 | 8,441 | 8,694 | 7,927 | 8,378 | 8,674 | 8,492 | 8,416 |
| 42 | 99 | 301 | 76,664 | 8,048 | 7,290 | 7,616 | 8,496 | 7,277 | 7,579 | 7,996 | 7,809 | 7,512 | 7,041 |
| 42 | 125 | 5 | 86,265 | 8,591 | 8,619 | 8,542 | 8,681 | 8,593 | 8,703 | 8,452 | 8,756 | 8,624 | 8,704 |
| 42 | 125 | 200 | 68,561 | 8,734 | 8,680 | 8,654 | 8,668 | 8,677 | 8,184 | 8,736 | 8,228 | | |
| 42 | 125 | 5001 | 64,554 | 8,318 | 6,897 | 8,376 | 8,361 | 8,252 | 8,045 | 8,476 | 7,829 | | |
| 42 | 129 | 8 | 66,658 | 7,755 | 8,484 | 8,622 | 8,437 | 8,441 | 8,487 | 8,508 | 7,924 | | |
| 42 | 133 | 8 | 86,184 | 8,710 | 8,676 | 8,525 | 8,707 | 8,646 | 8,564 | 8,475 | 8,598 | 8,637 | 8,646 |
| 44 | 3 | 2 | 12,286 | 560 | 1,259 | 989 | 814 | 2,016 | 1,860 | 1,462 | 988 | 1,214 | 1,124 |
| 44 | 7 | 12 | 79,298 | 8,019 | 8,035 | 8,038 | 7,599 | 8,166 | 7,100 | 7,990 | 8,254 | 8,084 | 8,013 |
| 44 | 7 | 1010 | 19,094 | 1,684 | 1,932 | 1,528 | 1,867 | 1,978 | 2,072 | 1,999 | 2,032 | 1,973 | 2,029 |

| State | County | | | | | | | | | | | | |
|-------|--------|---------|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Code | Code | Site ID | Total Obs | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 46 | 33 | 132 | 11,822 | | | | | | | 5,411 | 1,070 | 4,473 | 868 |
| 46 | 71 | 1 | 13,490 | | | | | | | 2,918 | 5,178 | 3,121 | 2,273 |
| 46 | 99 | 7 | 6,071 | | | | | | | 6,071 | | | |
| 46 | 99 | 8 | 24,484 | | | | | | | | 8,079 | 7,810 | 8,595 |
| 46 | 109 | 4003 | 5,701 | 1,019 | 4,682 | | | | | | | | |
| 46 | 127 | 1 | 11,276 | | | | | | | | | 4,103 | 7,173 |
| 46 | 127 | 2 | 11,127 | | | | | | | | | 3,602 | 7,525 |
| 47 | 9 | 101 | 4,115 | | | | | | | | 966 | 1,946 | 1,203 |
| 47 | 31 | 4 | 8,075 | | | | 6,722 | 1,353 | | | | | |
| 47 | 63 | 3 | 16,549 | 6,636 | 1,552 | | | | 8,361 | | | | |
| 47 | 75 | 3 | 17,215 | | 5,105 | 2,749 | | | 6,753 | 2,608 | | | |
| 47 | 85 | 20 | 7,765 | | | 7,765 | | | | | | | |
| 47 | 93 | 1030 | 823 | 823 | | | | | | | | | |
| 47 | 121 | 104 | 16,270 | | 5,846 | 2,172 | | 6,914 | 1,338 | | | | |
| 47 | 125 | 1010 | 10,877 | 2,054 | | | | 6,804 | 2,019 | | | | |
| 47 | 131 | 4 | 8,055 | | | 6,679 | 1,376 | | | | | | |
| 47 | 141 | 4 | 8,348 | | | | | | | 6,973 | 1,375 | | |
| 47 | 157 | 2005 | 8,315 | | | | | 5,569 | 2,746 | | | | |
| 48 | 29 | 46 | 79,505 | 8,421 | 8,287 | 7,038 | 8,026 | 8,342 | 8,499 | 8,420 | 8,208 | 8,283 | 5,981 |
| 48 | 29 | 52 | 71,911 | 7,194 | 3,515 | 5,531 | 7,699 | 8,120 | 8,371 | 8,360 | 8,369 | 7,265 | 7,487 |
| 48 | 29 | 55 | 20,949 | | | | | | | | 8,124 | 8,555 | 4,270 |
| 48 | 29 | 59 | 73,166 | 5,323 | 7,140 | 8,343 | 6,903 | 6,791 | 8,513 | 8,322 | 7,975 | 6,723 | 7,133 |
| 48 | 29 | 622 | 20,750 | | | | | | | | 6,222 | 8,274 | 6,254 |
| 48 | 39 | 1004 | 70,917 | 2,777 | 7,437 | 8,364 | 8,403 | 7,761 | 7,446 | 7,528 | 7,584 | 6,706 | 6,911 |
| 48 | 39 | 1016 | 51,490 | | | 2,967 | 6,710 | 6,865 | 6,943 | 7,451 | 7,032 | 6,349 | 7,173 |
| 48 | 113 | 69 | 80,653 | 7,729 | 8,341 | 8,489 | 8,494 | 8,054 | 7,799 | 7,854 | 7,465 | 8,147 | 8,281 |
| 48 | 113 | 75 | 80,998 | 6,263 | 8,203 | 8,322 | 8,253 | 8,137 | 8,521 | 8,215 | 8,510 | 8,048 | 8,526 |
| 48 | 113 | 87 | 75,542 | 7,769 | 6,892 | 7,776 | 8,482 | 6,359 | 8,242 | 8,203 | 7,575 | 6,488 | 7,756 |
| 48 | 121 | 34 | 79,612 | 7,393 | 6,531 | 8,556 | 8,472 | 8,519 | 8,517 | 8,409 | 8,100 | 7,259 | 7,856 |
| 48 | 139 | 15 | 44,542 | 3,906 | 4,344 | 8,247 | 8,298 | 8,225 | 7,239 | 4,283 | | | |
| 48 | 139 | 16 | 62,197 | | | 6,630 | 6,884 | 7,632 | 8,508 | 8,393 | 8,550 | 8,024 | 7,576 |
| 48 | 139 | 17 | 12,056 | | | | 1,652 | 8,419 | 1,985 | | | | |
| 48 | 139 | 1044 | 25,716 | | | | | | | 2,593 | 8,509 | 7,812 | 6,802 |
| 48 | 141 | 37 | 83,667 | 8,203 | 8,565 | 8,595 | 8,610 | 8,562 | 8,258 | 8,402 | 8,333 | 8,453 | 7,686 |
| 48 | 141 | 44 | 77,153 | 6,848 | 8,398 | 7,677 | 7,266 | 7,658 | 8,366 | 8,219 | 6,810 | 8,506 | 7,405 |
| 48 | 141 | 55 | 76,873 | 6,936 | 7,857 | 7,215 | 8,510 | 6,481 | 7,628 | 7,843 | 8,118 | 8,070 | 8,215 |
| 48 | 141 | 5 | | | | | | | | | | | |

List of AQS NOx Monitoring Stations and Number of Valid Observations by Year

| State | County | | | | | | | | | | | | |
|-------|--------|---------|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| Code | Code | Site ID | Total Obs | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 48 | 361 | 1001 | 75,567 | 7,417 | 6,362 | 8,318 | 7,315 | 7,676 | 7,941 | 8,376 | 7,798 | 7,012 | 7,352 |
| 48 | 423 | 7 | 79,708 | 8,353 | 7,352 | 8,099 | 6,211 | 8,254 | 8,291 | 8,209 | 8,469 | 8,311 | 8,159 |
| 48 | 439 | 57 | 4,347 | 4,347 | | | | | | | | | |
| 48 | 439 | 1002 | 79,929 | 6,732 | 7,622 | 7,888 | 8,480 | 8,400 | 8,526 | 8,434 | 8,496 | 8,232 | 7,119 |
| 48 | 439 | 3009 | 83,230 | 8,060 | 7,598 | 8,494 | 8,449 | 8,497 | 8,482 | 8,444 | 8,510 | 8,294 | 8,402 |
| 48 | 439 | 3011 | 74,868 | | 7,790 | 8,559 | 8,491 | 8,008 | 8,367 | 8,510 | 8,485 | 8,333 | 8,325 |
| 48 | 453 | 20 | 75,674 | 6,223 | 6,920 | 8,199 | 8,047 | 8,365 | 8,150 | 7,546 | 7,944 | 6,781 | 7,499 |
| 48 | 453 | 613 | 14,593 | | | 4,967 | 823 | 4,497 | 4,306 | | | | |
| 49 | 5 | 4 | 62,787 | | 1,446 | 8,520 | 8,346 | 5,977 | 7,016 | 7,990 | 8,375 | 7,441 | 7,676 |
| 49 | 11 | 1 | 20,681 | 8,344 | 8,476 | 3,861 | | | | | | | |
| 49 | 11 | 4 | 62,444 | | | 3,417 | 8,521 | 8,396 | 8,338 | 8,376 | 8,370 | 8,479 | 8,547 |
| 49 | 35 | 3 | 83,525 | 8,256 | 8,535 | 8,493 | 8,550 | 8,463 | 7,372 | 8,511 | 8,593 | 8,346 | 8,406 |
| 49 | 35 | 3006 | 81,937 | 8,520 | 8,593 | 8,584 | 8,631 | 8,579 | 8,571 | 8,565 | 8,527 | 8,523 | 4,844 |
| 49 | 47 | 2002 | 10,565 | | | | | | | | | 3,186 | 7,379 |
| 49 | 47 | 2003 | 9,241 | | | | | | | | | 2,739 | 6,502 |
| 49 | 49 | 2 | 81,152 | 8,498 | 8,174 | 7,683 | 8,256 | 8,483 | 7,332 | 8,385 | 8,418 | 8,411 | 7,512 |
| 49 | 53 | 6 | 8,650 | | | | | | | | 1,876 | 3,218 | 3,556 |
| 49 | 57 | 2 | 77,474 | 3,770 | 8,495 | 8,210 | 8,603 | 8,581 | 6,949 | 8,609 | 8,320 | 8,152 | 7,785 |
| 50 | 7 | 13 | 9,819 | 2,772 | 7,047 | | | | | | | | |
| 50 | 7 | 14 | 62,389 | | | 7,714 | 8,274 | 7,788 | 8,156 | 8,028 | 7,692 | 7,354 | 7,383 |
| 50 | 21 | 2 | 76,380 | 7,837 | 7,646 | 6,019 | 7,185 | 8,215 | 7,762 | 8,091 | 8,045 | 7,766 | 7,814 |
| 51 | 13 | 20 | 33,922 | | | | | | | 8,442 | 8,362 | 8,604 | 8,514 |
| 51 | 33 | 1 | 23,734 | 7,738 | 6,191 | | 2,335 | | | 2,154 | 1,863 | 1,617 | 1,836 |
| 51 | 36 | 2 | 31,252 | | | | | | | 7,774 | 7,868 | 7,599 | 8,011 |
| 51 | 87 | 14 | 19,663 | | | | | | | | 3,015 | 8,086 | 8,562 |
| 51 | 107 | 1005 | 31,603 | | | | | | | 8,237 | 7,950 | 7,262 | 8,154 |
| 51 | 153 | 9 | 31,154 | | | | | | | 7,558 | 7,404 | 8,038 | 8,154 |
| 51 | 161 | 1004 | 32,568 | | | | | | | 7,854 | 7,967 | 8,488 | 8,259 |
| 51 | 165 | 3 | 33,327 | | | | | | | 8,173 | 8,480 | 8,324 | 8,350 |
| 51 | 510 | 9 | 31,846 | | | | | | | 6,823 | 8,424 | 8,524 | 8,075 |
| 51 | 700 | 13 | 2,658 | | | | | | | | | 542 | 2,116 |
| 51 | 710 | 24 | 23,020 | | | | | | | 8,067 | 8,564 | 6,389 | |
| 51 | 760 | 24 | 33,712 | | | | | | | 8,262 | 8,582 | 8,532 | 8,336 |
| 53 | 33 | 23 | 1,029 | | | 1,029 | | | | | | | |
| 53 | 33 | 32 | 13,444 | 8,101 | 5,343 | | | | | | | | |
| 53 | 33 | 80 | 38,652 | 8,179 | 7,519 | 7,305 | 8,004 | 6,444 | 1,201 | | | | |
| 55 | 71 | 7 | 13,289 | | | | 2,026 | 2,187 | 1,757 | 2,019 | 1,944 | 1,342 | 2,014 |
| 55 | 79 | 7 | 1,327 | 1,327 | | | | | | | | | |
| 55 | 79 | 26 | 64,133 | | | 5,109 | 8,648 | 8,499 | 8,115 | 8,529 | 8,169 | 8,461 | 8,603 |
| 55 | 79 | 41 | 8,490 | 8,097 | 393 | | | | | | | | |
| 55 | 89 | 9 | 4,455 | | | | 2,180 | 2,275 | | | | | |
| 55 | 111 | 7 | 9,112 | | | 6,299 | 2,813 | | | | | | |
| 56 | 5 | 123 | 58,130 | 5,416 | 6,844 | 6,015 | 7,116 | 6,821 | 6,504 | 6,533 | 3,379 | 3,706 | 5,796 |
| 56 | 5 | 456 | 49,432 | | | 2,778 | 7,747 | 7,387 | 5,618 | 6,692 | 6,480 | 5,504 | 7,226 |
| 56 | 5 | 789 | 1,227 | | | | 1,227 | | | | | | |
| 56 | 5 | 892 | 23,750 | | | 5,782 | 2,209 | 3,608 | 4,851 | | | 2,061 | 5,239 |
| 56 | 7 | 99 | 5,526 | | | | | | | 836 | 3,618 | 1,072 | |
| 56 | 9 | 819 | 12,853 | | | 6,471 | 1,188 | 1,898 | 1,642 | | | 558 | 1,096 |
| 56 | 13 | 99 | 7,908 | | | | | | | 2,737 | 2,280 | 1,282 | 1,609 |
| 56 | 35 | 98 | 24,869 | | | | 1,300 | 6,871 | 7,193 | 7,195 | 2,310 | | |
| 56 | 35 | 99 | 23,147 | | | | | 5,535 | 3,712 | 2,669 | 2,648 | 4,775 | 3,808 |
| 56 | 35 | 100 | 5,652 | | | | | 245 | 1,326 | 1,269 | 362 | 1,534 | 916 |
| 56 | 35 | 101 | 8,776 | | | | | | | | | 2,437 | 6,339 |
| 56 | 37 | 200 | 36,121 | | | | | | 6,759 | 7,783 | 7,596 | 6,942 | 7,041 |
| 56 | 41 | 101 | 19,013 | | | | | | | 3,846 | 4,642 | 6,091 | 4,434 |

APPENDIX B

Bin Data and Ambient Ratios for Data Sets

ARM2 Poly Curve Fitting Analysis

ALL Sites

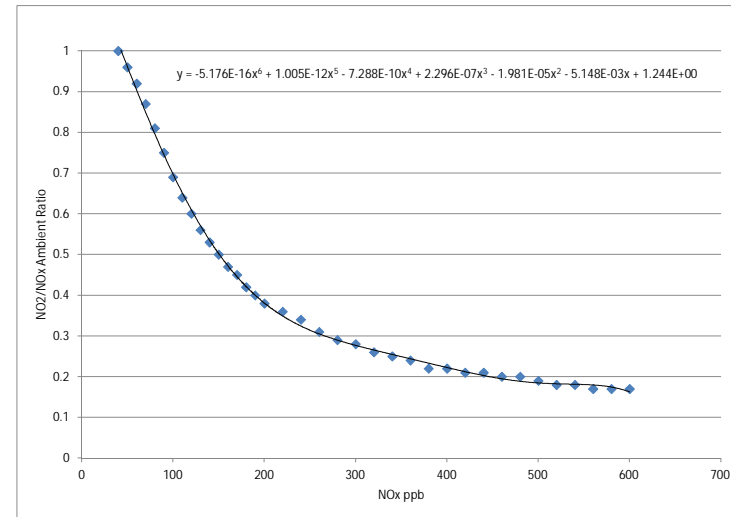
Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 6804348

| Upper Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb | Delta Change ppb |
|----------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|---------------------|
| 30 | 2330786 | 1 | 1.077 | 0.900 | 27 | |
| 40 | 1310767 | 1 | 1.019 | 0.900 | 36 | 9.0 |
| 50 | 817380 | 0.96 | 0.962 | 0.900 | 45 | 9.0 |
| 60 | 546757 | 0.92 | 0.905 | 0.900 | 54 | 9.0 |
| 70 | 385505 | 0.87 | 0.850 | 0.850 | 59 | 5.5 |
| 80 | 282446 | 0.81 | 0.796 | 0.796 | 64 | 4.2 |
| 90 | 214245 | 0.75 | 0.746 | 0.746 | 67 | 3.4 |
| 100 | 164999 | 0.69 | 0.697 | 0.697 | 70 | 2.6 |
| 110 | 130316 | 0.64 | 0.652 | 0.652 | 72 | 2.0 |
| 120 | 104093 | 0.6 | 0.610 | 0.610 | 73 | 1.5 |
| 130 | 84056 | 0.56 | 0.571 | 0.571 | 74 | 1.0 |
| 140 | 68350 | 0.53 | 0.535 | 0.535 | 75 | 0.7 |
| 150 | 56801 | 0.5 | 0.502 | 0.502 | 75 | 0.4 |
| 160 | 46956 | 0.47 | 0.473 | 0.473 | 76 | 0.3 |
| 170 | 39220 | 0.45 | 0.446 | 0.446 | 76 | 0.2 |
| 180 | 32769 | 0.42 | 0.422 | 0.422 | 76 | 0.1 |
| 190 | 27696 | 0.4 | 0.400 | 0.400 | 76 | 0.1 |
| 200 | 23436 | 0.38 | 0.381 | 0.381 | 76 | 0.2 |
| 220 | 36618 | 0.36 | 0.349 | 0.349 | 77 | 0.6 |
| 240 | 26373 | 0.34 | 0.325 | 0.325 | 78 | 1.1 |
| 260 | 19349 | 0.31 | 0.305 | 0.305 | 79 | 1.5 |
| 280 | 14204 | 0.29 | 0.290 | 0.290 | 81 | 1.8 |
| 300 | 10468 | 0.28 | 0.277 | 0.277 | 83 | 1.9 |
| 320 | 7816 | 0.26 | 0.266 | 0.266 | 85 | 1.9 |
| 340 | 5774 | 0.25 | 0.255 | 0.255 | 87 | 1.6 |
| 360 | 4293 | 0.24 | 0.244 | 0.244 | 88 | 1.2 |
| 380 | 3287 | 0.22 | 0.233 | 0.233 | 89 | 0.8 |
| 400 | 2524 | 0.22 | 0.223 | 0.223 | 89 | 0.4 |
| 420 | 1880 | 0.21 | 0.212 | 0.212 | 89 | 0.1 |
| 440 | 1401 | 0.21 | 0.203 | 0.203 | 89 | 0.1 |
| 460 | 1079 | 0.2 | 0.195 | 0.200 | 92 | 2.7 |
| 480 | 810 | 0.2 | 0.189 | 0.200 | 96 | 4.0 |
| 500 | 658 | 0.190 | 0.185 | 0.200 | 100 | 4.0 |
| 520 | 413 | 0.18 | 0.183 | 0.200 | 104 | 4.0 |
| 540 | 310 | 0.18 | 0.182 | 0.200 | 108 | 4.0 |
| 560 | 239 | 0.17 | 0.180 | 0.200 | 112 | 4.0 |
| 580 | 159 | 0.17 | 0.175 | 0.200 | 116 | 4.0 |
| 600 | 115 | 0.17 | 0.162 | 0.200 | 120 | 4.0 |

6804348

ARM2 Poly Curve Fitting Analysis - ALL Sites



Linest and Logest Functions

Cubic =LINEST(X_1^{1,2,3,4,5,6})

a

-5.176E-16

b

1.005E-12

c

-7.288E-10

d

2.296E-07

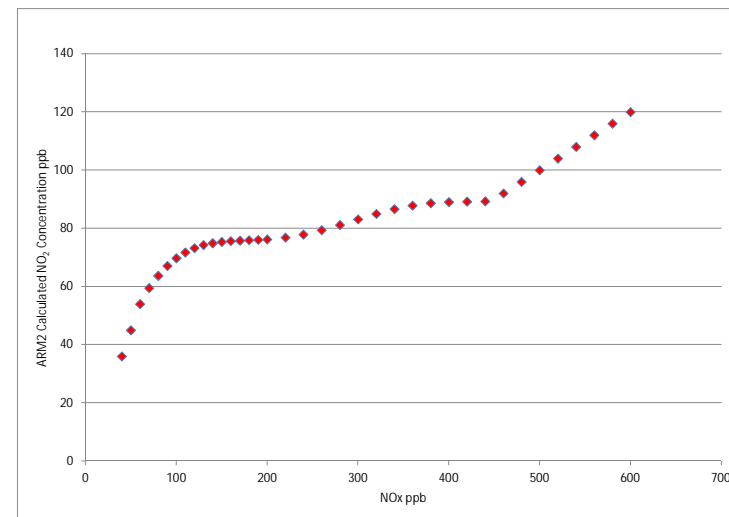
-1.98144E-05

-0.00514793

1.24409

equation

$=(((a7^3 \cdot -5.17597743179288E-16) + (a7^2 \cdot 1.00504244687937E-12) + (a7 \cdot -7.2884004876291E-10) + 2.29606193106293E-07)$



ARM2 Poly Curve Fitting Analysis

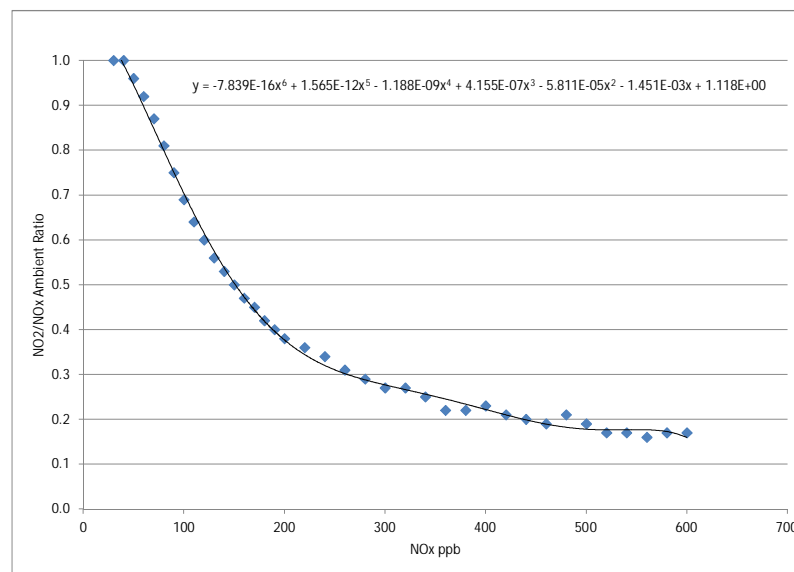
Rural & Suburban

Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 3600763

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 1295731 | 1.000 | 1.080 | 0.900 | 27 |
| 40 | 703035 | 1.000 | 1.020 | 0.900 | 36 |
| 50 | 426294 | 0.960 | 0.962 | 0.900 | 45 |
| 60 | 280390 | 0.920 | 0.904 | 0.900 | 54 |
| 70 | 194496 | 0.870 | 0.849 | 0.849 | 59 |
| 80 | 142124 | 0.810 | 0.796 | 0.796 | 64 |
| 90 | 107859 | 0.750 | 0.745 | 0.745 | 67 |
| 100 | 82582 | 0.690 | 0.697 | 0.697 | 70 |
| 110 | 64674 | 0.640 | 0.652 | 0.652 | 72 |
| 120 | 52123 | 0.600 | 0.610 | 0.610 | 73 |
| 130 | 41692 | 0.560 | 0.571 | 0.571 | 74 |
| 140 | 34307 | 0.530 | 0.536 | 0.536 | 75 |
| 150 | 28242 | 0.500 | 0.503 | 0.503 | 75 |
| 160 | 23148 | 0.470 | 0.473 | 0.473 | 76 |
| 170 | 19288 | 0.450 | 0.446 | 0.446 | 76 |
| 180 | 15995 | 0.420 | 0.422 | 0.422 | 76 |
| 190 | 13403 | 0.400 | 0.401 | 0.401 | 76 |
| 200 | 11235 | 0.380 | 0.381 | 0.381 | 76 |
| 220 | 17754 | 0.360 | 0.349 | 0.349 | 77 |
| 240 | 12477 | 0.340 | 0.324 | 0.324 | 78 |
| 260 | 9007 | 0.310 | 0.304 | 0.304 | 79 |
| 280 | 6700 | 0.290 | 0.288 | 0.288 | 81 |
| 300 | 4853 | 0.270 | 0.275 | 0.275 | 82 |
| 320 | 3571 | 0.270 | 0.263 | 0.263 | 84 |
| 340 | 2538 | 0.250 | 0.253 | 0.253 | 86 |
| 360 | 1883 | 0.220 | 0.242 | 0.242 | 87 |
| 380 | 1430 | 0.220 | 0.232 | 0.232 | 88 |
| 400 | 1029 | 0.230 | 0.221 | 0.221 | 89 |
| 420 | 769 | 0.210 | 0.211 | 0.211 | 89 |
| 440 | 578 | 0.200 | 0.201 | 0.201 | 89 |
| 460 | 447 | 0.190 | 0.193 | 0.200 | 92 |
| 480 | 336 | 0.210 | 0.186 | 0.200 | 96 |
| 500 | 283 | 0.190 | 0.180 | 0.200 | 100 |
| 520 | 173 | 0.170 | 0.177 | 0.200 | 104 |
| 540 | 122 | 0.170 | 0.175 | 0.200 | 108 |
| 560 | 91 | 0.160 | 0.173 | 0.200 | 112 |
| 580 | 57 | 0.170 | 0.170 | 0.200 | 116 |
| 600 | 47 | 0.170 | 0.163 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Rural & Suburban



Linest and Logest Functions

Cubic =LINEST(X_1^(1,2,3,4,5,6))

a -4.214E-16 b 8.317E-13 c -6.091E-10 d 1.899E-07
-1.33492E-05 -0.005623261 1.256021

ARM2 Poly Curve Fitting Analysis

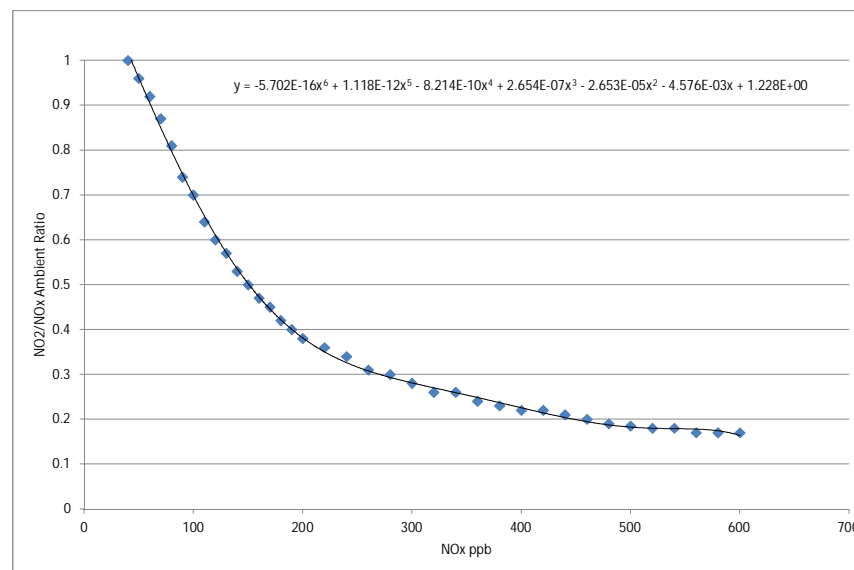
Urban & City Center

Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 3177338

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 1020825 | 1 | 1.073 | 0.900 | 27 |
| 40 | 601939 | 1 | 1.017 | 0.900 | 36 |
| 50 | 388295 | 0.96 | 0.961 | 0.900 | 45 |
| 60 | 264973 | 0.92 | 0.905 | 0.900 | 54 |
| 70 | 190229 | 0.87 | 0.851 | 0.851 | 60 |
| 80 | 139879 | 0.81 | 0.798 | 0.798 | 64 |
| 90 | 106102 | 0.74 | 0.747 | 0.747 | 67 |
| 100 | 82251 | 0.7 | 0.699 | 0.699 | 70 |
| 110 | 65527 | 0.64 | 0.653 | 0.653 | 72 |
| 120 | 51895 | 0.6 | 0.611 | 0.611 | 73 |
| 130 | 42318 | 0.57 | 0.572 | 0.572 | 74 |
| 140 | 34009 | 0.53 | 0.536 | 0.536 | 75 |
| 150 | 28542 | 0.5 | 0.503 | 0.503 | 75 |
| 160 | 23785 | 0.47 | 0.473 | 0.473 | 76 |
| 170 | 19921 | 0.45 | 0.446 | 0.446 | 76 |
| 180 | 16761 | 0.42 | 0.422 | 0.422 | 76 |
| 190 | 14290 | 0.4 | 0.401 | 0.401 | 76 |
| 200 | 12192 | 0.38 | 0.382 | 0.382 | 76 |
| 220 | 18854 | 0.36 | 0.350 | 0.350 | 77 |
| 240 | 13893 | 0.34 | 0.326 | 0.326 | 78 |
| 260 | 10338 | 0.31 | 0.308 | 0.308 | 80 |
| 280 | 7503 | 0.3 | 0.293 | 0.293 | 82 |
| 300 | 5615 | 0.28 | 0.281 | 0.281 | 84 |
| 320 | 4245 | 0.26 | 0.270 | 0.270 | 86 |
| 340 | 3236 | 0.26 | 0.259 | 0.259 | 88 |
| 360 | 2409 | 0.24 | 0.249 | 0.249 | 89 |
| 380 | 1857 | 0.23 | 0.237 | 0.237 | 90 |
| 400 | 1494 | 0.22 | 0.226 | 0.226 | 90 |
| 420 | 1111 | 0.22 | 0.214 | 0.214 | 90 |
| 440 | 823 | 0.21 | 0.204 | 0.204 | 90 |
| 460 | 632 | 0.2 | 0.195 | 0.200 | 92 |
| 480 | 474 | 0.19 | 0.188 | 0.200 | 96 |
| 500 | 375 | 0.185 | 0.183 | 0.200 | 100 |
| 520 | 240 | 0.18 | 0.180 | 0.200 | 104 |
| 540 | 188 | 0.18 | 0.179 | 0.200 | 108 |
| 560 | 148 | 0.17 | 0.178 | 0.200 | 112 |
| 580 | 102 | 0.17 | 0.174 | 0.200 | 116 |
| 600 | 68 | 0.17 | 0.164 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Urban & City Center



Linest and Logest Functions

Cubic =LINEST(X_1^(1,2,3,4,5,6))

a

-5.702E-16

b

1.118E-12

c

-8.214E-10

d

2.654E-07

-2.65265E-05

-0.004576385

1.227829

ARM2 Poly Curve Fitting Analysis

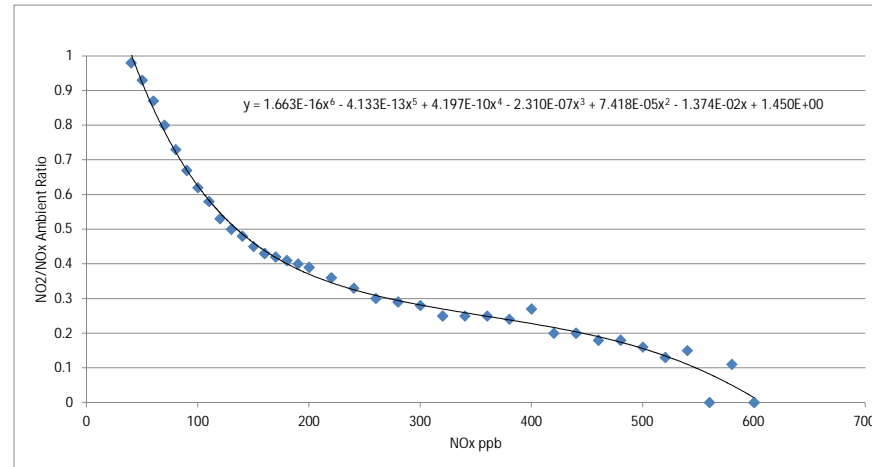
Mountain Region

Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 304578

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 84000 | 1 | 1.099 | 0.900 | 27 |
| 40 | 51663 | 0.98 | 1.006 | 0.900 | 36 |
| 50 | 35255 | 0.93 | 0.922 | 0.900 | 45 |
| 60 | 25925 | 0.87 | 0.848 | 0.848 | 51 |
| 70 | 19801 | 0.8 | 0.782 | 0.782 | 55 |
| 80 | 15568 | 0.73 | 0.723 | 0.723 | 58 |
| 90 | 12591 | 0.67 | 0.671 | 0.671 | 60 |
| 100 | 10148 | 0.62 | 0.625 | 0.625 | 62 |
| 110 | 8258 | 0.58 | 0.584 | 0.584 | 64 |
| 120 | 6866 | 0.53 | 0.547 | 0.547 | 66 |
| 130 | 5671 | 0.5 | 0.515 | 0.515 | 67 |
| 140 | 4531 | 0.48 | 0.486 | 0.486 | 68 |
| 150 | 3961 | 0.45 | 0.461 | 0.461 | 69 |
| 160 | 3313 | 0.43 | 0.439 | 0.439 | 70 |
| 170 | 2781 | 0.42 | 0.419 | 0.419 | 71 |
| 180 | 2278 | 0.41 | 0.401 | 0.401 | 72 |
| 190 | 1953 | 0.4 | 0.385 | 0.385 | 73 |
| 200 | 1630 | 0.39 | 0.370 | 0.370 | 74 |
| 220 | 2584 | 0.36 | 0.346 | 0.346 | 76 |
| 240 | 1784 | 0.33 | 0.326 | 0.326 | 78 |
| 260 | 1270 | 0.3 | 0.309 | 0.309 | 80 |
| 280 | 849 | 0.29 | 0.295 | 0.295 | 82 |
| 300 | 631 | 0.28 | 0.282 | 0.282 | 85 |
| 320 | 422 | 0.25 | 0.270 | 0.270 | 86 |
| 340 | 308 | 0.25 | 0.259 | 0.259 | 88 |
| 360 | 199 | 0.25 | 0.249 | 0.249 | 90 |
| 380 | 125 | 0.24 | 0.239 | 0.239 | 91 |
| 400 | 74 | 0.27 | 0.228 | 0.228 | 91 |
| 420 | 63 | 0.2 | 0.217 | 0.217 | 91 |
| 440 | 24 | 0.2 | 0.204 | 0.204 | 90 |
| 460 | 25 | 0.18 | 0.190 | 0.200 | 92 |
| 480 | 12 | 0.18 | 0.174 | 0.200 | 96 |
| 500 | 5 | 0.16 | 0.156 | 0.200 | 100 |
| 520 | 2 | 0.13 | 0.135 | 0.200 | 104 |
| 540 | 4 | 0.15 | 0.110 | 0.200 | 108 |
| 560 | 1 | 0 | 0.082 | 0.200 | 112 |
| 580 | 2 | 0.11 | 0.050 | 0.200 | 116 |
| 600 | 1 | 0 | 0.014 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Mountain Region



Linest and Logest Functions

Cubic =LINEST(X_1^{1,2,3,4,5,6})

a

1.663E-16

b

-4.133E-13

c

4.197E-10

d

-2.310E-07

7.41799E-05

-0.013744247

1.450436

ARM2 Poly Curve Fitting Analysis

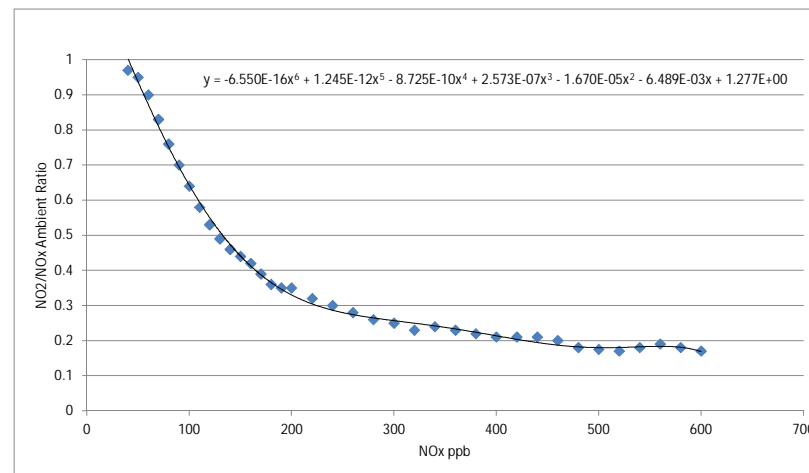
Northeast Region

Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 1886112

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 643420 | 1 | 1.073 | 0.900 | 27 |
| 40 | 375722 | 0.97 | 1.005 | 0.900 | 36 |
| 50 | 236081 | 0.95 | 0.938 | 0.900 | 45 |
| 60 | 157607 | 0.9 | 0.873 | 0.873 | 52 |
| 70 | 110023 | 0.83 | 0.810 | 0.810 | 57 |
| 80 | 79065 | 0.76 | 0.751 | 0.751 | 60 |
| 90 | 58733 | 0.7 | 0.695 | 0.695 | 63 |
| 100 | 44084 | 0.64 | 0.643 | 0.643 | 64 |
| 110 | 34197 | 0.58 | 0.595 | 0.595 | 65 |
| 120 | 26539 | 0.53 | 0.550 | 0.550 | 66 |
| 130 | 21049 | 0.49 | 0.510 | 0.510 | 66 |
| 140 | 16665 | 0.46 | 0.474 | 0.474 | 66 |
| 150 | 13535 | 0.44 | 0.442 | 0.442 | 66 |
| 160 | 10935 | 0.42 | 0.413 | 0.413 | 66 |
| 170 | 9027 | 0.39 | 0.388 | 0.388 | 66 |
| 180 | 7477 | 0.36 | 0.366 | 0.366 | 66 |
| 190 | 6202 | 0.35 | 0.347 | 0.347 | 66 |
| 200 | 5107 | 0.35 | 0.330 | 0.330 | 66 |
| 220 | 8093 | 0.32 | 0.305 | 0.305 | 67 |
| 240 | 5695 | 0.3 | 0.287 | 0.287 | 69 |
| 260 | 4234 | 0.28 | 0.274 | 0.274 | 71 |
| 280 | 3116 | 0.26 | 0.265 | 0.265 | 74 |
| 300 | 2312 | 0.25 | 0.257 | 0.257 | 77 |
| 320 | 1772 | 0.23 | 0.250 | 0.250 | 80 |
| 340 | 1322 | 0.24 | 0.242 | 0.242 | 82 |
| 360 | 960 | 0.23 | 0.233 | 0.233 | 84 |
| 380 | 772 | 0.22 | 0.224 | 0.224 | 85 |
| 400 | 606 | 0.21 | 0.214 | 0.214 | 86 |
| 420 | 444 | 0.21 | 0.204 | 0.204 | 86 |
| 440 | 332 | 0.21 | 0.195 | 0.200 | 88 |
| 460 | 282 | 0.2 | 0.187 | 0.200 | 92 |
| 480 | 187 | 0.18 | 0.182 | 0.200 | 96 |
| 500 | 173 | 0.175 | 0.180 | 0.200 | 100 |
| 520 | 111 | 0.17 | 0.181 | 0.200 | 104 |
| 540 | 82 | 0.18 | 0.183 | 0.200 | 108 |
| 560 | 74 | 0.19 | 0.184 | 0.200 | 112 |
| 580 | 41 | 0.18 | 0.181 | 0.200 | 116 |
| 600 | 36 | 0.17 | 0.169 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Northeast Region



Linest and Logest Functions
Cubic =LINEST(X_1^{1,2,3,4,5,6})

| | | | | | | | |
|------------|-----------|------------|-----------|--------------|--------------|----------|--|
| a | b | c | d | | | | |
| -6.550E-16 | 1.245E-12 | -8.725E-10 | 2.573E-07 | -1.66953E-05 | -0.006489282 | 1.276762 | |

ARM2 Poly Curve Fitting Analysis

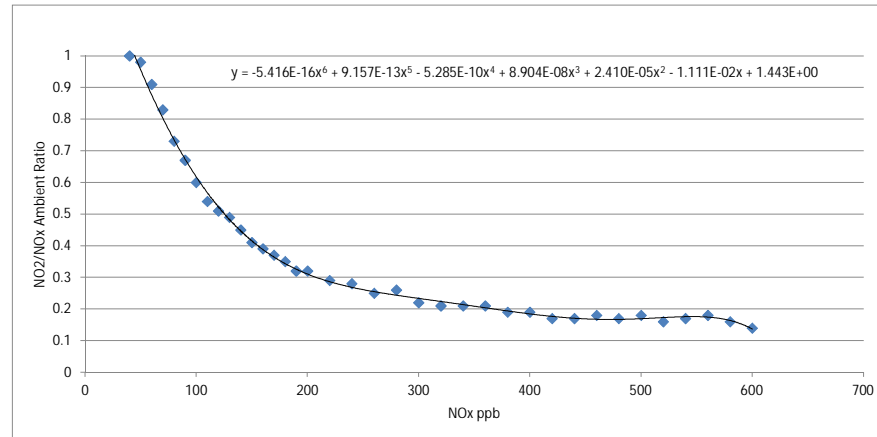
Southeast Region

Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 419681

| Mid Pt. | Number in | Observed | Step 1 Fitted | Final Adjusted | ARM2 Calculated |
|-------------|-----------|--------------|---------------|----------------|-----------------|
| Nox BIN ppb | Bin | 98% Perc ARM | ARM Ratio | ARM Ratio | NO2 ppb |
| 30 | 179401 | 1.05 | 1.133 | 0.900 | 27 |
| 40 | 82728 | 1 | 1.042 | 0.900 | 36 |
| 50 | 45511 | 0.98 | 0.956 | 0.900 | 45 |
| 60 | 27839 | 0.91 | 0.876 | 0.876 | 53 |
| 70 | 18693 | 0.83 | 0.803 | 0.803 | 56 |
| 80 | 13197 | 0.73 | 0.735 | 0.735 | 59 |
| 90 | 9866 | 0.67 | 0.674 | 0.674 | 61 |
| 100 | 7427 | 0.6 | 0.618 | 0.618 | 62 |
| 110 | 5635 | 0.54 | 0.568 | 0.568 | 62 |
| 120 | 4433 | 0.51 | 0.523 | 0.523 | 63 |
| 130 | 3688 | 0.49 | 0.482 | 0.482 | 63 |
| 140 | 2971 | 0.45 | 0.447 | 0.447 | 63 |
| 150 | 2566 | 0.41 | 0.416 | 0.416 | 62 |
| 160 | 2110 | 0.39 | 0.388 | 0.388 | 62 |
| 170 | 1762 | 0.37 | 0.364 | 0.364 | 62 |
| 180 | 1475 | 0.35 | 0.344 | 0.344 | 62 |
| 190 | 1350 | 0.32 | 0.326 | 0.326 | 62 |
| 200 | 1156 | 0.32 | 0.311 | 0.311 | 62 |
| 220 | 1775 | 0.29 | 0.287 | 0.287 | 63 |
| 240 | 1331 | 0.28 | 0.269 | 0.269 | 64 |
| 260 | 1026 | 0.25 | 0.255 | 0.255 | 66 |
| 280 | 813 | 0.26 | 0.244 | 0.244 | 68 |
| 300 | 583 | 0.22 | 0.234 | 0.234 | 70 |
| 320 | 512 | 0.21 | 0.224 | 0.224 | 72 |
| 340 | 356 | 0.21 | 0.214 | 0.214 | 73 |
| 360 | 318 | 0.21 | 0.204 | 0.204 | 73 |
| 380 | 275 | 0.19 | 0.194 | 0.200 | 76 |
| 400 | 228 | 0.19 | 0.185 | 0.200 | 80 |
| 420 | 156 | 0.17 | 0.177 | 0.200 | 84 |
| 440 | 130 | 0.17 | 0.171 | 0.200 | 88 |
| 460 | 99 | 0.18 | 0.168 | 0.200 | 92 |
| 480 | 80 | 0.17 | 0.168 | 0.200 | 96 |
| 500 | 61 | 0.18 | 0.170 | 0.200 | 100 |
| 520 | 44 | 0.16 | 0.173 | 0.200 | 104 |
| 540 | 32 | 0.17 | 0.176 | 0.200 | 108 |
| 560 | 27 | 0.18 | 0.175 | 0.200 | 112 |
| 580 | 14 | 0.16 | 0.164 | 0.200 | 116 |
| 600 | 13 | 0.14 | 0.137 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Southeast Region



Linest and Logest Functions

Cubic =LINEST(X_1^{1,2,3,4,5,6})

a

-5.416E-16

b

9.157E-13

c

-5.285E-10

d

8.904E-08

2.41047E-05

-0.011108125

1.443045

ARM2 Poly Curve Fitting Analysis

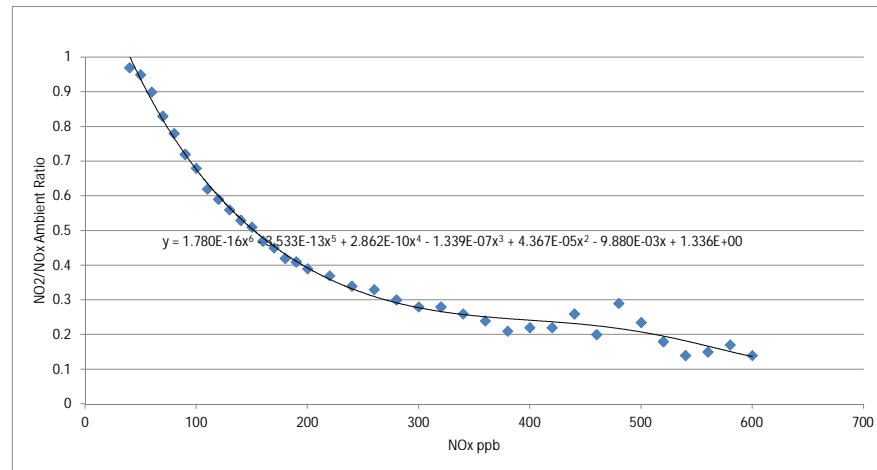
Midwest Region

Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 778336

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 283397 | 1 | 1.076 | 0.900 | 27 |
| 40 | 156803 | 0.97 | 1.003 | 0.900 | 36 |
| 50 | 95480 | 0.95 | 0.937 | 0.900 | 45 |
| 60 | 62403 | 0.9 | 0.875 | 0.875 | 53 |
| 70 | 43440 | 0.83 | 0.819 | 0.819 | 57 |
| 80 | 30934 | 0.78 | 0.768 | 0.768 | 61 |
| 90 | 22693 | 0.72 | 0.720 | 0.720 | 65 |
| 100 | 16907 | 0.68 | 0.677 | 0.677 | 68 |
| 110 | 12983 | 0.62 | 0.636 | 0.636 | 70 |
| 120 | 10210 | 0.59 | 0.599 | 0.599 | 72 |
| 130 | 7780 | 0.56 | 0.565 | 0.565 | 74 |
| 140 | 6191 | 0.53 | 0.534 | 0.534 | 75 |
| 150 | 4955 | 0.51 | 0.505 | 0.505 | 76 |
| 160 | 3948 | 0.47 | 0.479 | 0.479 | 77 |
| 170 | 3332 | 0.45 | 0.454 | 0.454 | 77 |
| 180 | 2692 | 0.42 | 0.432 | 0.432 | 78 |
| 190 | 2158 | 0.41 | 0.411 | 0.411 | 78 |
| 200 | 1816 | 0.39 | 0.393 | 0.393 | 79 |
| 220 | 2811 | 0.37 | 0.360 | 0.360 | 79 |
| 240 | 2010 | 0.34 | 0.332 | 0.332 | 80 |
| 260 | 1387 | 0.33 | 0.310 | 0.310 | 81 |
| 280 | 952 | 0.3 | 0.292 | 0.292 | 82 |
| 300 | 780 | 0.28 | 0.278 | 0.278 | 83 |
| 320 | 536 | 0.28 | 0.266 | 0.266 | 85 |
| 340 | 387 | 0.26 | 0.258 | 0.258 | 88 |
| 360 | 310 | 0.24 | 0.251 | 0.251 | 91 |
| 380 | 224 | 0.21 | 0.246 | 0.246 | 94 |
| 400 | 200 | 0.22 | 0.242 | 0.242 | 97 |
| 420 | 148 | 0.22 | 0.237 | 0.237 | 100 |
| 440 | 111 | 0.26 | 0.232 | 0.232 | 102 |
| 460 | 100 | 0.2 | 0.226 | 0.226 | 104 |
| 480 | 91 | 0.29 | 0.218 | 0.218 | 104 |
| 500 | 99 | 0.24 | 0.208 | 0.208 | 104 |
| 520 | 25 | 0.18 | 0.196 | 0.200 | 104 |
| 540 | 10 | 0.14 | 0.182 | 0.200 | 108 |
| 560 | 16 | 0.15 | 0.167 | 0.200 | 112 |
| 580 | 10 | 0.17 | 0.151 | 0.200 | 116 |
| 600 | 7 | 0.14 | 0.137 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Midwest Region



Linest and Logest Functions

Cubic =LINEST(X_1^{1,2,3,4,5,6})

a

1.780E-16

b

-3.533E-13

c

2.862E-10

d

-1.339E-07

4.36678E-05

-0.00987988

1.336483

ARM2 Poly Curve Fitting Analysis

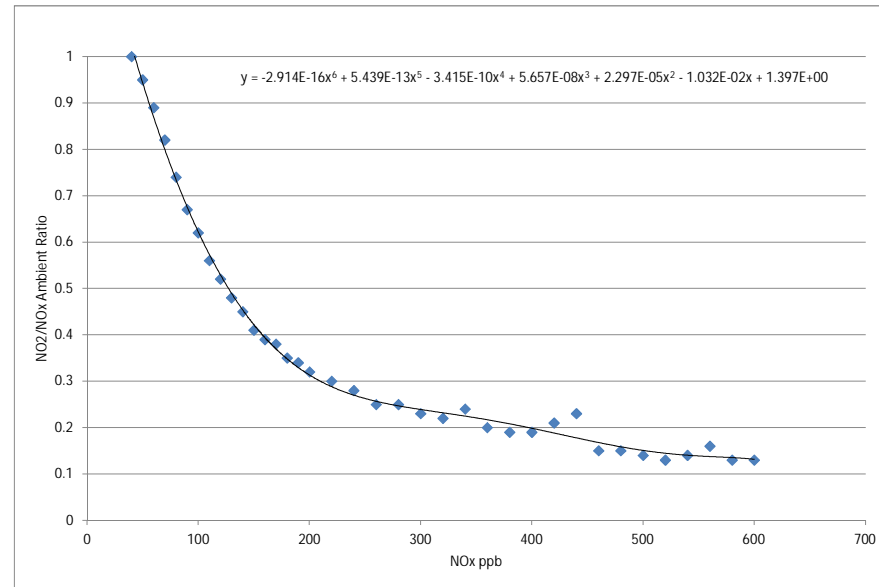
Southwest Region

Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 819626

| Mid Pt. | Number in | Observed | Step 1 Fitted | Final Adjusted | ARM2 Calculated |
|-------------|-----------|--------------|---------------|----------------|-----------------|
| Nox BIN ppb | Bin | 98% Perc ARM | ARM Ratio | ARM Ratio | NO2 ppb |
| 30 | 324695 | 1.01 | 1.109 | 0.900 | 27 |
| 40 | 161405 | 1 | 1.024 | 0.900 | 36 |
| 50 | 92570 | 0.95 | 0.944 | 0.900 | 45 |
| 60 | 58213 | 0.89 | 0.869 | 0.869 | 52 |
| 70 | 39442 | 0.82 | 0.799 | 0.799 | 56 |
| 80 | 28437 | 0.74 | 0.735 | 0.735 | 59 |
| 90 | 21170 | 0.67 | 0.676 | 0.676 | 61 |
| 100 | 16697 | 0.62 | 0.622 | 0.622 | 62 |
| 110 | 12993 | 0.56 | 0.573 | 0.573 | 63 |
| 120 | 10358 | 0.52 | 0.529 | 0.529 | 63 |
| 130 | 8458 | 0.48 | 0.489 | 0.489 | 64 |
| 140 | 6830 | 0.45 | 0.454 | 0.454 | 64 |
| 150 | 5833 | 0.41 | 0.422 | 0.422 | 63 |
| 160 | 4844 | 0.39 | 0.394 | 0.394 | 63 |
| 170 | 4107 | 0.38 | 0.370 | 0.370 | 63 |
| 180 | 3519 | 0.35 | 0.348 | 0.348 | 63 |
| 190 | 2952 | 0.34 | 0.330 | 0.330 | 63 |
| 200 | 2524 | 0.32 | 0.314 | 0.314 | 63 |
| 220 | 3780 | 0.3 | 0.288 | 0.288 | 63 |
| 240 | 2748 | 0.28 | 0.270 | 0.270 | 65 |
| 260 | 2048 | 0.25 | 0.257 | 0.257 | 67 |
| 280 | 1487 | 0.25 | 0.247 | 0.247 | 69 |
| 300 | 1124 | 0.23 | 0.239 | 0.239 | 72 |
| 320 | 874 | 0.22 | 0.232 | 0.232 | 74 |
| 340 | 653 | 0.24 | 0.225 | 0.225 | 76 |
| 360 | 494 | 0.2 | 0.217 | 0.217 | 78 |
| 380 | 374 | 0.19 | 0.208 | 0.208 | 79 |
| 400 | 271 | 0.19 | 0.199 | 0.200 | 80 |
| 420 | 200 | 0.21 | 0.189 | 0.200 | 84 |
| 440 | 155 | 0.23 | 0.178 | 0.200 | 88 |
| 460 | 122 | 0.15 | 0.168 | 0.200 | 92 |
| 480 | 85 | 0.15 | 0.159 | 0.200 | 96 |
| 500 | 53 | 0.140 | 0.151 | 0.200 | 100 |
| 520 | 41 | 0.13 | 0.145 | 0.200 | 104 |
| 540 | 22 | 0.14 | 0.141 | 0.200 | 108 |
| 560 | 29 | 0.16 | 0.138 | 0.200 | 112 |
| 580 | 14 | 0.13 | 0.135 | 0.200 | 116 |
| 600 | 5 | 0.13 | 0.132 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Southwest Region



Linest and Logest Functions

Cubic =LINEST(X_1^(1,2,3,4,5,6))

a

-2.914E-16

b

5.439E-13

c

-3.415E-10

d

5.657E-08

2.29691E-05

-0.010318222

1.396971

ARM2 Poly Curve Fitting Analysis

Within 1 km of 100 tpy Sources

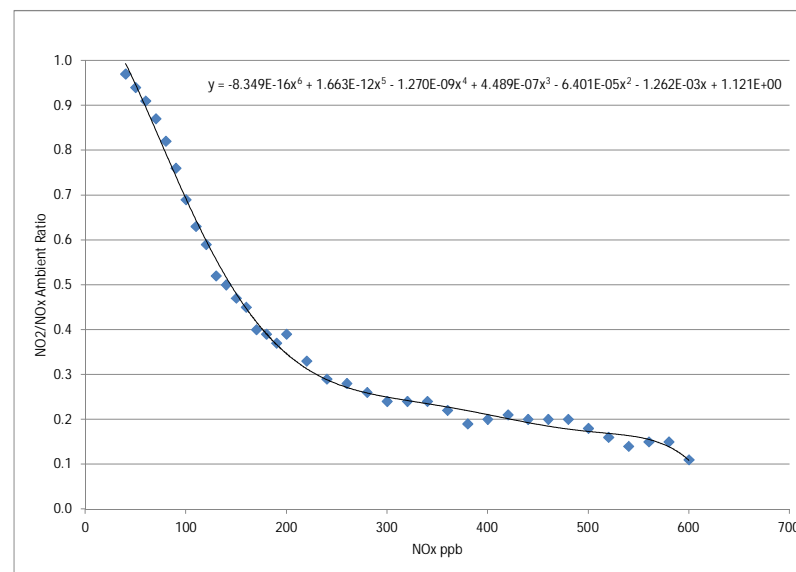
Min Ratio 0.2 Maximum Ratio

0.9

Total # Obs 277498

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 83331 | 1.000 | 1.036 | 0.900 | 27 |
| 40 | 52394 | 0.970 | 0.994 | 0.900 | 36 |
| 50 | 35868 | 0.940 | 0.946 | 0.900 | 45 |
| 60 | 25553 | 0.910 | 0.896 | 0.896 | 54 |
| 70 | 18735 | 0.870 | 0.845 | 0.845 | 59 |
| 80 | 13573 | 0.820 | 0.793 | 0.793 | 63 |
| 90 | 10272 | 0.760 | 0.742 | 0.742 | 67 |
| 100 | 7653 | 0.690 | 0.692 | 0.692 | 69 |
| 110 | 5898 | 0.630 | 0.644 | 0.644 | 71 |
| 120 | 4460 | 0.590 | 0.599 | 0.599 | 72 |
| 130 | 3442 | 0.520 | 0.556 | 0.556 | 72 |
| 140 | 2702 | 0.500 | 0.517 | 0.517 | 72 |
| 150 | 2234 | 0.470 | 0.480 | 0.480 | 72 |
| 160 | 1735 | 0.450 | 0.447 | 0.447 | 72 |
| 170 | 1420 | 0.400 | 0.417 | 0.417 | 71 |
| 180 | 1200 | 0.390 | 0.391 | 0.391 | 70 |
| 190 | 979 | 0.370 | 0.367 | 0.367 | 70 |
| 200 | 858 | 0.390 | 0.346 | 0.346 | 69 |
| 220 | 1361 | 0.330 | 0.313 | 0.313 | 69 |
| 240 | 934 | 0.290 | 0.289 | 0.289 | 69 |
| 260 | 652 | 0.280 | 0.271 | 0.271 | 71 |
| 280 | 480 | 0.260 | 0.259 | 0.259 | 73 |
| 300 | 434 | 0.240 | 0.250 | 0.250 | 75 |
| 320 | 300 | 0.240 | 0.242 | 0.242 | 78 |
| 340 | 224 | 0.240 | 0.235 | 0.235 | 80 |
| 360 | 169 | 0.220 | 0.228 | 0.228 | 82 |
| 380 | 129 | 0.190 | 0.220 | 0.220 | 83 |
| 400 | 118 | 0.200 | 0.211 | 0.211 | 84 |
| 420 | 103 | 0.210 | 0.202 | 0.202 | 85 |
| 440 | 65 | 0.200 | 0.193 | 0.200 | 88 |
| 460 | 57 | 0.200 | 0.185 | 0.200 | 92 |
| 480 | 51 | 0.200 | 0.178 | 0.200 | 96 |
| 500 | 45 | 0.180 | 0.173 | 0.200 | 100 |
| 520 | 25 | 0.160 | 0.169 | 0.200 | 104 |
| 540 | 13 | 0.140 | 0.164 | 0.200 | 108 |
| 560 | 19 | 0.150 | 0.155 | 0.200 | 112 |
| 580 | 5 | 0.150 | 0.139 | 0.200 | 116 |
| 600 | 7 | 0.110 | 0.109 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Within 1 km of 100 tpy Sources



Linest and Logest Functions

Cubic =LINEST(X_1^(1,2,3,4,5,6))

a

b

c

d

-8.349E-16

1.663E-12

-1.270E-09

4.489E-07

-6.40058E-05

-0.001261712

1.120815

ARM2 Poly Curve Fitting Analysis

1 to 5 km from 100 tpy Sources

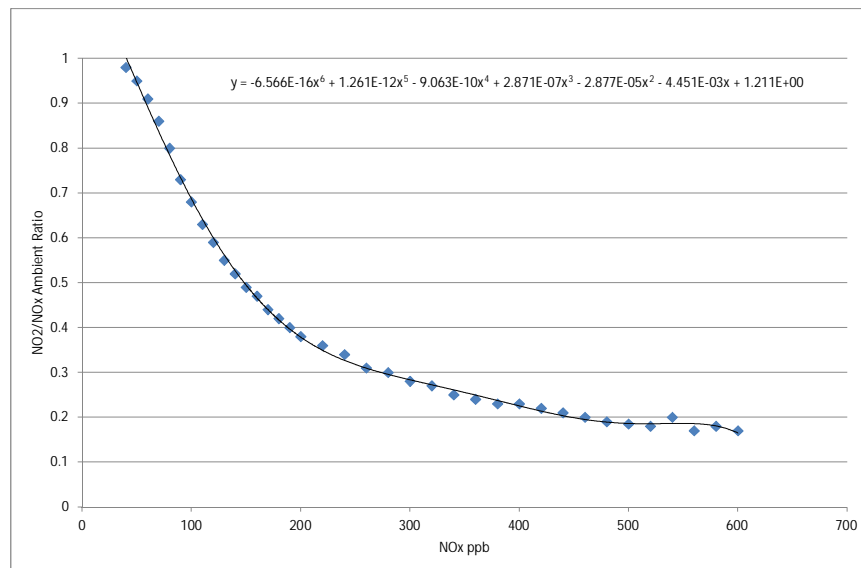
Min Ratio 0.2 Maximum Ratio

0.9

Total # Obs 2860546

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 953087 | 1 | 1.059 | 0.900 | 27 |
| 40 | 557450 | 0.98 | 1.003 | 0.900 | 36 |
| 50 | 352830 | 0.95 | 0.947 | 0.900 | 45 |
| 60 | 237786 | 0.91 | 0.892 | 0.892 | 54 |
| 70 | 168694 | 0.86 | 0.837 | 0.837 | 59 |
| 80 | 122978 | 0.8 | 0.785 | 0.785 | 63 |
| 90 | 92705 | 0.73 | 0.735 | 0.735 | 66 |
| 100 | 70982 | 0.68 | 0.687 | 0.687 | 69 |
| 110 | 55110 | 0.63 | 0.642 | 0.642 | 71 |
| 120 | 43202 | 0.59 | 0.601 | 0.601 | 72 |
| 130 | 34459 | 0.55 | 0.562 | 0.562 | 73 |
| 140 | 27769 | 0.52 | 0.527 | 0.527 | 74 |
| 150 | 22799 | 0.49 | 0.495 | 0.495 | 74 |
| 160 | 18601 | 0.47 | 0.466 | 0.466 | 75 |
| 170 | 15572 | 0.44 | 0.440 | 0.440 | 75 |
| 180 | 12847 | 0.42 | 0.417 | 0.417 | 75 |
| 190 | 10823 | 0.4 | 0.397 | 0.397 | 75 |
| 200 | 9084 | 0.38 | 0.379 | 0.379 | 76 |
| 220 | 14067 | 0.36 | 0.349 | 0.349 | 77 |
| 240 | 10177 | 0.34 | 0.327 | 0.327 | 78 |
| 260 | 7555 | 0.31 | 0.310 | 0.310 | 80 |
| 280 | 5442 | 0.3 | 0.296 | 0.296 | 83 |
| 300 | 4136 | 0.28 | 0.284 | 0.284 | 85 |
| 320 | 3028 | 0.27 | 0.272 | 0.272 | 87 |
| 340 | 2337 | 0.25 | 0.261 | 0.261 | 89 |
| 360 | 1748 | 0.24 | 0.250 | 0.250 | 90 |
| 380 | 1289 | 0.23 | 0.238 | 0.238 | 90 |
| 400 | 1034 | 0.23 | 0.225 | 0.225 | 90 |
| 420 | 798 | 0.22 | 0.214 | 0.214 | 90 |
| 440 | 573 | 0.21 | 0.204 | 0.204 | 90 |
| 460 | 472 | 0.2 | 0.195 | 0.200 | 92 |
| 480 | 326 | 0.19 | 0.189 | 0.200 | 96 |
| 500 | 291 | 0.185 | 0.186 | 0.200 | 100 |
| 520 | 171 | 0.18 | 0.186 | 0.200 | 104 |
| 540 | 111 | 0.2 | 0.186 | 0.200 | 108 |
| 560 | 107 | 0.17 | 0.186 | 0.200 | 112 |
| 580 | 62 | 0.18 | 0.181 | 0.200 | 116 |
| 600 | 44 | 0.17 | 0.165 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - 1 to 5 km from 100 tpy Sources



Linest and Logest Functions

Cubic =LINEST(X_1^(1,2,3,4,5,6))

a

-6.566E-16

b

1.261E-12

c

-9.063E-10

d

2.871E-07

-2.87686E-05

-0.00445067 1.211233

ARM2 Poly Curve Fitting Analysis

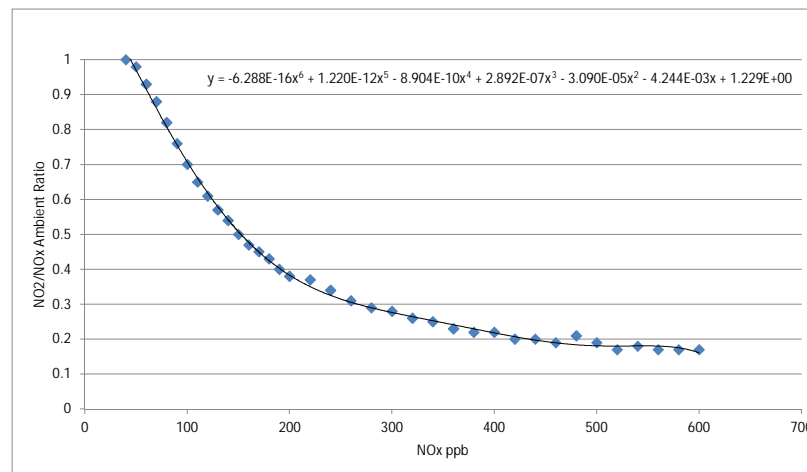
Greater than 5 km from 100 tpy Sources

Min Ratio 0.2 Maximum Ratio 0.9

Total # Obs 3900337

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 1357621 | 1 | 1.081 | 0.900 | 27 |
| 40 | 744075 | 1 | 1.027 | 0.900 | 36 |
| 50 | 459651 | 0.98 | 0.971 | 0.900 | 45 |
| 60 | 306175 | 0.93 | 0.915 | 0.900 | 54 |
| 70 | 215051 | 0.88 | 0.861 | 0.861 | 60 |
| 80 | 158346 | 0.82 | 0.808 | 0.808 | 65 |
| 90 | 120705 | 0.76 | 0.756 | 0.756 | 68 |
| 100 | 93413 | 0.7 | 0.708 | 0.708 | 71 |
| 110 | 74751 | 0.65 | 0.662 | 0.662 | 73 |
| 120 | 60579 | 0.61 | 0.619 | 0.619 | 74 |
| 130 | 49353 | 0.57 | 0.579 | 0.579 | 75 |
| 140 | 40396 | 0.54 | 0.542 | 0.542 | 76 |
| 150 | 33857 | 0.5 | 0.508 | 0.508 | 76 |
| 160 | 28214 | 0.47 | 0.478 | 0.478 | 76 |
| 170 | 23554 | 0.45 | 0.450 | 0.450 | 76 |
| 180 | 19841 | 0.43 | 0.425 | 0.425 | 77 |
| 190 | 16808 | 0.4 | 0.403 | 0.403 | 77 |
| 200 | 14292 | 0.38 | 0.383 | 0.383 | 77 |
| 220 | 22451 | 0.37 | 0.351 | 0.351 | 77 |
| 240 | 16135 | 0.34 | 0.326 | 0.326 | 78 |
| 260 | 11747 | 0.31 | 0.306 | 0.306 | 80 |
| 280 | 8728 | 0.29 | 0.290 | 0.290 | 81 |
| 300 | 6311 | 0.28 | 0.277 | 0.277 | 83 |
| 320 | 4766 | 0.26 | 0.265 | 0.265 | 85 |
| 340 | 3414 | 0.25 | 0.253 | 0.253 | 86 |
| 360 | 2532 | 0.23 | 0.242 | 0.242 | 87 |
| 380 | 1992 | 0.22 | 0.230 | 0.230 | 87 |
| 400 | 1482 | 0.22 | 0.218 | 0.218 | 87 |
| 420 | 1080 | 0.2 | 0.207 | 0.207 | 87 |
| 440 | 825 | 0.2 | 0.198 | 0.200 | 88 |
| 460 | 605 | 0.19 | 0.190 | 0.200 | 92 |
| 480 | 482 | 0.21 | 0.184 | 0.200 | 96 |
| 500 | 364 | 0.190 | 0.181 | 0.200 | 100 |
| 520 | 242 | 0.17 | 0.181 | 0.200 | 104 |
| 540 | 199 | 0.18 | 0.181 | 0.200 | 108 |
| 560 | 132 | 0.17 | 0.181 | 0.200 | 112 |
| 580 | 97 | 0.17 | 0.176 | 0.200 | 116 |
| 600 | 71 | 0.17 | 0.161 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - Greater than 5 km from 100 tpy Sources



Linest and Logest Functions

Cubic =LINEST(X_1^{1,2,3,4,5,6})

a

-6.288E-16

b

1.220E-12

c

-8.904E-10

d

2.892E-07

-3.09044E-05

-0.00424388

1.229419

ARM2 Poly Curve Fitting Analysis

2001-2003

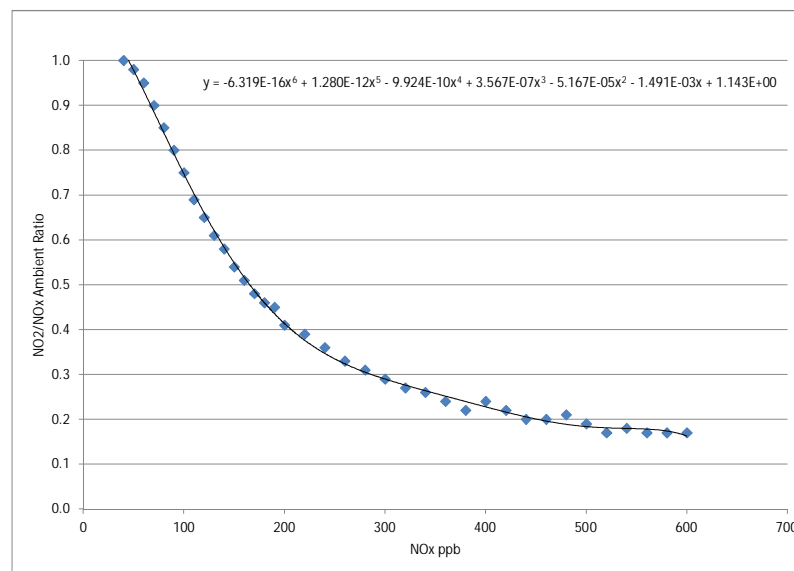
Min Ratio 0.2 Maximum Ratio

0.9

Total # Obs 2383063

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 744455 | 1.000 | 1.061 | 0.900 | 27 |
| 40 | 443379 | 1.000 | 1.021 | 0.900 | 36 |
| 50 | 287844 | 0.980 | 0.978 | 0.900 | 45 |
| 60 | 197928 | 0.950 | 0.933 | 0.900 | 54 |
| 70 | 142352 | 0.900 | 0.886 | 0.886 | 62 |
| 80 | 105246 | 0.850 | 0.839 | 0.839 | 67 |
| 90 | 80624 | 0.800 | 0.793 | 0.793 | 71 |
| 100 | 63110 | 0.750 | 0.747 | 0.747 | 75 |
| 110 | 50459 | 0.690 | 0.703 | 0.703 | 77 |
| 120 | 41019 | 0.650 | 0.661 | 0.661 | 79 |
| 130 | 33297 | 0.610 | 0.621 | 0.621 | 81 |
| 140 | 27659 | 0.580 | 0.584 | 0.584 | 82 |
| 150 | 23192 | 0.540 | 0.549 | 0.549 | 82 |
| 160 | 19517 | 0.510 | 0.516 | 0.516 | 83 |
| 170 | 16682 | 0.480 | 0.487 | 0.487 | 83 |
| 180 | 14130 | 0.460 | 0.460 | 0.460 | 83 |
| 190 | 12315 | 0.450 | 0.436 | 0.436 | 83 |
| 200 | 10569 | 0.410 | 0.414 | 0.414 | 83 |
| 220 | 16807 | 0.390 | 0.377 | 0.377 | 83 |
| 240 | 12542 | 0.360 | 0.347 | 0.347 | 83 |
| 260 | 9452 | 0.330 | 0.324 | 0.324 | 84 |
| 280 | 7207 | 0.310 | 0.305 | 0.305 | 85 |
| 300 | 5515 | 0.290 | 0.290 | 0.290 | 87 |
| 320 | 4223 | 0.270 | 0.276 | 0.276 | 88 |
| 340 | 3259 | 0.260 | 0.264 | 0.264 | 90 |
| 360 | 2452 | 0.240 | 0.252 | 0.252 | 91 |
| 380 | 1907 | 0.220 | 0.240 | 0.240 | 91 |
| 400 | 1501 | 0.240 | 0.228 | 0.228 | 91 |
| 420 | 1163 | 0.220 | 0.216 | 0.216 | 91 |
| 440 | 822 | 0.200 | 0.206 | 0.206 | 91 |
| 460 | 672 | 0.200 | 0.196 | 0.200 | 92 |
| 480 | 510 | 0.210 | 0.189 | 0.200 | 96 |
| 500 | 414 | 0.190 | 0.184 | 0.200 | 100 |
| 520 | 267 | 0.170 | 0.181 | 0.200 | 104 |
| 540 | 222 | 0.180 | 0.180 | 0.200 | 108 |
| 560 | 153 | 0.170 | 0.179 | 0.200 | 112 |
| 580 | 116 | 0.170 | 0.174 | 0.200 | 116 |
| 600 | 82 | 0.170 | 0.163 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - 2001-2003



Linest and Logest Functions

Cubic =LINEST($X_1^{(1,2,3,4,5,6)}$)

a

b

c

d

-6.319E-16

1.280E-12

-9.924E-10

3.567E-07

-5.16661E-05

-0.001490704

1.143189

ARM2 Poly Curve Fitting Analysis

2004-2007

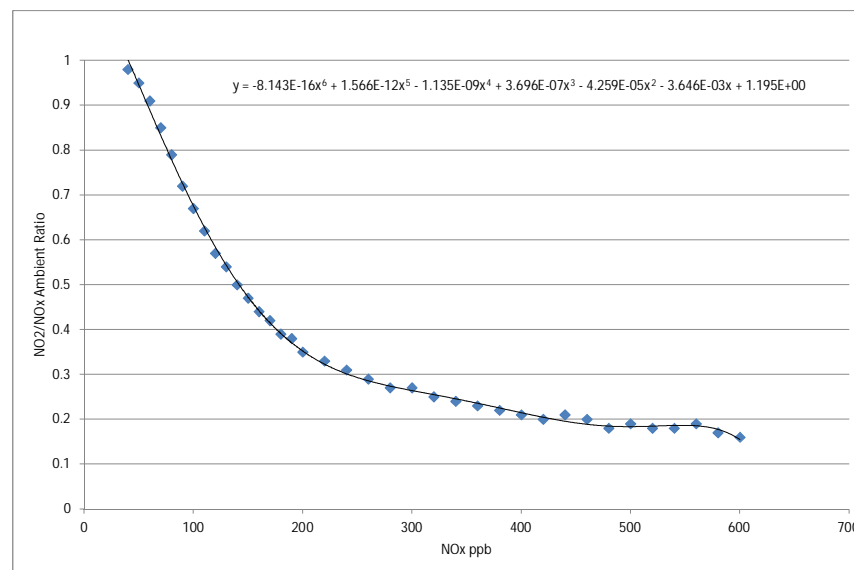
Min Ratio 0.2 Maximum Ratio

0.9

Total # Obs 2762848

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 955828 | 1 | 1.057 | 0.900 | 27 |
| 40 | 538672 | 0.98 | 1.002 | 0.900 | 36 |
| 50 | 334416 | 0.95 | 0.946 | 0.900 | 45 |
| 60 | 221739 | 0.91 | 0.890 | 0.890 | 53 |
| 70 | 155190 | 0.85 | 0.833 | 0.833 | 58 |
| 80 | 113430 | 0.79 | 0.779 | 0.779 | 62 |
| 90 | 85658 | 0.72 | 0.726 | 0.726 | 65 |
| 100 | 65412 | 0.67 | 0.676 | 0.676 | 68 |
| 110 | 51708 | 0.62 | 0.628 | 0.628 | 69 |
| 120 | 40833 | 0.57 | 0.584 | 0.584 | 70 |
| 130 | 32973 | 0.54 | 0.544 | 0.544 | 71 |
| 140 | 26895 | 0.5 | 0.506 | 0.506 | 71 |
| 150 | 22473 | 0.47 | 0.473 | 0.473 | 71 |
| 160 | 18526 | 0.44 | 0.442 | 0.442 | 71 |
| 170 | 15296 | 0.42 | 0.415 | 0.415 | 71 |
| 180 | 12752 | 0.39 | 0.391 | 0.391 | 70 |
| 190 | 10583 | 0.38 | 0.370 | 0.370 | 70 |
| 200 | 9014 | 0.35 | 0.352 | 0.352 | 70 |
| 220 | 14198 | 0.33 | 0.323 | 0.323 | 71 |
| 240 | 10120 | 0.31 | 0.302 | 0.302 | 72 |
| 260 | 7437 | 0.29 | 0.286 | 0.286 | 74 |
| 280 | 5386 | 0.27 | 0.274 | 0.274 | 77 |
| 300 | 3869 | 0.27 | 0.264 | 0.264 | 79 |
| 320 | 2845 | 0.25 | 0.255 | 0.255 | 82 |
| 340 | 2028 | 0.24 | 0.246 | 0.246 | 84 |
| 360 | 1464 | 0.23 | 0.236 | 0.236 | 85 |
| 380 | 1140 | 0.22 | 0.225 | 0.225 | 86 |
| 400 | 824 | 0.21 | 0.215 | 0.215 | 86 |
| 420 | 595 | 0.2 | 0.205 | 0.205 | 86 |
| 440 | 459 | 0.21 | 0.196 | 0.200 | 88 |
| 460 | 331 | 0.2 | 0.189 | 0.200 | 92 |
| 480 | 234 | 0.18 | 0.185 | 0.200 | 96 |
| 500 | 196 | 0.19 | 0.184 | 0.200 | 100 |
| 520 | 113 | 0.18 | 0.185 | 0.200 | 104 |
| 540 | 77 | 0.18 | 0.186 | 0.200 | 108 |
| 560 | 73 | 0.19 | 0.186 | 0.200 | 112 |
| 580 | 35 | 0.17 | 0.177 | 0.200 | 116 |
| 600 | 26 | 0.16 | 0.155 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - 2004-2007



Linest and Logest Functions

Cubic =LINEST(X_1^(1,2,3,4,5,6))

a

-8.143E-16

b

1.566E-12

c

-1.135E-09

d

3.696E-07

-4.25882E-05 -0.003645685 1.195252

ARM2 Poly Curve Fitting Analysis

2008-2010

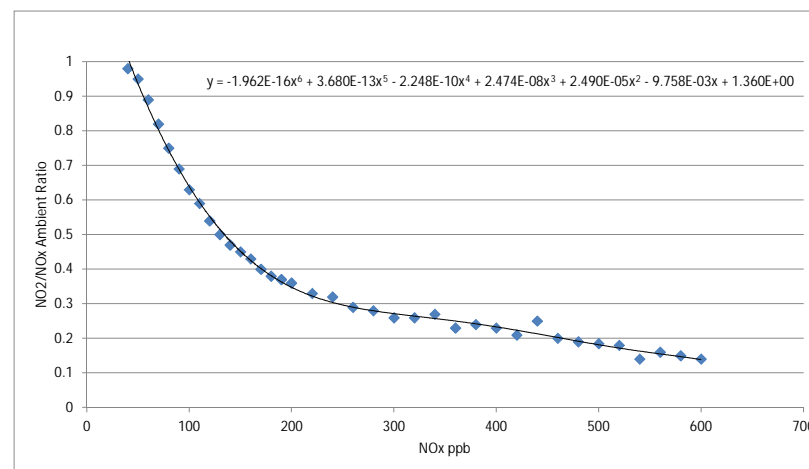
Min Ratio 0.2 Maximum Ratio

0.9

Total # Obs 1658437

| Mid Pt. Nox BIN ppb | Number in Bin | Observed 98% Perc ARM | Step 1 Fitted ARM Ratio | Final Adjusted ARM Ratio | ARM2 Calculated NO2 ppb |
|------------------------|------------------|--------------------------|----------------------------|-----------------------------|----------------------------|
| 30 | 630503 | 1 | 1.090 | 0.900 | 27 |
| 40 | 328716 | 0.98 | 1.011 | 0.900 | 36 |
| 50 | 195120 | 0.95 | 0.936 | 0.900 | 45 |
| 60 | 127090 | 0.89 | 0.867 | 0.867 | 52 |
| 70 | 87963 | 0.82 | 0.803 | 0.803 | 56 |
| 80 | 63770 | 0.75 | 0.744 | 0.744 | 59 |
| 90 | 47963 | 0.69 | 0.689 | 0.689 | 62 |
| 100 | 36477 | 0.63 | 0.639 | 0.639 | 64 |
| 110 | 28149 | 0.59 | 0.594 | 0.594 | 65 |
| 120 | 22241 | 0.54 | 0.553 | 0.553 | 66 |
| 130 | 17786 | 0.5 | 0.515 | 0.515 | 67 |
| 140 | 13796 | 0.47 | 0.482 | 0.482 | 67 |
| 150 | 11136 | 0.45 | 0.452 | 0.452 | 68 |
| 160 | 8913 | 0.43 | 0.426 | 0.426 | 68 |
| 170 | 7242 | 0.4 | 0.402 | 0.402 | 68 |
| 180 | 5887 | 0.38 | 0.382 | 0.382 | 69 |
| 190 | 4798 | 0.37 | 0.364 | 0.364 | 69 |
| 200 | 3853 | 0.36 | 0.348 | 0.348 | 70 |
| 220 | 5613 | 0.33 | 0.323 | 0.323 | 71 |
| 240 | 3711 | 0.32 | 0.304 | 0.304 | 73 |
| 260 | 2460 | 0.29 | 0.290 | 0.290 | 75 |
| 280 | 1611 | 0.28 | 0.280 | 0.280 | 78 |
| 300 | 1084 | 0.26 | 0.272 | 0.272 | 82 |
| 320 | 748 | 0.26 | 0.265 | 0.265 | 85 |
| 340 | 487 | 0.27 | 0.258 | 0.258 | 88 |
| 360 | 377 | 0.23 | 0.250 | 0.250 | 90 |
| 380 | 240 | 0.24 | 0.242 | 0.242 | 92 |
| 400 | 199 | 0.23 | 0.234 | 0.234 | 93 |
| 420 | 122 | 0.21 | 0.224 | 0.224 | 94 |
| 440 | 120 | 0.25 | 0.213 | 0.213 | 94 |
| 460 | 76 | 0.2 | 0.203 | 0.203 | 93 |
| 480 | 66 | 0.19 | 0.192 | 0.200 | 96 |
| 500 | 48 | 0.185 | 0.182 | 0.200 | 100 |
| 520 | 33 | 0.18 | 0.172 | 0.200 | 104 |
| 540 | 11 | 0.14 | 0.163 | 0.200 | 108 |
| 560 | 13 | 0.16 | 0.155 | 0.200 | 112 |
| 580 | 8 | 0.15 | 0.147 | 0.200 | 116 |
| 600 | 7 | 0.14 | 0.139 | 0.200 | 120 |

ARM2 Poly Curve Fitting Analysis - 2008-2010



Linest and Logest Functions

Cubic =LINEST(X_1^{1,2,3,4,5,6})

| a | b | c | d |
|------------|-----------|------------|-----------|
| -1.962E-16 | 3.680E-13 | -2.248E-10 | 2.474E-08 |

| | | |
|-------------|--------------|----------|
| 2.48951E-05 | -0.009757621 | 1.360263 |
|-------------|--------------|----------|

APPENDIX C

User Instructions for AERMOD-ARM2

User Instructions for AERMOD-ARM2

A new model option is available in AERMOD-ARM2 version 12345. The model option keyword “ARM2” can be selected (not in combination with either of the PVMRM or OLM options). The ARM2 maximum and minimum conversion ratios are set by default to 0.9 and 0.2; however, the keywords NO2EQUIL and NO2STACK can be used to change these default settings if necessary (if these keywords are missing, the default values will be used). The following AERMOD control input example illustrates the setup of the ARM2 option:

```
CO STARTING
  TITLEONE AERMOD-ARM2 Model Run
  TITLETWO
  MODELOPT CONC ARM2
  AVERTIME 1
  POLLUTID NO2
  RUNORNOT RUN
  NO2STACK 0.2
  NO2EQUIL 0.9
CO FINISHED
```

Since the ARM2 ambient ratio is determined from the cumulative modeled NO_x concentration, there must be an “ALL” source group in the AERMOD run (if an “ALL” source group is not included in a run, AERMOD halts with an error message).

When the ARM2 option is enabled, AERMOD calculates the cumulative 1-hr NO_x concentration (i.e., based on the source group “ALL”) at each receptor on an hour-by-hour basis. Each hourly NO_x concentration for the “ALL” source group is input to the ARM2 equation to determine the ARM2 ambient ratio for that hour. The ARM2 ambient ratio for that hour is then multiplied by the NO_x concentration for any other source groups in the model run to determine that hour’s NO₂ concentration for each source group. The hourly NO₂ concentrations for each receptor and source group are then stored as usual into the appropriate AERMOD data array for subsequent output processing.
